Nuclear Power in Canada:



An Examination of Risks, Impacts and Sustainability





Mark Winfield, PhD Alison Jamison Rich Wong Paulina Czajkowski

Acknowledgements

The Pembina Institute wishes to thank the EJLB Foundation and the Oak Foundation for their financial support of this project.

We are grateful to our external reviewers who provided extremely helpful and constructive comments throughout the development of this report.

We are also grateful to the staff of nuclear facilities and regulatory agencies who provided prompt responses to our many, and often very detailed questions.

Design and Layout by Green Living Communications

Cover photos: Rabbit Lake Uranium Mine: MiningWatch Canada; Bruce Nuclear Station: Bruce Power; Elliot Lake landscape: Jamie Kneen, MiningWatch Canada.

Publication Date: December 2006

Authors: Mark Winfield, with contributions by Alison Jamison, Rich Wong, and Paulina Czajkowski

ISBN 0-921719-87-6

About the Pembina Institute

The Pembina Institute is an independent, not-for-profit environmental policy research and education organization specializing in the fields of sustainable energy, community sustainability, climate change and corporate environmental management. Founded in 1985 in Drayton Valley Alberta, the Institute now has offices in Calgary, Edmonton, Vancouver, Ottawa and Toronto.

For more information on the Institute's work, please visit our website at www.pembina.org.



Sustainable Energy Solutions

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Executive summary

This study examines the environmental impacts of the use of nuclear energy for electricity generation in Canada through each of the four major stages of nuclear energy production: uranium mining and milling; uranium refining, conversion and fuel fabrication; nuclear power plant operation; and waste fuel management. It is intended to inform public debate over the future role of nuclear energy in Canada, and to facilitate comparisons of nuclear energy with other potential energy sources.

The study examines waste generation, atmospheric releases, impacts on water quality and water use, and landscape and ecosystem impacts of nuclear energy production. It also examines the occupational and community health impacts of nuclear power and key long-term challenges to its sustainability, including security and weapons proliferation risks. Specific environmental impacts are examined in the context of CANDU nuclear technology, the only reactor type currently in use in Canada.*

The study findings likely underestimate the overall impacts of the use of nuclear energy for electricity production in Canada. This is a result of significant gaps in the publicly available information on releases of pollutants and contaminants, as well as on the fate of certain waste streams related to the nuclear industry. In addition, the study relies on what are likely conservative estimates in a number of key areas, particularly with respect to the generation of greenhouse gas (GHG) emissions.

Is nuclear power clean?

The study finds that nuclear power, like other nonrenewable energy sources, is associated with severe environmental impacts. Each stage of the nuclear energy production process generates large amounts of uniquely difficult-to-manage wastes that will effectively require perpetual care, imposing costs and risks arising from current energy consumption onto future generations. The process also has severe impacts on surface water and groundwater water quality via a range of radioactive and hazardous pollutants, and results in releases to the atmosphere of a wide range of criteria (i.e. smog and acid-rain causing), radioactive and hazardous pollutants and greenhouse gases. Effluent from uranium mines and mills was found by Health Canada and Environment Canada to be 'toxic' for the purposes of the *Canadian Environmental Protection Act* in 2004.

What is particularly noteworthy about the radioactive waste streams produced at every stage of the nuclear life cycle are the timeframes over which these materials will need to be managed. Secure containment will be required for not hundreds, but hundreds of thousands of years - timeframes over which it is extremely difficult, if not virtually impossible, to predict outcomes with any level of assurance. There are no approved long-term strategies for the management of these wastes in place. The federally mandated Nuclear Waste Management Organization expects it will take over 300 years to implement its proposed "phased adaptive management" approach to containing waste nuclear fuel. As well, the effectiveness and adequacy of tailings management facilities at mine sites in Canada has been subject to serious question. There is a long history of uranium mine tailings management facility failures in Canada and elsewhere in the world, resulting in severe surface water and groundwater contamination.

Is nuclear power sustainable?

Nuclear energy is no more a renewable energy source than oil or gas. It relies on a finite and non-renewable fuel supply – uranium. World uranium prices have increased more than sixfold since 2001. Current Canadian uranium reserves are estimated to be sufficient for 40 years at current levels of consumption (compared to estimated natural gas reserves of

^{*(}Different types of reactors are associated with different impacts and risks. Light-water reactors, employing enriched uranium fuel, for example, are associated with the generation of lower volumes of waste fuel. However, the process of producing enriched uranium fuel for these types of reactors is associated with much higher emissions of greenhouse gases, particularly where gas diffusion based enrichment processes are employed, as well as higher atmospheric releases of uranium and the generation of large volumes of depleted uranium (DU) wastes.)

approximately 70 years). The exploitation of lowergrade uranium deposits in the future would increase the already substantial emissions (including greenhouse gas emissions) from uranium mining and milling operations, as well as significantly expanding the enormous amounts of waste rock and tailings generated by uranium mines and mills.

Efforts to increase the available fuel supply through the reprocessing of waste fuel or the use of fast breeder reactors are seen to present serious waste management, technological and weapons proliferation risks. Other suggested fuel sources, such as thorium or extraction of uranium from seawater, face major technological, environmental and economic hurdles.

Is nuclear power greenhouse gas 'emissions free'?

The study finds that GHG emissions arise at each stage of the nuclear energy cycle, with power plant construction being the most significant source of releases. Further releases of GHGs occur as a result of the operation of equipment in the uranium mining process, the milling of uranium ore, mill tailings management activities, and refining and conversion operations. The generation of greenhouse gases from mining and milling operations would increase proportionally with the use of lower grade uranium ores, as larger amounts of ore would have to be extracted and processed to produce the same amount of uranium concentrate.

The road transportation of uranium between milling, refining and conversion facilities results in additional releases. As with criteria air pollutants, the management of waste nuclear fuel along with other radioactive wastes could involve significant transportation activities, leading to further generation of GHG emissions.

In Canada, total GHG emissions associated with uranium mining, milling, refining, conversion and fuel fabrication are between 240,000 and 366,000 tonnes of CO_2 per year. Total emissions associated with the sector, including the emissions associated with power plant construction, are in the range of 468,000 and 594,000 tonnes of CO_2 per year, equivalent to the emissions of between 134,000 and 170,000 cars per year. Total annual GHG emissions that are primarily associated with domestic power production are estimated at between 267,000 and 289,000 tonnes of CO_2 per year. This total is almost certainly an underestimate, due to a lack of complete informa-

tion. Other recent estimates suggest total GHG emissions associated with nuclear power in Canada are in the range of at least 840,000 tonnes per year.

These figures relate to the CANDU-type reactors used in Canada. The process of producing enriched uranium fuel for other types of reactors is associated with much higher emissions of greenhouse gases, particularly where gas diffusion-based enrichment processes are employed.

Is Nuclear Power Reliable?

The Ontario CANDU reactor fleet has been subject to severe performance and maintenance problems. Over the past decade, some Ontario facilities have had average operating capacities below 40 per cent rather than the expected 85–90 per cent range. Reactors expected to have operational lifetimes in the range of 40 years have turned out to require major refurbishments after approximately 25 years of service. Refurbishment projects themselves have run seriously over budget and behind schedule.

Heavy reliance on coal-fired electricity to backstop under-performing or offline nuclear units has been associated with major increases in releases of greenhouse gases and other air pollutants. The shutdown of eight reactors between 1995 and 2001under the 1997 Nuclear Asset Optimization Plan led to emissions of GHGs from the province's coal-fired power plants increasing by a factor of 2.3, sulphur dioxide emissions by a factor of 2, and nitrogen oxide emissions by a factor of 1.7, significantly exacerbating the severe air quality problems regularly experienced in southern Ontario.

Is it a cost-effective solution?

Nuclear power generating facilities are subject to very high capital costs and long construction times relative to other electricity supply options. In addition, in Ontario there is a history of serious delays and cost overruns on nuclear generating facility projects, accounting for \$15 billion of the nearly \$20 billion "stranded debt" left by Ontario Hydro.

Nuclear energy also brings with it a unique set of risks, largely arising from the very high costs and levels of uncertainty involved in handling, storing and managing waste fuel and other radioactive wastes. Implementation of the Nuclear Waste Management Organization's proposed strategy for managing waste fuel from existing reactors is estimated to be likely to have a total cost in the range of \$24 billion. This would be in addition to the costs for the development and management of facilities for low and intermediate level radioactive waste and for managing waste rock and tailings at uranium mine sites. The costs of decommissioning Ontario's existing reactors have been estimated at \$7.474 billion.

Even with extensive subsidies and financial guarantees provided by governments, these costs, timelines and risks make it difficult for nuclear power projects to compete for private capital investments against potential investments that will bring much more rapid and secure returns.

Is it safe?

Much has changed in our understanding of radiation risks since the construction of Canada's first commercial reactors in the early 1970s. For example, recent research on the effects of even very low levels of ionizing radiation suggests that no level is safe to health. The International Agency for Research on Cancer (IARC) lists a number of radionuclides as carcinogenic to humans, including isotopes produced in uranium mining and milling, fuel production and nuclear power plant operations.

Yet despite our improved understanding of these risks, Canadian standards and practices appear to have not kept pace with this changing knowledge. It has been suggested, for example, that existing standards in Canada for cancer risks arising from radiological hazards permit much higher levels of acceptable risk than is the case for chemical and other hazards. Current Canadian standards in some areas are substantially weaker than those in place in other comparable jurisdictions. The existing drinking water standard in Ontario for tritium (of which discharges from nuclear power plants are the primary source), for example, of 7,000 Bq/L is significantly weaker than the standards in the United States of 740 Bq/L and in the European Union of 100 Bq/L.

Workers in the mining and refining, conversion and fuel fabrication sub-sectors are also found to be routinely exposed to levels of radiation above those that would be considered acceptable to members of the general public. There is a history of significant occupational health effects, particularly elevated incidences of lung cancer, among uranium miners attributed to radon exposure. Increased mortality among uranium miners is also attributed to exposure to silica, solvents, asbestos and radiation.

As well, substantial health risks have been identified in relation to the consumption of certain types of "country" food, particularly caribou, in the vicinity of uranium mine/mill operations as a result of contamination by radionuclides.

While nuclear generating facility operators argue that the levels of public exposure to radiation arising from facility operations are trivial in comparison to other sources, recent studies suggest that health impacts of low-level radiation exposure may be more significant than previously thought, and that children and infants may be particularly at risk from such exposures.

Nuclear generating facilities are additionally subject to uniquely severe accident and security risks. A serious accident or incident could result in the release of large amounts of radioactive material to the atmosphere, which could be distributed over a large area. By comparison, the impacts of major incidents or accidents at facilities employing other generating technologies would be short term and largely limited to the facility site itself. It has been estimated that the monetized value of the off-site environmental, health and economic impacts of a major accident at the Darlington generating facility east of the City of Toronto, for example, would exceed \$1 trillion (1991 \$Cdn).

Nuclear energy's shared origins with nuclear weapons programs raises the potential for -- and reality of -- links between technologies and materials used for energy production and for nuclear weapons development. Concerns about these connections have grown in the past few years as a result of nuclear programs in North Korea, Iran, India and Pakistan. Any large-scale expansion of reliance on nuclear energy would carry significant risks of the proliferation of materials and technologies that could be applied to weapons development. India's 1974 nuclear bomb test, a project developed in part using Canadian-supplied technology and uranium, demonstrated this problem clearly.

The big picture

Any life-cycle analysis of an energy source is likely to identify previously unrecognized or un-quantified impacts. However, the range and scale of impacts and risks associated with nuclear power production make it unique among energy sources.

While the greenhouse gas emissions associated with nuclear power are less than those that would be associated with conventional fossil fuel energy use, no other energy source combines the generation of a range of conventional pollutants and waste streams – including heavy metals, smog and acid rain precursors, and water contaminants – with the generation of extremely large volumes of radioactive wastes that will require care and management over hundreds of thousands of years. The combination of these environmental challenges, along with security, accident and weapons proliferation risks that are simply not shared by any other energy source, place nuclear energy in a unique category relative to all other energy supply options. In essence, reliance on nuclear power as a response to climate change would involve trading one problem – greenhouse gas emissions – for which a wide range of other solutions exist, for a series of other complex and difficult problems for which solutions are generally more costly and difficult and for which the outcomes are much less certain.

In this context, proposals for the retention and expansion of the role of nuclear power must be approached with the greatest of caution. Such proposals must be examined in the full light of their environmental, economic and security implications, not only for Canada, but the rest of the world as well. They must also be examined in the context of the full range of available alternatives. Such an examination is likely to conclude that better options are readily available. These options range from making the most efficient use possible of existing energy resources to expanding the role of low-impact renewable energy sources that offer far safer, cheaper, more reliable and more sustainable options for meeting society's energy needs.

Nuclear energy production waste streams - a synopsis

Solid and Liquid Wastes

Uranium mining and milling

- An estimated 575,000 tonnes of tailings per year, of which 90–100,000 tonnes can be attributed to uranium production for domestic energy purposes. Uranium mill tailings are acidic or potentially acid generating, and contain a range of long-lived radionuclides, heavy metals and other contaminants. Tailings generation would increase proportionally with the use of lower grade uranium ores, as larger amounts of ore would have to be processed to produce the same amount of uranium concentrate.
- Up to 18 million tonnes of waste rock, which may also contain radionuclides, heavy metals, and be acid generating. Of this total, up to 2.9 million tonnes can be attributed to uranium mining for domestic energy purposes.

• It is estimated that there are more than 213 million tonnes of uranium mine tailings in storage facilities in Canada, and 109 million tonnes of waste rock.

Refining and conversion operations

• It is estimated that nearly 1,000 tonnes of solid wastes and 9,000 m3 of liquid wastes are produced per year as a result of uranium refining, conversion and fuel production for domestic energy generation purposes. Information on the precise character and fate of these wastes could not be obtained.

Power Plant operation

- Approximately 85,000 waste fuel bundles are generated by Canadian nuclear reactors each year. As of 2003, 1.7 million bundles were in storage at reactor sites. It is estimated that these wastes will have to be secured for approximately one million years for safety, environmental and security reasons.
- Approximately 6,000 cubic metres of lower level radioactive wastes are generated each year in Ontario as a result of power plant operations, maintenance, and refurbishment.
- Power plant maintenance and refurbishment also result in the generation of substantial amounts of additional hazardous wastes, including heavy metals and asbestos.
- Very large amounts of low-, intermediate- and high-level radioactive wastes will be produced as a result of the eventual decommissioning of refining, conversion and fabrication facilities as well as power plants.

Water

- Severe contamination of groundwater with radionuclides, heavy metals, and other contaminants has occurred at tailings management facilities and waste rock storage areas.
- Uranium mining and milling facility surface water discharges have resulted in the contamination of the receiving environment with radionuclides and heavy metals. Effluent from historic and operating uranium mines and mills, particularly uranium discharges, have been determined to be toxic for the purposes of the *Canadian Environmental Protection Act*.
- Uranium mining operations are associated with the extensive removal of groundwater (in excess of 16 billion litres per year).

- Routine and accidental releases of radionuclides to surface waters occur in the course of power plant operations, with tritium oxide and carbon-14 being key radioactive pollutants of concern. Groundwater contamination with tritium has occurred at the Pickering generating facility in Ontario.
- Ontario's nuclear power plants are found to be the leading source of discharges of hydrazine, an extremely hazardous pollutant, to surface waters in Canada. Nuclear generating facilities have also been sources of discharges of metals (copper, zinc, and chromium) and ammonia to surface waters.
- Nuclear power is a major consumer of water. Uranium mining operations involve extensive dewatering, in the range of at least 16–17 billion litres per year, with the implication of impacts on groundwater and surface water storage and flows.
- Generating facilities require large amounts of cooling water. The Darlington and Pickering facilities in Ontario are alone estimated to use approximately 8.9 trillion litres of water for cooling purposes per year — more than 19 times the annual water consumption of the City of Toronto. Adverse thermal impacts of cooling water discharges on fish populations in the vicinity of nuclear power plants have been observed.

Air

- Atmospheric releases of a range of radionuclides occur at all stages of nuclear power production. Atmospheric releases of radon gas result from mining and milling operations and from tailings management facilities. Windblown dust from mine sites and tailings management facilities (TMFs) contains a range of radionuclides. Atmospheric releases (principally uranium) also arise from refining and conversion activities.
- Routine and accidental releases of radiation and radionuclides occur from power plant operations, including tritium oxide, carbon-14, noble gases, iodine-131, radioactive particulate and elemental tritium.
- The incineration of low and intermediate-level radioactive wastes from power plant operations and maintenance in Ontario has resulted in further atmospheric releases of radionuclides, particularly tritium. A wide range of hazardous air pollutants have been released by the Bruce Western Waste Management facility. A new incinerator installed in 2003, has reduced emissions of hazardous, but not of radiological, pollutants .

- Windblown dust from mine sites and TMFs contains a range of heavy metals. In addition, releases of a number of hazardous air pollutants, including dioxins and furans, hexachlorobenzene, heavy metals (principally lead) ammonia and hydrogen fluoride arise from uranium refining and conversion operations.
- Ontario nuclear power plants are the only National Pollutant Release Inventory reported source of releases of hydrazine to the air in Canada.
- Uranium mining and milling operations are found to be significant sources of releases of sulphur dioxide (SO₂), volatile organic compounds (VOCs) and nitrogen oxides (NOx). Releases of NOx, particulate matter (PM) and sulphuric acid arise from refining and conversion activities.
- The road transportation of uranium from mill sites in northern Saskatchewan to the Blind River refinery in Northern Ontario and then on to the Port Hope conversion facility in Southern Ontario produces additional releases of NOx and PM. Further transportation related releases of criteria air pollutants would arise from the longterm management of waste nuclear fuel and other radioactive wastes arising from facility operations, maintenance and decommissioning, particularly if the management strategies for these materials require the movement of wastes from reactor sites to centralized facilities.

Climate

- Total greenhouse gas (GHG) emissions associated with uranium mining, milling, refining, conversion and fuel fabrication in Canada are estimated at between 240,000 and 366,000 tonnes of CO₂ per year.
- Total emissions associated with the sector, including the emissions associated with power plant construction, are in the range of 468,000 and 594,000 tonnes of CO₂ per year, equivalent to the emissions of between 134,000 and 170,000 cars per year.
- Total annual GHG emissions associated with domestic power production alone are estimated at between 267,000 and 289,000 tonnes of CO₂ per year. Other recent estimates suggest total GHG emissions associated with nuclear power in Canada are in the range of at least 840,000 tonnes per year.

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1. Introduction

1.1. Project Overview and Rationale

This study has its origins in the Pembina Institute's *Power for the Future* project on future electricity policy options for Ontario, published in May 2004.¹ In attempting to examine the environmental and economic risks, costs and benefits associated with different potential sources of electricity supply, it became apparent that there was no publicly accessible overview of the life cycle environmental impacts of nuclear power in Canada.

The most comprehensive recent document available was the assessment completed by Environment Canada and Health Canada of the toxicity of radionuclide releases from nuclear facilities in Canada. The assessment dealt with all subsectors of the industry, including uranium mines and mills, refining, conversion and fuel fabrication facilities, power reactors, and waste management facilities. While concluding that releases from uranium mines and mills were "toxic" as defined by the Canadian Environmental Protection Act, the assessment did not consider the generation of non-radioactive waste substances and their impacts.² Nor did it examine impacts beyond the release of potentially toxic substances, such as the disruption of landscapes and of surface water and groundwater flows by uranium mining activities, or impacts on human health.

The Pembina Institute concluded that the availability of such an overview was essential to understand the costs and risks associated with nuclear power and to compare nuclear power with other potential sources of energy supply. The need for such an assessment became particularly acute in the context of the Ontario Power Authority's (OPA)³ December 2005 recommendations that the Province of Ontario commit \$30–40 billion to create 9,400 to 12,400 MW of new or refurbished nuclear generation capacity over the next 20 years.⁴ Proposals have also been advanced for the use of nuclear energy to support the exploitation of Alberta's oil sands.⁵

On the basis of the OPA's December 2005 advice, in June 2006 the Government of Ontario directed the

authority to develop a 20-year Integrated Power System Plan for the province, including 14,000 MW of nuclear generating capacity.⁶ Ontario Power Generation (OPG)⁷ was directed at the same time to undertake feasibility studies for refurbishing a number of its existing nuclear generating facilities, and to begin the work needed for an environmental assessment of the construction of new units at an existing nuclear facility⁸ in order to fulfill this direction.

The OPA's recommendation was based in part on an analysis of the environmental performance of different electricity supply options that was, in the view of observers ranging from the Pembina Institute9 to the City of Toronto's Medical Office of Health,¹⁰ flawed in terms of overall methodology. The authority's approach to weighting certain types of impacts, such greenhouse gas (GHG) generation, much more heavily than other impacts, such as waste generation and water pollution, and its failure to consider the justifiability of the transfer of risks and costs from current electricity consumers to future generations were the target of particular criticism. The analysis also overlooked major environmental impacts associated with nuclear power generation, such as waste generation and water consumption and water pollution associated with uranium mining and milling.11

The Canadian nuclear industry has recently put considerable effort into presenting itself as a 'clean' source of energy. A life cycle approach to the evaluation of risks, costs and benefits is particularly important in the case of nuclear energy. Key impacts may occur at locations other than the actual electricity generating facilities, and may last well beyond the time of power generation. A life cycle approach considers the full range of impacts across media and time, from the extraction and processing of fuel sources to the management of wastes resulting from fuel consumption and the decommissioning of extraction, processing and electricity production facilities. As such it provides a comprehensive basis for comparing the costs and risks associated with different energy technologies, and a better understanding of the trade-offs among the different types of impacts and risks that may occur.

The goal of the project, in this context, is to provide a comprehensive overview description of the major environmental impacts and risks associated with each stage of nuclear energy production in Canada, from uranium mining through to waste fuel disposal. Economic, community and occupational health costs and risks are noted where information on such impacts was available. A full economic costing of nuclear energy is beyond the scope of the study, in part due to the non-accessibility of key economic information with respect to the nuclear sector.

A comparison of the impacts of nuclear energy with those of other energy sources is also beyond the scope of the study. However, this study is intended to inform such comparisons in the future by providing as comprehensive as possible a picture of the risks and impacts associated with the use of nuclear energy for electricity generation in a Canadian context.

Implicit in this effort is an attempt to measure how much specific information on these issues is actually available in the public realm to support the completion of such an assessment, particularly in comparison to other energy sources. The identification of these information gaps is a significant sub-goal of the project.

Nuclear Power in Canada

As of September 2006, there are five commercial nuclear power generating stations in Canada: three in Ontario (Pickering, Bruce and Darlington), one in Quebec (Gentilly-2) and one in New Brunswick (Point Lepreau). All Canadian commercial nuclear generating stations operate Canada Deuterium Uranium (CANDU) reactors. CANDU reactors use un-enriched uranium as fuel, and heavy water (deuterium) as a moderator.

The Pickering Nuclear Generation Station was Canada's first large-scale nuclear power plant. The Pickering A station consists of four reactors that went into service in 1971–73. The Pickering B station consists of four reactors brought into service in 1983–86. At their peak, the Pickering A and B stations had a combined total capacity of 4,120 megawatts (MW).¹² Two Pickering A units are now permanently out of service.

The Bruce Nuclear Generation Station is located in Tiverton, Ontario. The four Bruce A reactors went into service in 1977–79 and the four Bruce B units entered service in 1985–87. The Bruce Power Ltd. consortium currently operates the Bruce Station.¹³ The station has a capacity of 6,140 MW,¹⁴ although two units are currently undergoing refurbishment.

The Darlington Nuclear Generation Station consists of four reactors commissioned in 1990–1993. It has a total generating capacity of 3,524 MW.¹⁵

The Gentilly-2 Nuclear Generation Station is located near Bécancour in Quebec and is operated by Hydro-Québec. The station consists of a single reactor that was declared in-service in 1983. Gentilly-2 has a total generating capacity of 675 MW.¹⁶

The Point Lepreau Nuclear Generation Station is located on the north shore of the Bay of Fundy in New Brunswick. It was declared in-service in 1983. Point Lepreau has a generation capability of 635 MW.¹⁷

1.2. Project Methodology

Nuclear energy differs from other energy sources in that many of its key impacts and risks can occur at locations other than the actual generating facilities, and well beyond the time of power generation. For this reason, the study considers the impacts of nuclear power at each of the four major phases of its production:

- Mining and milling of uranium ore into uranium oxide (U3O8)
- Refining U3O8 into uranium trioxide (UO3), conversion to uranium dioxide (UO2), and fabrication into pellets and assembly of fuel bundles (fuel processing).
- Nuclear power plant operation, where the fuel bundles are used in CANDU reactors to produce heat to fuel electricity generation.
- Waste fuel management.

Figure 1.1 provides a geographical overview of where activities in each phase of nuclear energy production occur in Canada. Uranium mine and mill sites currently operational in Canada are McClean Lake, Key Lake, Rabbit Lake and McArthur River, all of which are located in the Athabasca basin in northern Saskatchewan. There is a distance of over 4,000 km from the uranium mine and mill sites to the refining, fabrication and conversion facilities to the generating

stations in Ontario. Not displayed in Figure 1.1 are the generating stations located in Quebec and New Brunswick. Currently, Canada's generating stations also contain storage facilities for the waste nuclear fuel after it has been removed from the reactors.

1.3. Impacts Considered

The impacts of nuclear power production at each of its four phases were considered in terms of waste generation, atmospheric releases, water quality and quantity, land and ecosystem effects, and occupational and community health. Examples of the specific types of impacts examined are summarized in **Table 1.1**.

The findings related to each phase of nuclear energy production are presented in a separate chapter. The specific types of impacts considered in the study are outlined below.

1.3.1. Radioactivity and Radionuclides¹⁸

A radionuclide is an atom with an unstable nucleus. Radionuclides may occur naturally, but can also be artificially produced.

Four types of radiation are given off by radionuclides: alpha, beta and neutron particles and gamma rays. They are all hazardous, but they differ in their



Figure 1.1: Nuclear Energy Production in Canada

Waste Generation	Atmospheric Releases	Water Impacts	Landscape and ecosystem impacts	Occupational and Community Health
Waste fuel, and high-, medium- and low-level radioactive wastes Hazardous wastes High volume wastes such as waste rock from mining opera- tions	Radiation Radioactive contaminants Conventional air pollutants · Sulphur and nitrogen oxides · Particulate matter · Volatile organic compounds Hazardous air pollutants · Persistent organic pollutants · Heavy metals Greenhouse gases	Releases to surface water and ground- water of radioactive, conventional and hazardous contami- nants Water consumption Impacts on ground and surface water storage and flows Thermal impacts	Land footprint of facilities Impacts on affected biota and ecosystems	Occupational impacts from radiation exposure Conventional occupational risks Community health impacts of nuclear facilities including accidents

Table 1.1: Impacts of Nuclear Energy in Canada

power of penetration. While radioactive elements can give off two or more types of radiation, they generally give off only one.

Alpha particles are positively charged particles. They are the weakest form of ionizing radiation, and can be stopped by a sheet of paper, layer of skin or a few millimetres of air. However, if swallowed or inhaled, alpha particles can be extremely toxic.

Beta particles are fast-moving electrons. They can penetrate paper or skin, and can travel through a few centimetres of human tissue, but can be stopped by a few millimetres of metal.

Neutron particles are highly penetrating particles released by nuclear fission reactions. They can be stopped by thick shields of concrete or water.

Gamma rays are rays of energy somewhat similar to light rays, although they cannot be seen by the naked eye. Gamma rays can penetrate flesh, bone, and metal. It takes one metre of concrete or three metres of water to stop gamma rays.

Radiation loses energy as it passes through matter. The energy is transferred to and excites the atoms of materials it contacts, disturbing the way the material's electrons are arranged or causing the addition or loss of electrons (referred to as "ionizing radiations"). This may cause chemical changes that are harmful to living cells. Even small amounts of radiation can affect the chemistry of healthy cells, causing them to grow in an uncontrolled manner, producing a cancer. Alternatively their genetic structure may be altered, resulting in mutations in future generations. There is strong evidence that radiation has harsher effects on fetuses and young children than on adults.

The hazard to life and health depends on the length of exposure time, the amount of energy emitted by the radiation, and its ability to penetrate body issues. Short-term exposure to very high levels of radiation can cause burns and even death. The most penetrating form of radiation—gamma radiation—is most hazardous externally. Alpha and beta radiation can do less harm externally, but are extremely dangerous if inhaled or ingested and absorbed into particularly sensitive parts of the body, such as bone marrow.

The danger also depends on how quickly the radioactive material decays. Radionuclides with short half-lives release more energy in a shorter time than those with longer half-lives, causing more immediate chemical and biological changes. Radionuclides with long half-lives emit energy at a lower rate, but can be of more concern than those with short half-lives as they may persist over extremely long time periods—up to millions of years.

The International Agency for Research on Cancer (IARC) lists a number of radionuclides as carcinogenic to humans, including isotopes produced in uranium mining and milling, fuel production and nuclear power plant operations.¹⁹ It has been argued that existing standards for cancer risks arising from radiological hazards permit much higher levels of acceptable risk than is the case for chemical and other hazards.²⁰

More broadly, recent research on the effects of even very low levels of ionizing radiation suggests that no level is safe to health. The risk of cancer has been found to be greatest for women and children and to be higher for younger children.²¹

1.3.2. Generation of Low-, Intermediateand High-level Radioactive Wastes

Nuclear energy generation results in the production of low-, intermediate- and high-level radioactive wastes.

Low-level waste includes items that have become contaminated with radioactive material or have become radioactive through exposure to neutron radiation. This waste typically consists of contaminated protective shoe covers and clothing, wiping rags, mops, filters, reactor water treatment residues, equipment and tools, luminous dials, medical tubes, swabs, injection needles, syringes, and laboratory animal carcasses and tissues.²²

Low-level waste is further categorized according to source and includes wastes from fuel manufacturing, electricity generation, radioisotope production and use, nuclear research and, and historic low-level waste.²³ Low-level waste is considered by the Canadian Nuclear Safety Commission (CNSC) as safe enough to handle without any radiation protection.

Intermediate-level wastes are of a level where shielding is required to protect workers and can include ion exchange resins, filters, and irradiated core components.²⁴ Waste rock and tailings from uranium mining and milling operation can also be radioactive.

High-level radioactive wastes include waste fuel from reactors, and reactor components that may be removed as part of refurbishment or decommissioning projects.

1.3.3. Generation and Release of Hazardous Contaminants

Hazardous contaminants generated in the production of nuclear energy include the substances listed in **Table 1.2**.

1.3.4. Nutrients

Nutrients generated in the production of nuclear energy include nitrates and phosphorus. In most freshwater bodies, phosphorus is the primary nutrient that limits plant and algae growth. Excessive phosphorus and nitrate levels can lead to changes in numbers and types of plants, decline of oxygen levels in the water and increased buildup rates of dead organic material.^{29,30} High concentrations of nitrates in drinking water are associated with health impacts.³¹

1.3.5. Criteria Air Pollutants

Criteria air pollutants generated in uranium mining and milling, fuel production and nuclear power plant operations include sulphur and nitrogen oxides (SOx and NOx), particulate matter (PM) and volatile organic compounds (VOCs). SOx and NOx are important precursors for acid rain and smog. PM less than 10 µm in diameter is commonly referred to as inhalable or thoracic particles as it can penetrate into the thoracic compartment of the human respiratory tract. Such particles are known to cause human health impacts. In addition, particles 10 µm in diameter and smaller can scatter light and therefore generate atmospheric haze. SOx, NOx, respirable PM, and PM containing metals from certain sources are classified as toxic substances for the purposes of the Canadian Environmental Protection Act.³²

1.3.5.1. Volatile Organic Compounds (VOCs)

VOCs are smog precursors and can have significant hazardous properties of their own, including being recognized as carcinogens. VOCs participate in atmospheric photochemical reactions and a number of individual VOCs such as benzene, have been classified as toxic substances for the purposes of the Canadian Environmental Protection Act.³³

1.3.6. Greenhouse Gases

Greenhouse gases (GHGs) contribute to global climate change. These gases include carbon dioxide (CO₂), nitrous oxide (N₂O), ozone, methane (CH₄), hydrofluorocarbons, perfluorocarbons, and water vapour. The Intergovernmental Panel on Climate Change (IPCC) in 2002 stated that "there is new and stronger evidence that most of the warming observed over the last 50 years is attributable to human activities." GHGs have been classified as toxic substances for the purposes of the *Canadian Environmental Protection Act.*³⁴

1.3.7. Landscape and Water Impacts

The production of nuclear energy may result in landscape disturbance as a result of the construction and operation of mining, processing, energy production or waste management facilities. Facilities may also use or require the removal or use of large amounts of surface water and groundwater.

Substance Name	Canadian Environmental Protection Act (CEPA), 1999 Schedule 1 Status ²⁵	Affected Media	Comments ²⁶
Heavy Metals			
Arsenic	Toxic (inorganic arsenic compounds)	Water releases	Recognized carcinogen and developmental toxicant
Cadmium	Toxic (inorganic cadmium compounds)	Water releases, waste generation	Recognized carcinogen and development and reproductive toxicant
Chromium	Toxic (hexavalent chromium)	Water releases	Recognized carcinogen and suspected respiratory toxicant
Lead	Toxic	Water and atmospheric releases, waste generation	Recognized carcinogen and developmental and reproductive toxicant
Mercury	Toxic	Waste generation	Recognized developmental toxicant; wide range of other suspected toxic effects
Molybdenum	Not assessed	Water releases	Suspected reproductive and neurotoxicant
Nickel	Toxic	Water releases	Recognized carcinogen; wide range of other suspected toxic effects
Selenium	Not assessed	Water releases	Suspected cardiovascular and blood, devel- opmental, gastrointestinal or liver, kidney, musculoskeletal, reproductive, respiratory and skin toxicant; suspected neurotoxicant
Uranium	Toxic (uranium and uranium com- pounds contained in effluent from uranium mines and mills)	Water and atmospheric releases, waste generation	Recognized carcinogen; generally regarded as having greater potential to cause chemical rather than radiological toxicity ²⁷
Persistent Organic Po	llutants		
Dioxins and furans	Toxic	Atmospheric releases	Recognized carcinogens; suspected developmental toxicants
Hexachlorobenzene	Тохіс	Atmospheric releases	Recognized carcinogen and development toxicant

Table 1.2: Hazardous Contaminants Generated in the Production of Nuclear Energy in Canada

Fable 1.2: continued				
Other Hazardous Contaminants				
Asbestos	Toxic	Waste generation	Recognized carcinogen	
Ammonia	Toxic (gaseous ammonia and ammonia dissolved in water)	Water and atmospheric releases	Associated with eutrophication and acidification risks ²⁸	
Hydrazine	Not assessed	Water and atmospheric releases	Recognized carcinogen, and suspected reproductive, developmental, cardiovascu- lar, neurological and respiratory toxicant	
Hydrogen fluoride	Not assessed	Atmospheric releases	Suspected cardiovascular and blood, developmental, gastrointestinal or liver, musculoskeletal, reproductive, respiratory, skin or sense organ toxicant; suspected neurotoxicant	
Hydrogen sulphide	Not assessed	Atmospheric releases	Suspected cardiovascular and blood toxicant, neurotoxicant, reproductive and respiratory toxicant	

1.3.8. Occupational and Community Health

In addition to normal occupational risks associated with mining, industrial processes and plant operations, workers in the nuclear industry are permitted be exposed to much higher levels of radiation higher than that permitted for workers in most sectors and well above the accepted rate of exposure for the general public.

The *Nuclear Safety and Control Act* defines a nuclear energy worker as "a person who is required, in the course of the person's business or occupation in connection with nuclear substances or a nuclear facility, to perform duties in such circumstances that there is a reasonable probability that the person may receive a dose of radiation that is greater than the prescribed limit for the general public."³⁵ Nuclear energy workers are required to be informed of their status and the risks associated with exposure to radiation, and must confirm, in writing, that they have been made aware of their status and the risks.

Table 1.3 compares the radiation dose limits for nuclear energy workers and the general public. The figures are expressed in millisieverts (mSv), which is a measure of radiation dose.

As **Table 1.3** indicates, current regulatory occupational effective radiation dose limits for nuclear energy workers are 50 mSv in a calendar year and 100 mSv in a five-year block. The dose limit for pregnant nuclear energy workers is 4 mSv during the balance of the **Table 1.3:** Comparison of Radiation Dose Limits forNuclear Energy Workers and the General Public

	Nuclear Energy Worker (mSv)	General Public (mSv)
Dose per year	50	1
Dose per five years	100	n/a
Skin, hands and feet dose per year	500	50

Note: 'n/a' indicates that the measurement is not applicable.

pregnancy, while the dose limit for a member of the public is only 1 mSv per year.³⁶ Acceptable radiation limits for nuclear energy workers are ten to 50 times higher than for the general public.

Community health risks of nuclear energy production include exposure to routine releases of radiation, radionuclides and conventional pollutants from nuclear facilities, and the consumption of food or water containing radionuclides or other contaminants. Communities locate d near nuclear facilities or along transportation routes where nuclear materials or wastes are moved may be at risk due to releases of contaminants as a result of accidents.

1.4. Sustainability Challenges

The study also examines number of cross cutting issues that have major implications for the long-term sustainability of nuclear energy as a source of electricity. These issues include the following:

- Generating facility construction costs and timeframes.
- Generating facility performance, reliability and maintenance costs.
- The long-term security of fuel supplies and fuel costs.
- Security and weapons proliferation issues.

The focus of the study is on Canadian CANDU operations. CANDU reactors are the only type of reactor used for electricity production in Canada, and are therefore the most relevant to decision making with respect to nuclear power in Canada. Different reactor technologies are used in other countries, resulting in different types and levels of impacts. Light water reactors, which rely on enriched uranium fuel, for example, produce a lower volume of spent fuel waste than CANDU reactors, but are associated with much higher greenhouse gas emissions arising from the uranium enrichment process, particularly where gaseous diffusion processes are used to enrich the uranium.³⁷

The authors had originally hoped to quantify and normalize the impacts of nuclear power production on a per kilowatt-hour basis. A per kilowatt-hour approach would facilitate comparisons with other electricity sources. Unfortunately, due to significant gaps in the available data, this largely proved impossible, although per kilowatt-hour estimates are provided for some wastes and pollutants.

Information was collected from a variety of sources

CANDU Reactors

All Canadian commercial nuclear generating stations operate Canada Deuterium Uranium (CANDU) reactors. CANDU reactors use un-enriched uranium as fuel, and heavy water (deuterium) as the moderator.

To understand the basis for the design of the CANDU reactor requires a basic understanding of some reactor physics. A nuclear fission reaction is a chain-reaction in which a neutron collides with an atom and causes it to split, thus releasing more neutrons. Those neutrons in turn collide with other atoms and cause them to split. With every split atom, large amounts of heat are generated. This heat can be used to generate steam to rotate turbines, which generate electricity in a nuclear power plant. The only type of atom in nature that can sustain this type of chain reaction is an isotope³⁸ of uranium known as U-235. Natural uranium is comprised primarily of another isotope, U-238, which does not readily split in a fission reaction. U-235 makes up a very small percentage (0.7 per cent) of the total uranium mix.³⁹

One way to make uranium into fuel for a nuclear reaction is to increase the concentration of U-235 isotopes. This process, which has very high capital and operating costs, is known as uranium enrichment. It was developed in the United States during World War II to produce the highly concentrated U-235 required in atomic bombs.

Another way to sustain a nuclear reaction using natural uranium as fuel is to slow down the neutrons that cause the splitting in order to increase the probability that they will collide with a rare U-235 atom. This slowing down can be done with the help of a substance called a moderator. There are several substances that have the correct properties to moderate nuclear reactions. Two such moderators are deuterium (heavy water) and graphite.

During World War II, British, French and Canadian scientists were assigned the task of designing a deuterium-moderated nuclear reactor in Montreal under the direction of the western allies. After the war, the Canadian National Research Council (NRC) took over the project and the deuterium technology. Uranium enrichment facilities did not exist in Canada at the time and, due to their extreme expense and Canada's decision not to develop its own nuclear weapons program, there were not any plans to develop such facilities. Thus it was decided that existing Canadian technologies and manufacturing capabilities should be used. With the existing knowledge of deuterium moderation, the presence of uranium reserves and the lack of enrichment facilities, the CANDU reactor was developed.⁴⁰ including available literature and discussions with relevant and credible experts. Key sources of information included the Canadian Nuclear Safety Commission (CNSC), the National Pollutant Release Inventory (NPRI), and the operators of individual nuclear facilities, particularly Cameco and OPG. The exclusion of radionuclides and the exemption of stage one mining activities (i.e., extraction and primary crushing) from the NPRI significantly limits the usefulness of the inventory as an information source.

The authors are aware that estimating actual levels of hazard or risk posed by wastes, emissions and exposures associated with mining, fuel production, plant operation and waste disposal is often the subject of intense scientific debate. We have not sought to resolve these debates. Rather we simply highlight where they exist and leave it to readers to draw their own conclusions from the available literature.

Consistent with the sustainability principles "polluter pays" and "intergenerational justice,"⁴¹ particular note is given to situations where wastes or pollutants are generated that have the potential to transfer risks and impacts over time, beyond present consumers of the electricity and other benefits associated with their generation, on to future generations.

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2. Phase I: Uranium Mining and Milling

Summary of Key Findings

- The environmental impacts of uranium mining and milling are severe. They represent the most significant short-term environmental impacts of nuclear energy production in Canada. A number of jurisdictions in Canada and Australia have adopted bans on the establishment of new uranium mines due to concerns over the potential environmental and health impacts of such operations.
- The key impacts of uranium mining and milling include the following:
 - The generation of large quantities of waste rock and mill tailings. These are typically acidic or potentially acid generating, comprise long-lived radionuclides, heavy metals, and other contaminants.
 - Uranium mining milling to supply Canadian domestic power generation is estimated to result in the production of more than 90,000 tonnes of tailings, and up to 2.9 million tonnes of waste rock per year.
 - Canadian uranium mines and mills have an inventory of 109 million tonnes of waste rock, and 214 million tonnes of tailings.
 - There are major concerns regarding long-term integrity of tailings and waste rock containment facilities. These facilities will require perpetual care. The adequacy of current financial assurances required by governments for the closure and long-term care of containment facilities has been questioned.
 - Severe contamination of groundwater with radionuclides, heavy metals, and other contaminants has occurred at tailings management facilities and waste rock storage areas.
 - Uranium mining and milling facility surface water discharges have resulted in the contamination of the receiving environment with radionuclides and heavy metals. Effluent from historic and operating uranium mines and mills, particularly uranium discharges, have been determined to be toxic for the purposes of the *Canadian Environmental Protection Act* by Environment Canada and Health Canada.
 - Uranium mining operations are associated with the extensive removal of groundwater (in excess of 16 billion litres per year).
 - The environment and biota in the vicinity of uranium mines and mills has been contaminated with radionuclides particularly via windblown dust from tailings sites. Significant potential increases in cancer risks to humans from the consumption of caribou in the vicinity of uranium mines have been identified.

- Uranium mines and tailings storage areas have been identified as significant sources of atmospheric releases of radon gas.
- Major atmospheric releases of sulphur dioxide and VOCs are associated with the uranium milling process. In 2004, VOC emissions from the sector were equivalent to the average annual emissions of more than 300,000 cars. The Rabbit Lake facility acid plant reported releases of 43,000 tonnes of SO₂ in 2004.
- Atmospheric releases of NOx and PM result from the milling process and the operation of fossil fuel-powered machinery and equipment.
- Annual CO_2 emissions resulting from uranium mining, milling and tailings management activities in Canada are estimated at between 160,000 and 250,000 tonnes.
- The mining of lower grade ores would result in the generation of proportionally larger amounts of tailings, other wastes and emissions, as larger amounts of ore would have to be processed to produce the same amount of uranium concentrate. Processing of ore that is 0.01% uranium, for example, would generate approximately ten times the tailings of ore that is 0.1% uranium.

Workers at uranium mines and mills typically receive annual effective radiation doses higher than those considered acceptable to members of the general public. Increased incidences of lung cancer as well as deaths resulting from silica exposure are reported among uranium miners.

2.1. Introduction

The mining and milling of uranium is the first step in the production of nuclear energy. Canada is currently the world's largest uranium producer, extracting uranium from four Saskatchewan mines: McClean Lake, Key Lake, Rabbit Lake and McArthur River. Uranium milling occurs on-site at each mine with the exception of McArthur River, which trucks its uranium ore 80 km to Key Lake for milling. While all current and proposed uranium mining and milling operations are based in northern Saskatchewan, there are potentially developable uranium reserves in a number of Canadian provinces and territories. Historically, there have been uranium mines in both Ontario and the Northwest Territories.

Table 2.1 lists Canadian mining and milling sites, and their respective operational state as reported by NRCan.

The historical environmental and health impacts of uranium mining and milling in Canada have been severe. The effects have included the extensive contamination of surface water, groundwater and the surrounding environment in the vicinity of facilities with radioactive, toxic, and conventional pollutants, and the creation of major occupational health concerns. Although major uranium mining operations began in Ontario in the 1950s, for example, occupational health and safety requirements in the province were not fully established until 1984.²

As a result of such negative impacts, some Canadian provinces have imposed moratoriums on uranium

exploration and mining. After significant uranium exploration took place in Nova Scotia in the late 1970s, that province imposed a moratorium on uranium exploration and mining in 1981. An inquiry in 1985 recommended that the moratorium be renewed for another five years. An interdepartmental committee on uranium was then established, releasing a final report in 1994. The moratorium remains in place.³

In British Columbia, a seven-year moratorium was imposed in 1980 after a royal commission concluded that health risks associated with uranium mining made it too dangerous. The moratorium expired in 1987 and was not re-instated. No uranium mines have been established in British Columbia since then, although some exploration activity is occurring.⁴ A number of Australian states also have bans in place on the establishment of new uranium mines.⁵

The approval of uranium mining and milling projects has continued to be a source of major controversy. In 1993, the joint federal-provincial environmental assessment panel examining the then proposed McClean Lake mine recommended that approval of the facility be delayed for five years to permit a better understanding of the likely performance of the proposed tailings management facilities, community health and social impacts, and the cumulative biophysical and socio-economic impacts of the project.⁶ The Atomic Energy Control Board's (now Canadian Nuclear Safety Commission) subsequent decision to approve the McClean Lake tailings facility was successfully challenged by the Inter-Church Uranium

Producing Operations	Projects under Development	Past Producing Operations
Rabbit Lake (Northern SK)	Midwest (Northern SK)	Cluff Lake (Northern SK)
Key Lake (Northern SK)	Cigar Lake (Northern SK)	Port Radium (Port Radium, ON)
McClean Lake (Northern SK)	Kiggavik (Baker Lake, NWT)	Agnew Lake (Espanola, ON)
McArthur River (Northern SK)		Madawaska et al. (Bancroft, ON)
		Rayrock (Marian River, NWT)
		Beaverlodge et al. (Uranium City, SK)
		Quirke/Panel/Denison and Stanleigh et al. (Elliot Lake, ON)
		Gunnar and Lorado et al. (Uranium City, SK)

Table 2.1: Overview of Canadian Uranium Mining and Milling Operations¹

Committee Educational Cooperative (ICUCEC) before the Trial Division of the Federal Court of Canada. The challenge was, however, overturned on appeal to the Federal Court of Appeal. The Supreme Court of Canada declined the ICUCEC leave to appeal the Court of Appeal's decision in April 2005.⁷

2.2. Uranium and Milling Process: An Overview

The mining procedure for uranium depends on the depth of the uranium below ground level. If the deposit is relatively close to the surface (i.e., less than 100 m below ground level) an open pit mine is used to extract the ore. Any rock or overburden (i.e., overlying soil, surface features and biota) on top of the orebody is removed to permit access to the ore.⁸

If the uranium deposit lies more than 100 m below ground level, access to the deposit is gained by digging vertical shafts to the depth of the orebody. Once mined, the ore may be transported directly to the milling facility, or, as at the McArthur River mine, crushed in underground facilities to the consistency of fine sand, diluted with water and pumped to the surface as a slurry or mud.

At the mill, the ore is ground to a very fine consistency and mixed in either a highly acidic or alkaline solution, depending on its chemical characteristics, to extract it from the rock particles. Finally, the uranium is concentrated using an ion exchange process or solvent extraction and precipitated and dried into mixed uranium oxides (U_3O_8) called "yellowcake."

Tailings, which are the sand-like materials left over from the milling process, consist of ground rock particles, water and mill chemicals, and include radioactive and hazardous constituents. Tailings are stored at a tailings management facility (TMF).

During the mining and milling process, fossil fuels are used to fuel machinery for earth moving, transportation, heat and steam production, and electricity generation.

2.3. Uranium Mining and Milling Impacts

Radioactive, hazardous and other wastes, emissions and discharges are generated at each stage of the mining and milling process. Uranium mining and milling also involves significant disruptions of the surface landscape and of surface water and groundwater flows, both as a direct result of mining activities, and as a result of the construction and operation of waste rock and tailings storage facilities. The key impacts of uranium mining and milling are summarized in **Table 2.2**. The details of these impacts are discussed in the following sections.

2.3.1. Waste Generation

Both open pit mines and shaft mines create very large quantities of waste rock and tailings. These wastes are major sources of the biophysical impacts of uranium mining.

2.3.1.1. Waste Rock

Waste rock consists of any rock and overburden (i.e., overlying soil, biota and surface features) removed to permit access to the ore. The amount of waste rock generated depends on the type of mine. In an open pit mine, 40 tonnes of waste rock or more can be generated for every tonne of uranium ore extracted. In an underground mine, to access the same amount of ore may generate one tonne of waste rock or less.9 Approximately 75 per cent of Canadian uranium ore is obtained from open pit operations.¹⁰ In 2003 the Canadian uranium industry processed 587,000 tonnes of uranium ore,¹¹ suggesting waste rock generation of up to 18 million tonnes. Domestic reactor fuel requirements accounted for approximately 16per cent of total Canadian uranium production in 2003,12 suggesting associated waste rock production of up to 2.9 million tonnes.

Waste rock may contain both radionuclides (principally U_3O_8) and other heavy metals, such as nickel, copper, arsenic, molybdenum, selenium and cadmium. In addition, waste rock may contain sulphur materials, which will oxidize under weathering and bacterial action, forming sulphur oxides that mix with water to form acids. These acids can dissolve metals, thus mobilizing uranium and other heavy metals (e.g., copper, arsenic and cadmium) in the waste rock and other locations at the mine site that are harmful to humans and wildlife. The resulting discharges can acidify surface water and groundwater, and contaminate the water with heavy metals, including radionuclides. Acid mine drainage can persist for centuries.¹³

In addition, uranium ore contains radionuclides such as thorium and radium, which are intermediates in the uranium decay chain. One of the decay products of radium is radon gas, which emanates from the rock as radium and thorium decay. As a result, overburden and waste rock storage areas can be sources of releases of radon gas,¹⁴ as well as wind blown dust containing radionuclides, heavy metals and particulate matter.

Waste rock is managed at Canadian uranium mines according to its quality. Some of it is consid-

Occupational and Community Health	Occupational Health
Landscape and Ecosystem Impacts	 Land footprint of mines, facilities, tailings and waste rock storage facilities Increased concentration of radioactive material and heavy metals in flora, fauna and food chain near mines
Water Impacts	 Groundwater contamination from tailings and waste rock sites Surface water contamination from discharges of mine- water and process waters, and surface drainage from tailings, waste rock, and general run-off Disruption of surface water and groundwater flows Discharges from milling pro- cesses
Atmospheric Releases	 Radon in ventilation air being discharged from underground mines Radon releases from surface mining VOCs, radionuclides, PM, NOx and SOx emissions from milling operations and acid plants GHGs, particulates, NOx and SOx emissions from equipment and vehicles, milling and tailings manage- ment activities. PM, heavy metals and radio- nuclides in dust from surface mine sites, waste rock and tailings areas
Waste Generation	 Waste Rock Large volume May contain radioactive and hazardous contaminants Risk of acid mine drainage resulting in acidification of surface water and radioactive contaminants Dust contamining radionu-ventional and radioactive contaminants Dust contamining radionu-clides, heavy metals and PM Radon releases from waste rock May require perpetual care Contain radionuclides and hazardous contaminants Risk of ground water and surface water contaminants Risk of ground water and surface water contaminants Risk of catastrophic tailings radionuclides, heavy metals, and PM

ered "clean waste" (non-acid generating waste, less than 0.03per cent U_3O_8) and is used for surface work or backfill aggregate. Potentially acid generating waste rock (less than 0.03per cent U_3O_8) is stored on lined waste pads and disposed of at the bottom of flooded mine pits to prevent it from oxidizing.¹⁵ Mineralized waste (greater than 0.03per cent U_3O_8 but less than 2.0per cent U_3O_8) is used as blend materials for the milling feed stream.¹⁶

As of December 2001, there were 3.7 million tonnes of mineralized waste rock and 105 million tonnes of non-mineralized waste rock in inventory at the McClean Lake, Key Lake, Rabbit Lake, Cluff Lake, McArthur River and Cigar Lake sites.¹⁷

2.3.1.2. Tailings

Ore mined in open pit or underground mines is crushed and leached in a uranium mill, a chemical plant designed to extract uranium from the ore. Tailings are the waste by-product of the milling process. The amount of tailings produced is proportional to the grade and amount of the ore milled. At a grade of 0.3per cent uranium, for example, 99.7per cent of the material is left over.

Tailings consist of ground rock particles, water, and mill chemicals, and contain radioactive and hazardous constituents. Up to 85per cent of the radiological elements, contained in the uranium ore end up in the tailings.¹⁸ The following radionuclides may be present: polonium-210, bismuth-210, lead-210, polonium-214, bismuth-214, lead-214, polonium-218, radon-222, radium-226, thorium-230, uranium-234, protactinium-234m, thorium-234, and uranium-238.¹⁹ Radium-226 and uranium are considered the most significant radionuclides present in Canadian uranium mill tailings.²⁰

Radon gas will be released from the tailings if the containment area is not constructed properly. Due to the continuous production of radon from the decay of radium-226, which has a half-life of 1,600 years, radon presents a long-term hazard. Further, because the parent product of radium-226—thorium-230 (with a half-life of 80,000 years)—is also present, there is continuous production of radium-226. After about one million years, the radioactivity of the tailings and thus its radon emanation will have decreased so that it is only limited by the residual uranium contents, which continuously produce new thorium-230.²¹

In addition to radionuclides, tailings may contain heavy metals (e.g., arsenic, molybdenum, selenium) and residual mill process chemicals.²² The tailings may also be acidic as a result of the presence of sulphidic ore or chemicals introduced through milling, with the result that they present a risk of acid drainage to surface water and groundwater. Dust from tailings areas may contain radionuclides (e.g., uranium, radon-226 and lead-210, and other decay products), heavy metals (e.g., arsenic), PM and other contaminants.²³

Both the radioactive and chemically hazardous materials contained in uranium mill tailings need to be managed for an indefinite amount of time, with the radioactive material carrying risks that are more long term. As a result, tailings containment areas are considered perpetual environmental hazards.²⁴

At all of the newer operations in Saskatchewan, tailings are managed in mined-out pits converted to TMFs, which use hydraulic containment during operation. The pits are maintained in a partially dewatered state relative to the surrounding natural water table so that all groundwater flow is towards the TMF. Currently active TMFs are the Deilmann TMF at Key Lake operation, the Rabbit Lake In-Pit TMF at Rabbit Lake operation, and the Jeb TMF at McClean Lake operation.

The TMFs have treatment facilities that receive contaminated water feeds, remove dissolved metals, radionuclides and suspended solids, and discharge the treated water to a Treated Effluent Management System (TEMS).²⁵ The sludge produced is pumped directly to the tailings preparation circuit in the mill and disposed of in the TMF. The TEMS also receives clean groundwater from dewatering facilities at the TMF. The effluent from the TEMS is discharged into nearby water bodies.

Specific information on the contents and quantities of tailings being generated by Canadian uranium mining operations was sought from the relevant government agencies and individual facility managers, without success. Transfers for materials to tailings facilities are not reported by uranium mining companies to the NPRI. Natural Resources Canada reports that, in 2003, the Canadian uranium mining industry processed 587,000 tonnes of uranium ore, with an average grade of 18 kg of uranium per tonne (1.8per cent),26 suggesting tailings production in excess of 575,000 tonnes. Domestic reactor fuel requirements accounted for approximately 16 per cent of total Canadian uranium production in 2003,²⁷ suggesting the production of approximately 92,000 tonnes of tailings to fuel Canadian reactor operations in that year. The processing of lower grade ores would result in the generation of proportionally larger amounts of tailings, as larger amounts of ore would have to be processed to produce the same amount of uranium concentrate. Processing of ore that is 0.01% uranium,

for example, would generate approximately ten times the tailings of ore that is 0.1% uranium.

As of December 2003, there were 213 million tonnes of uranium mill tailings in storage at 24 tailings sites across Canada. This is enough material to fill the Toronto Rogers Centre (formerly the Sky-Dome) approximately 100 times.²⁸

The majority of these tailings are located at inactive or decommissioned operations in Ontario (82.8per cent), the last of which was closed in 1996. A small portion (0.5per cent) is located in the Northwest Territories at one decommissioned (1984) and another inactive (1959) operation. The remaining portion is located at active, inactive and decommissioned operations in Saskatchewan (16.7per cent).²⁹

2.3.2. Atmospheric Releases

Atmospheric releases of contaminants arise from many sources during the uranium mining and milling process. Dusts containing radionuclides, heavy metals and PM can be released from underground ventilation systems, waste rock and tailings storage areas, surface mining operations and milling operations. Radon gas may also be released from these sources. Milling operations produce releases of NOx, VOCs, CO₂ and PM. Acid plants producing acid for milling operations release large amounts of SO₂. The combustion of fossil fuels to operate equipment and vehicles for earth moving, transportation, heat and steam production and electricity supply generates releases of criteria air pollutants and GHGs.

2.3.2.1. Radionuclides

Radon gas is released as radium and thorium decay as part of the uranium decay chain. For underground mines, the release of radon has been estimated to be between 1 and 2,000 GBq/t of U_3O_8 produced, with a production average of 300 GBq/t.³³ Ventilation systems for underground mines have been identified as the leading source of radon emissions from uranium mining operations. Active open pit operations also produce significant atmospheric releases of radon.³⁴ Releases from mills have been estimated to be 13 GBq/ t of U_3O_8 produced.³⁵ Additional releases of radon occur from waste rock, tailings and ore storage areas.

Normal background radon concentrations in Saskatchewan are considered to be in the range of 37 Bq/ m^3 to 74 Bq/ $m^{3.36}$

The Saskatchewan Environment Cumulative Environmental Monitoring program monitors radon levels

Uranium Mine Tailings Management Failures

Uranium tailings management in the 1950s and 1960s was considered similar to other non-radioactive tailings management. The poor understanding of radiological and toxicological risks was clearly demonstrated by such practices as the use of uranium tailings as a building material in nearby cities.

Historical tailings practice exclusive to Canada included the disposal of tailings directly into deep lakes without controls or treatment. This includes the deep discharge from Port Radium (Northwest Territories) into Great Bear Lake, from Gunnar mill into Mudford Lake, and from Beaverlodge mill into Fookes/Marie Lakes. This practice is no longer permitted. In other cases, tailings were either disposed of as backfill in underground mines or placed in natural containment areas such as lakes and valleys and confined by permeable or water-retaining dams. Surface tailings were left bare or covered with soil and vegetated or flooded.¹⁵

As tailings management awareness grew, the complexity of tailings management increased as well. Tailings containment facilities were engineered and built, however, these were not without their own problems. Emissions of radioactive and non-radioactive substances occurred by several mechanisms: radon emanation, dust dispersal, and leaching into surface water and groundwater. Tailings discharges to water can occur when bottom-liners are absent and drainage waters are not collected from TMFs (dams and structures), or where these structures fail.³⁰

Examples of uranium mill tailings containment failures include dispersal of radioactive dust (Colorado, Germany, Hungary, Bulgaria, Kyrgyzstan, Australia, Kazakhstan), erosion of tailings containment (Australia, Germany), seepage through floors/walls (Hungary, Germany, Kyrgyzstan), effluent discharge (Rabbit Lake, Saskatchewan and Elliott Lake, Ontario, where effluent discharge resulted in significant radiological contamination and acidification of the 300 km² Quirke Lake) ³¹ and tailings dam overflows (Key Lake, Saskatchewan, 1984).³²

at eight stations in the Wollaston Lake area and has found that, for the most part, these levels have been continually low at an average of less than 30 Bq/m³ at each station over an eight-year period (1994–2002.) That being said, during this same period, three measurements of radon were over 100 Bq/m³, and two of these were at a considerable distance from the uranium mines. The accuracy of the system used to measure the radon concentrations (the track-etch cup system) has been questioned.³⁷

In addition to radon-226, other radionuclides, particularly uranium and lead-210, may be released in dust from waste rock, tailings and ore storage sites, ventilation system discharges and milling operations.³⁸ The Key Lake case study, highlighted below, demonstrates the extent to which the surrounding biota and environment can become contaminated as a result of windblown dust from a TMF.

2.3.2.2. Hazardous Air Pollutants

In addition to radionuclides dust from waste rock, tailings and ore storage sites, ventilation system discharges and milling operations may contain hazardous contaminants, particularly heavy metals. Uranium itself is regarded as having greater potential

Table	2.3:	Releases	of VOCs

	Rabbit Lake Operation (tonnes)	Key Lake Operation (tonnes)	Total (tonnes)
2004	36.1	1118.0	1154.1
2003	95.6	400.1	495.7
2002	45.0	416.1	461.1

to cause chemical than radiological toxicity.³⁹

Uranium milling results in significant fugitive airborne releases of VOCs. These substances can act as smog precursors, have significant hazardous properties of their own, and are classified as toxic substances under the *Canadian Environmental Protection Act*. **Table 2.3** lists annual releases of VOCs reported to the NPRI from uranium mine and milling operations in Canada. The high levels of releases, principally attributed to fugitive and stack sources at the Key Lake facility, are likely a result of the milling operation located there. The McClean Lake operation did not report any releases of VOCs.

The total VOC emissions from uranium mining and milling in 2004 is roughly equivalent to the annual VOC emissions from 302,000 cars.⁴⁰

2.3.2.3. Sulphur Dioxide, Nitrogen Oxides and Particulate Matter

 SO_2 and NOx are important precursors for acid rain and smog. PM less than 10 μ m in diameter is commonly referred to as inhalable or thoracic particles as it can penetrate into the thoracic compartment of the human respiratory tract. Such particles are known to cause human health impacts. In addition, particles 10

> µm in diameter and smaller can scatter light and therefore generate atmospheric haze. All of these substances are classified as toxic substances under the *Canadian Environmental Protection Act*.⁴¹

Sulphur Dioxide (SO₂)

Table 2.5 lists annual releases of SO_2 reported from the Key Lake mill for 2002–2004. These were the only reports of SO_2 releases

Case study: Fugitive tailings dust at Key Lake Mine

In a 2000 study of the ecosystem effects of uranium mining and milling, measurements for uranium-series radionuclides were taken at three sites near the Key Lake uranium mine in Northern Saskatchewan. The Key Lake Mine has an above-ground TMF, which was designed to minimize groundwater contamination,¹ yet it releases fugitive tailings dusts to the wind.² Compared to a control site, the sites impacted by windblown tailings and mill dust were found to have significantly higher concentrations of uranium and uranium-series radionuclides (Ra-226, Pb-210 and Po-210) in soils, litter, vegetation, tree needles, twigs, small mammals, and birds. Absorbed doses in small animals were highest at the tailings-impacted site, followed by the mill-impacted site. The study also examined atmospheric deposition rates of uranium series radionuclides, concluding that dry deposition was a more important transport mechanism for uranium, Ra-226 and Pb-210 than rainfall.³

	Key Lake Operation (tonnes)
2004	30.1
2003	32.5
2002	34.7

Table 2.5: Releases of SO₂ from Key Lake Operation⁴²

from uranium mining and milling to the NPRI.

The Rabbit Lake mine's 2004 annual environmental report states that 2004 SO_2 emissions from the Rabbit Lake facility's acid plant totalled 43,815 tonnes.⁴³

Nitrogen Oxides (NOx)

Table 2.4 lists annual airborne releases of NOx from uranium mining and milling operations for 2002–2004. Emissions were not reported for the McClean Lake operation.

Total NOx emissions from uranium mining in 2004 were roughly equivalent to the annual NOx emissions from the exhaust of 106,000 cars. 45

Particulate Matter (PM)

Uranium mining and milling activities release airborne PM from a variety of sources, particularly dust from mine operations and ventilation systems, tailings, waste rock and ore storage areas, and emissions from diesel engines. These particles, suspended in the air, are associated with many negative impacts on human health.

Table 2.6 lists annual releases of PM from uraniummining and milling operations for 2002–2004.

The total releases of PM from uranium mining and milling in 2004 were roughly equivalent to the PM released annually from the exhaust of 26,550 cars.⁴⁷

2.3.2.4. Greenhouse Gas (GHG) Emissions

GHGs contribute to global climate change. These gases include carbon dioxide (CO₂), nitrous oxide (N₂O), ozone, methane (CH₄), hydrofluorocarbons, perfluorocarbons, and water vapour. The Intergovernmental Panel on Climate Change (IPCC) in 2002 stated, "there is new and stronger evidence that most of the warming observed over the last 50 years is attributable to human activities."⁴⁸

A 1998 study based on 1996 data from uranium mines in Canada concluded that 12.1 tonnes of CO_2 are released for every tonne of uranium concentrate

Table	2.4: Release	s of NOx from	Uranium	Mining and	Milling O	perations ⁴⁴
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	Rabbit Lake Operation (tonnes)	Key Lake Operation (tonnes)	McArthur River Operation (tonnes)	Cluff Lake Project (tonnes)	Total (tonnes)
2004	47.7	72.2	55.7	220.7	396.3
2003	49.0	75.5	80.8	192.9	398.2
2002	-	97.8	32.8	382.7	513.3

Note: '-' indicates that the data was not available.

Table 2.6: Releases of PM from Uranium Mining and Milling Operations⁴⁶

	Rabbit Lake Operation (tonnes)	Key Lake Operation (tonnes)	McArthur River Operation (tonnes)	Cluff Lake Project (tonnes)	McClean Lake Operation (tonnes)	Total (tonnes)
2004	1.6	3.4	2.8	1.9	10.1	19.8
2003	1.8	3.5	5.0	1.4	8.1	19.8
2002	0.5	5.2	1.3	2.9	-	9.9

produced. This figure was based on the fact that certain mines obtain power from nearby hydropower dams, while others are too remote and must rely on on-site diesel generation. If all of the power consumed at these mines had been generated via fossil fuels (diesel), this figure would increase to 20.7 tonnes of CO_2 for every tonne of uranium concentrate produced. 49 During the milling process, there are additional CO_2 releases from acid leaching of the ore, which contains carbonate, and the use of lime, which is used to neutralize the tailings. In total, milling releases an additional 3.2 tonnes of CO₂ for every tonne of uranium concentrate produced.⁵⁰ These estimates do not appear to include the emissions associated with the transportation of ore 80 km by truck from the McArthur River mine to the Key Lake site for milling.⁵¹

On the basis of these estimates the 10.5 kilotonnes of uranium concentrate produced by Canadian uranium mines and mills in 2003 would have resulted in the release of between 160 and 250 kilotonnes of CO₂.⁵² This is roughly equivalent to the annual GHG emissions from 71,495 cars driving an average of 15,000 km /year.⁵³ The mining and milling of lower grade ores would require larger energy inputs as a larger volume of ore needs to be mined and processed to produce the same amount of uranium concentrate. The result is proportionally higher levels of emissions of greenhouse gases and other pollutants.⁵⁴

2.2.3. Water Impacts

Uranium mining and milling releases contaminants to groundwater and surface water through discharges of process and mine waters, leaching from TMFs and waste rock storage sites, and general run-off from mine sites. Releases can include radioactive conventional pollutants (e.g., total suspended solids) and hazardous pollutants (e.g., heavy metals), as well as acid drainage. Mining and milling operations can also disrupt surface water and groundwater features and flows.

In 2004 Health Canada and Environment Canada concluded that "releases of uranium and uranium compounds contained in effluent from uranium mines and mills are entering the environment in quantities or concentrations or under conditions that have or may have a harmful immediate or long-term effect on the environment or its biological diversity" and are therefore "toxic" as defined in s. 64 of the *Canadian Environmental Protection Act*, *1999*.⁵⁵

2.2.3.1 Groundwater Contamination

Groundwater contamination can occur as a result of leachate seeping into the ground from waste rock or TMFs. The tailings may be acidic as a result of the presence of sulphidic ore or chemicals introduced through milling, with the result that they present a risk of acid drainage to surface water and groundwa-

Case Study: Groundwater Quality Impacts at Cluff Lake Mine

The Cluff Lake project is the first modern uranium mine decommissioning project in Canada. The mine opened in 1980 and had accumulated 3.28 million tonnes¹ of tailings by the time it closed in 2002. The tailings are stored in an above-ground facility behind a dam.

Groundwater chemistry was monitored and compared to baseline (unimpacted) data as part of the Environmental Assessment for the Cluff Lake Mine decommissioning project. Adverse groundwater quality has been measured from releases of leachate from the tailings management area (TMA), backfill materials and waste rock piles. Decommissioning of the pits and mining areas is expected to have further impacts, as hydraulic containment (i.e., pumping) in these areas will no longer be used.

Groundwater in areas near the decommissioned TMA at Cluff Lake Mine was found to have elevated concentrations of major ions (potassium, magnesium, bicarbonate, chloride, calcium, sodium, sulphate) ranging from 10 to 200 times the levels in unimpacted groundwater, respectively. Concentrations of arsenic, cobalt, iron, lead, manganese, molybdenum, nickel, radionuclides Pb-210, Po-210 and Ra-226, uranium and vanadium were also found to be elevated from baseline data.²

Groundwater near the mine areas and waste rock storage facilities at Cluff Lake Mine was found to have significantly higher concentrations of major ions (chloride, potassium, calcium, sodium, magnesium) than were found in unimpacted groundwater. The concentrations ranged from 4.5 times higher (chloride) to 135 times higher (magnesium). The sulphate concentration of groundwater near the facilities was 1,450 times higher than that of unimpacted groundwater. Concentrations of arsenic, copper, lead, manganese, molybdenum, nickel and zinc were also found to be elevated from baseline data, with arsenic concentrations 66 times higher, manganese 1,100 times higher, and nickel 1,250 times higher.² ter. Waste sulphidic rock may also be subject to acid generation and drainage.

In addition to increasing the acidity of leachate, the presence or formation of acid in tailings or waste rock can help mobilize radionuclides and other heavy metals.

The experience with the Cluff Lake mine TMF illustrates the extent of the potential for groundwater contamination at uranium mine TMFs.

Concern over groundwater contamination prompted the Inter-Church Uranium Committee Educational Co-operative (ICUCEC), a watchdog group based in Saskatchewan, to take the CNSC to court in 2002 over the operational license for the Jeb tailings pit at Cogema's McClean Lake operation.⁵⁶ The ICUCEC argued that the license, issued in 1999, should have been assessed under the 1992 *Canadian Environmental Assessment Act.* The license had been granted on the completion of an eight-year environmental assessment, which began in 1991, under an older regulation. A federal court judge ruled in favour of ICUCEC, which invalidated Cogema's license to operate the McClean Lake mine. However, Cogema was able to obtain a stay on the ruling and continued to operate the mine and use the tailings pit. In June of 2004, Cogema's appeal of the ruling was heard and the ruling was overturned and the McClean Lake operating license was re-instated.⁵⁷ The Supreme Court of Canada declined the ICUCEC leave to appeal the Court of Appeal's decision in April 2005.⁵⁸

2.2.3.2. Discharges to Surface Water

Surface water discharges can occur from mine water, process waters, and surface drainage from tailings and waste rock as well as general run-off from the mine site.

Tables 2.7–2.10 presents mass loadings of contaminants to the environment (for which data is available) via wastewater treatment plant effluent from the McArthur River, and Rabbit Lake, Key and McClean Lake mines including collected leachate from TMFs.

Ammonia is also released to surface water from uranium milling. **Table 2.11** lists annual releases of ammonia to surface water from Saskatchewan uranium mills between 2001 and 2003. Ammonia dissolved in water is classified as a toxic substance under the *Canadian Environmental Protection Act*.⁶³

Most effluent outlets discharge into small tributaries and ponds, which act as mixing zones to dilute

Table 2.7:

	Mass	Loadings to	the Er	nvironment via	Wastewater	Treatment	Plant	Effluent	from	McArthur	River	Mine ⁵
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	2004	2003	2002	2001	2000
Total Suspended Solids (TSS) (kg)	3,425	9,113	3,240	2,914	2,240
Lead-210 – MBq	73.6	267.2	213.8	158.7	79.4
Radium-226 – MBq	147.8	739.3	223.6	160.5	36.0
Thorium-230 – MBq	28.7	158.7	156.7	58.3	29.0
Uranium (kg)	62.8	145.4	146.2	53.4	207.3
Arsenic (kg)	6.2	7.3	3.0	9.2	3.3
Copper (kg)	2.8	7.4	2.4	2.3	2.2
Nickel (kg)	5.3	7.3	2.6	2.5	3.0
Lead (kg)	4.9	10.7	4.5	4.3	3.2
Selenium (kg)	5.8	11.1	8.2	8.5	3.8
Vanadium (kg)	3.0	5.5	8.6	6.4	1.9
Zinc (kg)	17.8	47.4	26	29.1	28

Table 2.8:

Mass Loadings to the Environment via Wastewater Treatment Plant Effluent from Rabbit Lake Mine60

	2004	2003	2002	2001	2000
TSS (kg)	5,300	3,900	2,700	4,400	9,300
Uranium (kg)	1,230	1,430	1,860	2,470	3,190
Arsenic (kg)	40	60	30	90	230
Copper (kg)	8	6	4	10	10
Nickel (kg)	70	120	80	150	400
Lead (kg)	Below Detection	Below Detection	Below Detection	Below Detection	Below Detection
Zinc (kg)	20	20	30	20	20

Table 2.9: Mass I	Loadings to	the Environme	nt via Effluent	t from Ke	y Lake Mine	Mill ⁶¹
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	2004 (MMER Method)	2003 (MMER Method)	2002	2001	2000	1999
TSS (kg)	2593	3390.5	3545.1	2620.7	2192.2	3,091.1
Radium-226 (Bq)	73,488,445	83,504,0900	137,865,000	163,704,050	248,769,500	11,760,700
Lead-210	71,107,860 (previous method	87,948,000 (previous method)	78,780,000	92,463,130	92,463,000	35,504,000
Thorium-230	272,580,130 (previous method)	337,134,000 (previous method)	59,085,000	42,370,460	52,836,000	46,599,000
Molybdenum	1,252.6	862.2	2,127.1	1,117	3,940.7	7,300.5
Nickel	45.1	58.6	59.1	65.6	114.5	184.2
Selenium	33.2 (previous method)	30.8 (previous method)	27.6	46.2	39.6	117.5
Zinc	7.9	13.5	9.8	28.9	24.2	13.3
Vanadium	5.9 (previous method)	7.3 (previous method)	9.8	25	110.01	<11.1
Lead	17	15.7	19.7	<15.4	22	<22.2
Uranium	11.6	10.4	13.8	12.9	10.6	17.8
Copper	9	10.3	<9.8	7.7	24.2	17.8
Arsenic	10.1	9.4	5.9	5.8	30.8	146.5

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the effluent before it moves into the main water body. This method of dealing with toxic effluent can be considered outdated and problematic. Reliance on these tributaries as mixing zones means that the degree of dilution will vary seasonally. It also creates localized contamination impacts and doesn't take into account cumulative loadings, especially of persistent pollutants like heavy metals. While modern regulations require that concentrations of these substances remain within water quality guidelines, a number of recent studies indicate total loadings of these substances over the life of a mining operation may have significant cumulative impacts.

Surface water contamination has wide-ranging negative impacts on aquatic biota within the contaminated water body.65,66 A 2004 Canadian Environmental Protection Act (CEPA) Priority Substances List (PSL) toxicology assessment concluded that the effluents released from historical uranium mining and milling operations in Ontario, and both historic and current operations in northern Saskatchewan, particularly those containing uranium and uranium compounds, were toxic to benthic invertebrates, mink, muskrat, plankton, and fish.⁶⁷ The same study concluded that radionuclides from uranium mining and milling were being released into the environment in quantities or conditions that have either had or may have an immediate or long-term harmful effect on the environment and its biological diversity.68 Studies are ongoing to further understand the impacts of these contaminants, including the potential effects of milling effluent releases (specifically metals) on the health of native fish populations, and the impacts of mine effluents on aquatic invertebrates.69

The CNSC has reported for the Rabbit Lake Operation releases of effluent to the Horseshoe Creek drainage system, which in turn drains into Horseshoe Pond and Hidden Bay of Wollaston Lake. While the effluent quality has consistently been within Saskatchewan's limits, a substantial accumulation of uranium and

other metals (molybdenum, arsenic, nickel) in the sediment of Horseshoe Creek and Horseshoe Pond has occurred. Hidden Bay has also shown elevated levels of uranium and molybdenum in the sediment.⁷⁰

The result of the accumulation is such that wildlife and aquatic systems are now at risk due to chemical toxicity of uranium. While fish population numbers are stable in Horseshoe

Table 2.10:

Mass Loadings for McClean Lake Mine (2005)62

Contaminant	2005
TSS (kg)	7,715.6
Arsenic (kg)	37.2
Copper (kg)	6.1
Nickel (kg)	97.4
Lead (kg)	11.9
Zinc (kg)	5.1
Uranium (kg)	117.2
Ra-266 Bq	67,029,573

Pond, benthic invertebrates, which are simpler in structure than fish, that are specifically intolerant of radioactive contamination have been notably absent. Fish sampling in 1994 showed elevated levels of uranium in the flesh of fish. There is widespread concern (including from the CNSC) that the biological and ecological effects will expand downstream throughout the drainage system.⁷¹

The unexpected environmental contamination of Horseshoe Pond and Hidden Bay was reported to be caused by errors in modeling around the natural biological removal processes. It has come to light that the pre-development environmental assessment work (completed in the early 1990s) focused on radiological effects on humans and did not focus on radiological or chemical toxicity effects to non-human biota.⁷²

Cameco and the Natural Sciences and Engineering Research Council of Canada (NSERC) funded a 2001 study of the toxicity of uranium mine receiving

Table 2.11:

Releases of Ammonia from Uranium Mines and Mills to Surface Water⁶⁴

	Rabbit Lake Operation (tonnes)	Key Lake Operation (tonnes)	McClean Lake Operation (tonnes)
2003	12.6	45.5	7.2
2002	9.3	52.7	10.6
2001	8.1	56.1	35.6
water to fathead minnows near the Key Lake mine. The study found that caged minnows in lakes receiving mill effluent had higher mortalities than did minnows in lakes receiving dewatering effluent. The mortality rate amoung the latter group was not significantly higher at reference sites. The cause of these higher mortalities was suggested to be elevated levels of selenium in the mill effluents.⁷³

A 2002 study evaluated the impacts of metals on large-bodied fish and sediments in waters receiving mill effluents and mine dewatering discharges at the Key Lake mine. By analyzing sediments, the study found that in lakes receiving treated mill effluent, the highest degree of contamination was for arsenic, molybdenum and selenium. A study of the bioavailabilility of metals to fish revealed that one particular fish species-white suckers-had six to 38 times higher hepatic (liver) and renal (kidney) concentrations of molybdenum and selenium compared to those from reference (unimpacted) lakes. Similarly, hepatic arsenic was at least 15 times higher and renal nickel was over five times higher in the white suckers from the contaminated lakes than in those from the reference lakes. In a lake receiving mine dewatering discharges, the highest degree of contamination was for cobalt and nickel. In assessing the bioavailabilility of metals to fish, hepatic and renal concentrations of molybdenum, nickel, cobalt and cadmium in were highest in Northern Pike (1.5 to 43 times higher than reference).⁷⁴

2.2.3.3. Disruption of surface water and groundwater flows

The construction and operation of uranium mines, particularly open pit mines, may disrupt existing surface water and groundwater features and flows. The construction of TMFs and waste rock storage facilities may have similar impacts. Mining operations below the water table may require significant ongoing removal of groundwater.

The reported flow rates at mine treatment plants, as shown in **Table 2.12**, provides a partial estimate of the extent of dewatering activities taking place. The total is equivalent to the total water consumption of Toronto, Canada's largest city, over a two-week period, or enough to fill the Rogers Centre (formerly the Sky-Dome) roughly 14 times.⁷⁵

In addition, when a mine is decommissioned, the TMF is moved to a state of passive long-term containment. In this state, a zone of high hydraulic conductivity materials is constructed around the tailings to channel groundwater flow around, rather than through, them.⁷⁹

Table 2.12: Annual Flow Rates of Uranium Mine and Mill Water Treatment Plants

Facility	Annual Flow (million m3)
Rabbit Lake	2.4–3.876
Key Lake	8.1, including 4.7 clean water from dewatering and 2.9 contaminated water from wells ⁷⁷
McArthur River	5.7 78
Total	16.2–17.6

Case Study: Surface Water Contamination at Cluff Lake Mine

The Cluff Lake mine had closed by the time the 2003 CEPA Priority Substances List Assessment Report was released. The Cluff Lake mine is now being decommissioned, so while uranium emissions have decreased to acceptable levels, the water body that received the mine's treated effluents (Island Lake) contains concentrations of key contaminants that may pose a risk to non-human biota.¹ At the Cluff Lake mine, the release of treated effluent and associated contaminant loadings has resulted in adverse effects on water and sediment quality, and on the aquatic ecology of the water body downgradient of the TMA. Changes in the zooplankton, benthic macroinvertebrates and fish community composition, along with bioaccumulation of trace elements in certain aquatic species, have been observed.²

2.3.4 Landscape Impacts

Land footprint

A land footprint is the measure of the total land surface area used by a given facility. The land footprint of uranium mines and milling operations can be very large. The land footprint of uranium mine development includes the tailings and waste rock storage areas, the

mine and mill, and all associated infrastructure.

Table 2.13 displays the areas covered under the surface lease agreements for northern Saskatchewan uranium operations.

The surface lease of the mine represents the largest theoretical area that can be developed by a company. The total in **Table 2.13** represents an area equivalent to almost 26,000 football fields. The lease areas do not include the area taken up by access roads and power corridors.

Although a mine may not physically disturb the entire surface lease agreement area, given the potential impacts to the surrounding ecosystems, the actual area of disturbance from a uranium mine can be considered far greater than its surface lease agreement area. These impacts include contamination of the surrounding environment and biota (Sections 2.3.2. and 2.3.3.) and disruptions of surface water and groundwater flows (Section 2.3.3.3) that may extend throughout the affected watersheds. In addition, large areas may be affected by exploration activity before a mine is developed, including line cutting, trenching and the establishment of temporary roads and camps.

2.3.5 Occupational and Community Health Impacts

2.3.5.1. Occupational Health and Safety

Historically, radiation exposure levels of uranium mine workers often exceeded the occupational limits set by the mine itself.⁸¹ Dust is the main source of radiation exposure in open pit mines and in milling facilities. Radon gas is the primary radiation exposure in underground mines.⁸²

While today uranium mining and milling operations follow mandated radiation exposure limits much more closely, a number of studies link relatively low level radon and radiation exposure to cancers.^{83,84} The annual effective radiation doses received at Canadian uranium mine and mill sites are summarized in **Table 2.14**.

A 2005 report from the United States National Academy of Sciences presented the findings that even low levels of ionizing radiation are harmful.⁸⁹

Table 2.13: Surface Lease A	greement Areas for Northe	ern Saskatchewan	Uranium Operations ⁸
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Mine	Surface Lease Agreement Areas (ha)
McArthur River	651
Key Lake	3,476
Cigar Lake	984
Rabbit Lake	2,001
McClean Lake	3,677
Cluff Lake	1,631
Midwest Joint Venture	646
Parks Lake (essentially the first Rabbit Lake tailings area, post decommissioning monitoring)	800
Beaverlodge (water quality monitoring stations, post decommissioning monitoring)	Not available
TOTAL	13,866

The report also commented on findings from a combined analysis of data from nuclear workers designed to increase the sensitivity analysis of exposure studies and provide direct estimates of the effects of longterm low dose radiation. The analyses were found to be "compatible with a range of possibilities, from a reduction of risk at low doses to risks twice those upon which current radiation protection recommendations are based."⁹⁰

The health impacts of uranium mining have been the subject of over fifty studies globally. Most of these studies find the health impacts to occur over the long term and report levels of lung cancer between two and five times higher in workers who have been exposed to high levels of radon, or exposed over long periods of time to lower levels.⁹¹

Between 1995 and 2005, 44 fatalities were attributed to the uranium mining industry in Ontario alone (historical claims) as reported by the Ontario Workplace Safety and Insurance Board. All 44 fatalities were attributed to occupational disease, one-third of which were attributed to the effects of silica exposure.⁹² Silica is a compound commonly found in uranium ore, which enters the lungs via dust during the mining process.⁹³ Several of the claims were for exposure to solvents, several were for exposure to asbestos, and the remainder were for radiation exposure.⁹⁴

In addition to radiation exposure risks, mine workers are at risk of injury from dusty or noisy workplaces, rockfalls, exposure to chemicals, heavy lifting or repetitive tasks.⁹⁵ Over the 2000–2005 period, the uranium mining industry in Saskatchewan suffered two occupational fatalities, 403 claims for injuries without time losses and 172 claims with time losses.⁹⁶

2.3.5.2. Community Health Impacts

Releases of radiation and other contaminants through air and water also have an impact on the surrounding community. The main exposure pathways for radioactivity from tailings are direct gamma radiation, inhalation of radioactive particulates, and ingestion of radionuclides⁹⁷ through the food chain. While radiation has been shown to accumulate in the biota near uranium mines, the impacts of exposure to the health of the surrounding community are highly contested.

Saskatchewan Health does not study the health impacts of uranium mines on communities near the uranium mines due to confounding factors such as radon in homes and cigarette smoking. However, studies have been performed to assess the health of foodstuffs near uranium mines in northern Saskatchewan by toxicology researchers at the University of Saskatchewan.⁹⁸

In one study in this area, tissues from moose and cattle to be consumed as food were collected. The study concluded that moose and human radiation doses in the Wollaston area were two to three times higher than in control areas. ⁹⁹

Another tissue study from 18 Wollaston caribou concluded that an adult eating 100 g/day of caribou meat would receive annual effective doses of 0.85 mSv/year. Additional eating one liver and ten kidneys per year would double this dose to 1.7 mSv/year. A one-year-old child who consumed only 10per cent of the adult caribou intake would receive more than half the adult dose of radiation. These doses are predominantly from the presence of polonium-210 in the soft tissue. The lichen-caribou-human food chain is considered the most critical food chain in the world for concentrating airborne radionuclides.¹⁰⁰

Table 2.14: Annual Effective Rad	liation Doses at Mine ar	nd Mill Sites ^{85,86,87}
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	2004 (mSv)	2003 (mSv)	2002 (mSv)		
McArthur River	McArthur River				
Average effective dose per employee	1.0	1.6	1.4		
McClean Lake					
Site average effective dose	0.86	1.00	1.03		
Rabbit Lake ⁸⁸					
Average effective dose	2.41	2.12	1.21		

While the study concluded that consumption of moose did not carry with it significant health risks, the consumption of caribou was found to potentially increase the chance of developing cancer to as high as 0.6per cent over a 70-year lifetime, which is equivalent to a rate of six cancers per 1,000 people.¹⁰¹ This far exceeds the US Environmental Protection Agency (EPA) range of acceptable cancer risks of 1 in 10,000 to 1 in 1,000,000.¹⁰²

The issue of whether radiation risk models correctly estimate risks to human health is a subject of considerable debate. Recently, these debates have questioned whether current models are appropriate for assessing the effects of radioactive substances taken into the body. These models are based on significant uncertainties, and it has been suggested that the risks posed by radioactive sources inside the body must be judged carefully.¹⁰³

2.3.6. Mine Closure and Post Closure Care

Once uranium ore reserves are exhausted, the mine, mill and TMF must be decommissioned. Decommissioned mines must be managed in perpetuity to prevent the release of contaminated tailings and waste rock to the surrounding ecosystem and community. Therefore, mine decommissioning is one of the most important steps in the mining and milling process.

Decommissioning consists of removing equipment, materials and buildings, remediating, grading, and revegetating surrounding areas, refilling pits and sealing shafts, and preparing the TMF and waste rock management areas for long-term management.¹⁰⁴ While remediation is the goal of mine decommissioning, it is important to note that sites cannot be returned to their natural state, given the extent of the physical disruption of the landscape that occurs as a result of mining operations, the extent of permanent waste rock and tailings management areas (TMAs), and the presence of long-lived radionuclides and other contaminants at mine sites.

Current practice for the long-term management of TMFs usually involves construction of a zone of high hydraulic conductivity (pervious) material around the tailings, if it does not exist naturally, to channel the groundwater flow around rather than through the tailings.¹⁰⁵ The length and type of monitoring required for the decommissioned project will vary from site to site, and is the responsibility of the mine owner/operator. An Environmental Assessment of the project determines the effects and what mitigation needs to occur.¹⁰⁶

The management of all decommissioned mines must be considered in perpetuity. The Auditor General of Canada has observed that,

In Canada, a "walk-away" solution is not realistic for decommissioning most uranium tailings sites. Long-term storage requires longterm institutional care to monitor and maintain the containment structures and to control access to, and use of, the land.¹⁰⁷

Historically, uranium mines were abandoned without any decommissioning. Prior to the 1980s, uranium mines in northern Saskatchewan were abandoned with very little closure activities and practically no provisions for public safety, environmental protection, or aesthetics. In the 25 to 50 years since closure, many of the sites have deteriorated through natural degradation and vandalism to the point that they now pose "significant public safety hazards and possible long-term environmental concerns."¹⁰⁸ In 2001, the Saskatchewan Government identified 67 mines that had not been properly decommissioned and were in need of further investigation.¹⁰⁹

Of the sites assessed, the Gunnar mine near Uranium City poses the highest risk to environmental and public safety. Since closure of the Gunnar mine in 1964, 4.4 million tonnes of unconfined tailings containing radioactive waste have made their way into Lake Athabasca. The Lorado Mill site was also ranked among the top ten highest risk sites by the same study, as the tailings at this site cover almost 14 hectares and are leaching into two adjacent lakes.¹¹⁰

Federal and provincial governments have had to assume residual responsibility for abandoned uranium mine sites where the original producer of the waste or the current owner could not be held responsible for the waste, or was unwilling or unable to pay to manage it.¹¹¹

In 1994, Natural Resources Canada estimated that the total cost for decommissioning uranium tailings sites in Canada may exceed \$400 million (in 1994 dollars) and that the federal government's potential liabilities to meet its residual responsibilities could be in the "tens of millions of dollars."¹¹²

In 2005 the federal and Saskatchewan governments came to a \$24 million cost-sharing agreement for the clean-up of abandoned uranium mine sites (primarily the former Gunnar and Lorado mines) near Uranium City in northern Saskatchewan.¹¹³

Current requirements for planning for decommissioning of a mining/milling operation can be

Table 2.15: Typical Deco	nmissioning Work Packages ¹¹⁶
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Planning Envelopes	Work Packages		
Mine workings	 Remove salvageable equipment and hazardous materials Stabilize/fill underground workings/open pits Seal shafts, raises, declines and portals Remove ancillary structures and service/remediate contaminated soils Grade and revegetate immediate area 		
Mill site	 Remove coarse ore in storage Remove process chemicals and hazardous materials in storage Remove contaminated equipment and vessels for disposal Remove salvageable equipment and materials, decontaminate as needed Demolish remaining structures and tanks Remediate contaminated soils Grade and revegetate immediate area 		
Tailings mgmt area	 Construct/upgrade containment structures for long term Construct/improve water drainage or diversion works Recontour tailings Place final cover (soil, rock, water, etc.) Install/upgrade monitoring/treatment facilities Remove pipelines, pumps and other ancillary structures Grade and revegetate immediate area 		
Waste rock mgmt area	 Stabilize with respect to infiltration/acid generation Recontour/grade and vegetate or relocate for disposal as required 		
Hazardous material storage area	 Remove materials inventory Remove contaminated tanks and structures for disposal Demolish remaining structures and tanks Remediate contaminated soils Grade and revegetate immediate area 		
Effluent treatment	 Remove remaining effluents and chemicals in storage Remove unnecessary treatment plant, piping and other structures Remediate mine water, sewage and other effluent treatment ponds and sludges Grade and revegetate immediate area 		
Ancillary buildings and services	 Dismantle boiler and powerhouse Remove power lines and substations Remove potable water supply system Remove sewage treatment system Remove non-hazardous material and equipment warehousing Remove camp accommodations Remove mechanical shops Remove administration and security structures Regrade access roads and private airstrips and remove culverts Grade and revegetate immediate area 		

characterized by a two-step process: the preliminary decommissioning plan and the detailed commissioning plan. The preliminary plan is created and filed with the CNSC as early as possible in the life cycle of the facility (before beginning mining activities) and updated to suit changes in the operating conditions. The detailed commissioning plan is submitted to CNSC for approval when the decommissioning decision has been made but before the decommissioning process begins. Detailed plans incorporate procedural and organizational details to the preliminary plan. Upon approval by the CNSC, the detailed plan is changed to a license thereby authorizing decommissioning activity.¹¹⁴ Approvals are also required from the relevant provincial authorities.¹¹⁵

Decommissioning requires extensive work, as is illustrated in **Table 2.15**.

In the context of the past history of the abandonment of mines and the acceptance of liability by federal and provincial governments, financial assurances are now required to ensure funding will be available to carry out the decommissioning and reclamation plans in the event a mining company is unable or unwilling to carry out the required work.¹¹⁷

Table 2.16 shows the mine decommissioning financial assurances currently held against operating uranium mines in Canada.

Given the scale of uranium mining operations, the extent of their impacts, and the costs associated with the remediation of abandoned facilities in the past, the adequacy of these assurances is open to serious question. It has also been noted that the site closure plans and TMF designs make no provision for the effects of climate change (in a zone of discontinuous permafrost) or tectonic activity and that the designs of the TMFs are based on models that have never been tested on an operational scale, but are expected to function in perpetuity.¹¹⁹

2.4. Conclusions

The environmental impacts of uranium mining and milling activities in Canada are severe. These impacts represent the most significant immediate environmental effects of reliance on nuclear energy for electricity production in Canada. A number of jurisdictions in Canada and Australia have adopted bans on the establishment of new uranium mines due to concerns over the potential environmental and health impacts of such operations.

The key impacts include the generation of large quantities of waste rock and mill tailings. These are typically acidic or potentially acid generating, contain long-lived radionuclides, heavy metals, and other contaminants. To meet Canadian domestic power generation requirements, uranium milling operations are estimated to result in the production more than 90,000 tonnes of tailings, and up to 2.9 million tonnes of waste rock per year.

In addition, it was estimated that as of 2003 there were more than 109 million tonnes of waste rock, and 214 million tonnes of tailings in inventory at Canadian uranium mines and mills. Major concerns exist regarding the long-term integrity of tailings and waste rock containment facilities. These facilities will require perpetual care, and the adequacy of the financial assurances required by governments for their closure and long-term management has been questioned.

Severe contamination of groundwater with radionuclides (e.g. lead-210, polonium-210, radium-226), heavy metals (arsenic, manganese, nickel), and other contaminants has occurred at TMFs and waste rock storage areas. Uranium mining and milling facility surface water discharges have resulted in the contamination of the receiving environment with radionuclides and heavy metals (molybdenum, arsenic, nickel, selenium, cobalt, and cadmium). Discharges to surface waters in 2003 included over 1500 kg of uranium,

Mine Stage of Development		Value of Assurance (\$CAN)
Cigar Lake	Early	\$4.21 million
McArthur River	Early	\$0.55 million
Rabbit Lake	Mature	\$25.90 million
Cluff Lake	Mature	\$20.00 million

Table 2.16: Examples of Financial Assurances for Mine Decommissioning Activities¹¹⁸

860 kg of molybdenum, 70 kg of arsenic, 185 kg of nickel, and 40 kg of selenium. Effluent from historic and operating uranium mines and mills, particularly uranium discharges, have been determined to be toxic for the purposes of the *Canadian Environmental Protection Act* by Environment Canada and Health Canada.

Extensive dewatering operations (in excess of 16 billion litres per year) are associated with uranium mining in northern Saskatchewan.

Contamination of the surrounding environment and biota with radionuclides (uranium, radium-226, and lead-210) has occurred, particularly via windblown dust from tailings facilities and mill dust. A significant potential increase in cancer risk from consumption of caribou in the vicinity of uranium mines has been identified as a result.

Major releases of SO_2 and VOCs are associated with the milling process. VOC emissions from the Key Lake mill in 2004 were equivalent to the average annual emissions of more than 300,000 cars. Annual releases of over 43,000 tonnes of SO_2 are reported from the Rabbit Lake acid plant.

Atmospheric releases of NOx and PM also occur from milling processes, and the operation of fossil fuel-powered machinery and equipment.

Annual GHG emissions resulting from uranium mining, milling and tailings management activities in Canada are estimated at between 160,000 and 250,000 tonnes of CO_2 . The use of lower grade ores would result in the generation of proportionally larger amounts of tailings, other wastes and emissions, as larger amounts of ore would have to be processed to produce the same amount of uranium concentrate.

Workers at uranium mines and mills typically receive annual effective radiation doses higher than those considered acceptable to members of the general public. Increased incidences of lung cancer as well as deaths resulting from silica exposure are reported among uranium miners.

3. Phase II: Fuel Processing and Production

Summary of Findings

- The available information on the impacts of uranium refining, conversion and fuel fabrication is limited.
- The major impacts identified include the following:
 - The generation of low and intermediate levels of radioactive wastes associated with fuel production for domestic use in Canada is estimated in the range of 990 tonnes of solid wastes and 9,000 m³ of liquid wastes per year
 - The occupational exposure of workers to radiation in refining conversion and particularly fuel fabrication stages, although within the range acceptable for nuclear energy workers, exceeds levels that would be considered safe for members of the public.
 - Considerable debates exist about the extent of the community health impacts of facilities, particularly in relation to the Port Hope conversion facility, where there is a history of significant contamination of the surrounding community with radionuclides, heavy metals and other contaminants associated with the plant.
 - Releases to the air and water of uranium occur from refining and conversion facilities. Releases of other radionuclides may also occur. Total annual releases range between 17 and 60 kg of uranium to the air, and 6.8 kg to surface waters.
 - Environment Canada concluded in 2004 that releases of uranium and ionizing radiation from uranium refining and conversion facilities were not toxic to the environment for the purposes of the Canadian Environmental Protection Act. The assessment did not consider human health impacts.
 - Releases to the air of lead (approx 10 kg per year) are reported from the Port Hope conversion facility. Releases to the air of persistent organic pollutants, hydrogen fluoride, and criteria air pollutants are reported from refining and conversion facilities as are discharges of ammonia, nitrates and phosphorous to surface waters.
 - Atmospheric releases of NOx, PM and GHGs are associated with transportation of uranium from mill sites in northern Saskatchewan to the Blind River refinery, and then on to the Port Hope conversion facility.
 - Total GHG emissions associated with uranium refining, conversion and fuel fabrication activities in Canada are estimated at between 80,000 and 116,000 tonnes per year.

- The production of enriched uranium fuel, for use in light water, rather than CANDU, reactors is associated with much higher greenhouse gas emissions, particularly where gaseous diffusion processes are used to enrich the uranium. Uranium enrichment processes are also associated with higher atmospheric releases of uranium, and generation of depleted uranium (DU) waste streams.
- No refining, conversion or fuel fabrication facilities have been decommissioned to date. Decommissioning would likely result in the generation of large amounts of hazardous and low- and intermediate-level radioactive wastes. The adequacy of the financial assurances provided by facility operators for decommissioning, waste management and long-term care costs has been questioned.

3.1. Introduction

Once uranium is mined and milled (see Chapter 2), it must be further processed to prepare it for fuel use in nuclear reactors. In Canada, uranium fuel processing for use in CANDU reactors has three key steps, as outlined in **Figure 3.1**, with these processes occurring at the noted locations.

3.2. Fuel Processing and Production Overview

Yellowcake (dried concentrated uranium) is trucked 2,000 km from the mills in Saskatchewan to Canada's only uranium refinery located in Blind River, Ontario on the north shore of Lake Huron. At the refinery, the impurities are removed to produce high-purity uranium trioxide (UO_3) powder via a multi-step chemical and physical process utilizing a solvent extraction purification circuit. The majority of the uranium trioxide is shipped to the Port Hope facility.

In the next step of the fuel production process, the refined UO_3 is trucked 600 km to the Port Hope conversion facility. Some is also sold to the United States facilities for use in nuclear fuel blending.¹ The conversion facility primarily converts UO_3 powder produced at the Blind River refinery to uranium dioxide (UO_2) and uranium hexafluoride (UF_6). There is also a specialty metals plant on site that has been used to convert uranium tetrafluoride into uranium metal shapes for shielding and counterweights for certain types of aircraft. Finally, the facility also includes recycling and decontamination capabilities along with a stand-by plant for UO_2 production.²

 UO_2 is next trucked to fabrication facilities in Port Hope, Toronto, or Peterborough where it is made into fuel pellets for CANDU reactors. There, powdered UO_2 is pressed into small cylindrical shapes and baked at high temperatures (1600-1700 °C) to harden. UF₆ is exported for further processing into fuel for Light Water Reactors.

CANDU reactors require fuel bundles about the

size of a typical fireplace log. These fuel bundles consist of either 28 or 37 half-metre long rods of tubular zirconium alloy sheaths containing ceramic UO_2 pellets. Each fuel bundle weighs about 24 kilograms, including approximately 19 kilograms of uranium. Fuel pellets are processed into bundles at the Peterborough facility and at the Port Hope fabrication plant.³

General Electric (GE) Canada supplies fuel bundles to OPG for its reactors at the Pickering and Darlington plants, and Zircatec supplies fuel bundles to Bruce Power for its reactors at the Bruce plant (owned by OPG, operated by Bruce Power Ltd.).⁴

The Government of Saskatchewan has recently sought the construction of a uranium refinery and conversion by the French state-owned Areva Group.⁵

3.3. Refining, Conversion and Fabrication Facility Impacts

The key impacts of refining, conversion and fabrication processes are summarized in **Table 3.1**. The major impacts include the generation of low- and mediumlevel radioactive wastes, air emissions and the risks of radiation exposure to workers and host communities. The details of these impacts are discussed in the following sections.

Specific figures on production outputs were requested from regulatory agencies and refining, conversion and fabrication facilities. However, the Pembina Institute was informed that production figures are considered proprietary information and are therefore not available in the public domain. As a result, the annual total releases of contaminants from refining, conversion and fabrication facilities are reported. However it is known that approximately 16 per cent of total annual Canadian uranium production is used for domestic reactor fuel.⁶ Therefore this portion of emissions can conservatively be attributed to domestic energy production.

In addition, fuel fabrication facilities do not report on the NPRI. These include GE Canada's Toronto and



Figure 3.1: Fuel Conversion Process

Peterborough nuclear fuel fabrication facilities and Cameco's (Zircatec) Port Hope and Cobourg nuclear fuel fabrication facilities.⁷ The reasons for non-reporting are unknown.

The non-coverage of radionuclides by the NPRI also limits the inventory's usefulness in understanding the extent of releases and transfers of these substances from refining, conversion and fabrication facilities.

3.3.1. Waste Generation

No specific information on waste generation during the refining, conversion and refining of uranium for fuel purposes in Canada is publicly available. Low-level radioactive wastes known to be generated in these processes in Canada include scrap lumber, pallets, rags, paper, cardboard, rubber and plastic. These wastes are typically incinerated. Other radioactive wastes generated include contaminated air filters, fibreglass, polyvinyl chloride (PVC) ductwork, floor sweepings, sandblast sand, insulation, sample bottles, scrap metal, anodes, recyclable scrap metal, radioactive drain wastes and contaminated equipment.⁸

The International Atomic Energy Agency gives the following estimates of the amounts of wastes generated during the processing and production of uranium fuel, including processes similar to those used in the preparation of fuel for CANDU reactors.

With 2003 domestic reactor requirements of 1,650 tonnes of uranium,¹⁰ these estimates would suggest

990 tonnes of solid wastes, and 9,000 m^3 of liquid wastes from fuel production for domestic use.

The only off-site disposal of material reported to the NPRI by refining, conversion and fabrication facilities is the transfer in 2004 of 76 kg of lead to the Port Hope conversion facility, and in 2005 of 2,178 kg of lead to physical treatment off-site.¹¹

It has been reported that, in the production of UO₂, an ammonium nitrate by-product solution is produced. The solution is treated to reduce uranium and radium to levels less than 10 mgU/L and 370 mBq/L respectively. The batches of the solution are analysed for uranium, radium and selected heavy metals to ensure compliance with the Atomic Energy Control Board (AECB, now CNSC) limits, and then released to a local agricultural supply company for use as a fertilizer. It was reported in 1999 that approximately two million litres of solution are transferred every year for this purpose.¹² It has been estimated that this quantity of solution would contain up to 20 kg of uranium.¹³

It is further estimated that 3.5 million cubic metres of historic wastes associated with the Port Hope conversion facility, containing waste uranium, radium and their radioactive decay products, and various heavy metals, still remain within town. It has been suggested that another 872,000 cubic metres lie in the area immediately west of Port Hope at Welcome and Port Granby, awaiting a disposal solution. Approximately 200,000 tonnes of radioactive materials were trans-

Waste Generation	Atmospheric Releases	Water Impacts	Landscape Impacts	Occupational and Community Health
 Low-level radio- active waste—all stages 	 Release of uranium, heavy metals dioxins and furans, hexachlo- robenzene, ammonia, PM, sulphuric acid, NOx, hydrogen chloride and hydrogen fluoride GHGs and other air pollutants released as a result of transportation from mill to pro- cessing facilities and all stages of fuel processing 	• Release of uranium, ammonia, nitrates, phosphorus	 Facility land impacts. Off-site ecosystem impacts of pollutant releases 	 Radiation exposure for workers Historical radioactive contamination in Port Hope area

Table 3.1: Key Impacts of Refining, Conversion and Fabrication Processes

ported from Port Hope to Chalk River during major remedial actions carried out between 1977 and 1981.¹⁴

Port Hope is also designated as an Area of Concern under the Great Lakes Water Quality Agreement due to contaminated sediments in the harbour. Approximately 90,000 m3 of sediments are contaminated with uranium and thorium-series radionuclides, heavy metals, and polychlorinated biphenyls (PCBs). The contaminated sediment is the result of discharges from the Port Hope uranium processing facility (of the crown corporation Eldorado Nuclear Limited) in the 1930s and 1940s.¹⁵

It is important to note that the enrichment of uranium for use in light water reactors, rather than CANDU type reactors, results in the generation of additional waste streams. For each tonne of enriched uranium produced, for example, 7 tonnes of depleted uranium (DU) are generated. The DU is also referred to as "tails" (not to be confused with the mill tailings). DU still contains 0.2 - 0.35% of uranium-235. The fate of the DU generated as a result of fuel enrichment is largely unclear. Most of it is stored as UF₆ in steel containers in open yards near the enrichment plants, although it is known to be used in the production of armour piercing ammunition for military purposes. The radioactivity of depleted uranium increases in the long term, so it becomes highest after approx. 500,000 years.¹⁶

3.3.2. Atmospheric Releases

Releases to the atmosphere from refining, conversion and fabrication operations include uranium, persistent organic pollutants (hexachlorobenzene, dioxins and furans), hydrogen chloride, hydrogen fluoride, and sulphuric acid, as well as GHGs, NOx and PM.

In addition to the direct releases from refining, conversion and fabrication facilities, consideration also has to be given to the transportation emissions associated with the movement of yellowcake from the uranium mines and mills in northern Saskatchewan to the uranium refinery in Blind River, Ontario—a round trip distance of approximately 4,000 km. Additional emissions are associated with the transportation of refined material the 600 km distance from the Blind River refinery to the Port Hope conversion facility.

3.3.2.1. Radionuclides

The only radiological emissions from the Blind River Refinery are uranium particulates and radionuclides in secular equilibrium in the uranium-238 decay chain.¹⁷ The Blind River facility also operates an incinerator. Air emissions models for the facility, assuming use 60 times a year, for six hours each burn, releasing 0.033 mg U/s18, suggest total releases of 42.8 kg of uranium per year.

Table 3.2: Waste Generation from Uranium Processing (per 1,000 tonnes U processed)9

Arisings	Quantity	Classification
Drums	70 t	Material for recycling or waste
Insolubles and filter aids	50 t	Waste
Sludges	300 t	Waste
Liquid nitrate	200 t	By product
Ammonium nitrate solution	5,000 m ³	By product
Extraction residues	10 m ³	Material for treatment
Sludges	1 m ³	Material for treatment
Zircaloy	1 t	Material for treatment
Miscellaneous metal scrap	40 t	Material for treatment
Ventilation filters	100-200 m ³	Material for treatment
Mixed combustible materials	300 m ³	Material for treatment

	Refining (kg U) (Blind River)	Conversion (kg U) (Port Hope)	Fabrication (kg U) (Toronto, Peterborough, Port Hope, and Coburg)
2003	1.8	16	0.05
2002	15.8	16	0.05
2001	22.7	n/a	0.09
2000	34.1	n/a	0.06

Table 3.3: Annual	Uranium Relea	es to Air from	n Uranium Pi	rocessing F	acilities ^{19,20,21,22,23,24}
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Note: 'n/a' indicates that the measurement is not applicable.

Table 3.3 summarizes the annual air releases of uranium from the refinery, conversion facility and fabrication facilities. The annual releases of uranium for the conversion facility are from the UO_2 only. Uranium emissions from the UF_6 plant are over three times higher than those from the UO_2 plant, but are not reported here as UF_6 is not used in the CANDU reactor fuel cycle.

The 2004 Environment Canada and Health Canada PSL assessment report on releases of radionuclides from nuclear facilities concluded that releases of uranium and uranium compounds from Canadian uranium refining and conversion facilities were not entering the environment in quantities or concentrations that may have an immediate or long-term harmful effect on the environment or its biological diversity.²⁵ The report did not assess the human health impacts of uranium releases from these facilities.

The PSL assessment report notes the potential for atmospheric releases of thorium-232, thorium-230, radon-226 and radium-222 from UF₆ and UO₂ conversion plants in addition to uranium and uranium compounds, but no specific information on such releases from the Port Hope plant is publicly available.²⁶

3.3.2.2. Hazardous Air Pollutants

3.3.2.2.1. Persistent Organic Pollutants

Hexachlorobenzene

Hexachlorobenzene is a persistent organochlorine that is reasonably anticipated to be a human carcinogen.²⁷ Hexachlorobenzene is classified as a toxic substance for the purposes of the *Canadian Environmental Protection Act.*²⁸

Table 3.4 displays annual releases of hexachlorobenzene to air from facility stacks during uranium refining and conversion. The facilities are not a major sources of hexachlorobenzene releases relative to other sources in Canada.²⁹

Dioxins and Furans

Dioxins and furans are very toxic, highly persistent chemicals and are classified as toxic substances for the purposes of the *Canadian Environmental Protection Act*.³⁰

Table 3.5 summarizes air releases of dioxins and furans the during the refining and conversion processes reported to the NPRI. The uranium refining and conversion facilities are not major sources of dioxin and furan releases relative to other sources in Canada.³¹

Table 3.4:Annual Releases of Hexachlorobenzene to Air from Uranium Processing32

3.3.2.2.2. Heavy Metals

The Port Hope conversion facility reported air releases of 10.8 kg of lead to the NPRI for the 2004 reporting year. No other heavy metal air releases have been reported to the NPRI by refining, conversion or fabrication facilities.

	Cameco Blind River Refinery (g)	Cameco Conversion (g)
2004	0.044	0.192
2003	0.036	0.197
2002	0.040	0.197

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3.3.2.2.3. Ammonia

Ammonia gas can be absorbed on land by soil, water, and vegetation or it can react in the atmosphere to form ammonium (NH_4^+) compounds such as ammonium sulphate and ammonium nitrate, which fall to the ground in rainwater. While the majority of ammonia will deposit close to the emission source, it can travel hundreds of miles if it reaches the upper levels of the atmosphere.⁶⁰ Gaseous ammonia is classified as toxic for the purposes of the Canadian Environmental Protection Act.34

Table3.6presentsNPRIreportedannualstackreleasesofammonia to airfrom uraniumrefiningandconversionfacilities.

3.3.2.2.4. Hydrogen Chloride and Hydrogen Fluoride

Hydrogen chloride and hydrogen fluoride are emitted from the lab-

oratory and the incinerator at the Blind River refinery. The laboratory analytical tests involving these substances are undertaken a few times each week, while the incinerator runs about 60 times a year.³⁶ **Table 3.7** displays the NPRI reported annual releases of hydrogen fluoride to air from uranium conversion, which are released from the stack. The conversion facility is not a major source of hydrogen flouride releases relative to other industrial facilities in Canada.³⁷

No air releases of hydrogen chloride were reported to the NPRI by uranium refining and conversion and fuel fabrication facilities.

3.3.3.3. Criteria Air Contaminants

Minor releases of criteria air pollutants occur in the refining and conversion processes.

3.3.3.3.1. Nitrogen Oxides (NO_x)

During the refining process, nitrogen dioxide is released to air. **Table 3.8** lists annual releases of NO_2 to the air from the Blind River refinery.

Table	3.5:
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Annual Releases of Dioxin and Furan from Uranium Processing³²

	Cameco Blind River Refinery (g toxicity equivalent (TEQ) ³³)	Cameco Conversion (g TEQ)
2004	0.001	0.006
2003	0.000	0.007
2002	0.000	0.005

Table 3.6:

Annual Releases of Ammonia from Uranium Refining and Conversion³⁵

	Blind River Refinery (tonnes)	Conversion (Port Hope) (tonnes)
2004	0.140	22.37
2003	0.152	21.26
2002	0.089	15.99

 Table 3.7: Annual Releases of Hydrogen Fluoride

 from Uranium Processing³²

	Conversion (Port Hope) (tonnes)
2004	0.414
2003	0.537
2002	1.493

Table 3.8: /	Annual	Releases	of NOx	from	the	Blind
River Refine	ery ³²					

Year	Tonnes
2004	45.64
2003	70.80
2002	79.90

	Refining (Blind River) (tonnes)	Conversion (Port Hope)(tonnes)	Total (tonnes)
2004	0.93	3.96	4.89
2003	1.06	4.83	5.93
2002	1.10	4.12	5.22

Table 3.9: Annual Releases of PM from Uranium Processing Facilities^{38,39}

Table 3.10: Annual Releases of Sulphuric Acid fromUranium Processing Facilities³²

	Conversion (Port Hope)(tonnes)
2004	0.19
2003	0.22
2002	0.21

3.3.3.3.2. Particulate Matter (PM)

PM is a byproduct of both the refining process and the conversion process. **Table 3.9** lists annual quantities of PM released to the air from uranium processing facilities.

3.3.3.3. Sulphuric Acid

Sulphuric acid gas emissions (normally formed in the atmosphere from SO₂ emissions) can dissolve in water droplets and can be carried long distances by prevailing winds. The resulting water droplets are consequently acidic and fall to the ground as rainfall, fog, snow and other forms of precipitation where they are absorbed into soil and damage plants.⁴⁰ The Port Hope conversion facility had reported to the NPRI sulphuric acid releases to the air as summarized in **Table 3.10**.

In 2004, sulphuric acid emissions to air were reported as 0.174 tonnes as stack releases and 0.015

Table 3.11: Transportation related emissions – uranium refining,conversion and fuel fabrication

	GHGs	NOx	PM
	(tonnes CO ₂)48	(tonnes)49	(tonnes)50
Total	8,516	64.86	1.85

tonnes as fugitive releases. In 2003 and 2002, sulphuric acid emissions were reported entirely as fugitive releases.

3.3.2.4. Greenhouse Gases

GHGs are produced at all points in uranium production: refining, conversion and fabrication.

Fossil fuel inputs to the Blind River refinery are primarily natural gas but also include small

amounts of fuel oil, propane and gasoline, primarily to generate steam. In total, 1.33-2.80 units of CO₂ have been estimated to be released for every unit of uranium processed (i.e., 2.8 tonnes CO₂ per tonne of uranium), depending on the fuel source for electricity generation. Organic solvents are also used in the process and contribute to CO₂ emissions at 0.04 units per unit uranium.⁴¹

At the Port Hope conversion facility, electricity, natural gas, fuel oil, propane and gasoline are the major energy sources. In total, 2.80-4.84 units of CO₂ have been estimated to be released for every unit of uranium processed, depending on how the electricity was generated.⁴²

Based on data from fabrication facilities in Ontario, the production of 1,775 tonnes of uranium required 500,000 m³ of natural gas and 14,500 terawatt hours (TWh) of electricity.⁴³ This resulted in the production of 2.65 units of CO_2 per unit of uranium fabricated.⁴⁴

Cumulatively, uranium refining, processing and fuel fabrication releases 6.8–10.3 units of CO₂ per unit of uranium processed on a mass basis. On the basis of total primary production in 2003 of 10,455 tonnes of uranium produced,⁴⁵ and presumably processed, total releases would be in the range of 71,094–107,687 tonnes CO₂.

It is important to note that the production of enriched uranium fuel, for use in light water, rather than CANDU, reactors is associated with much higher greenhouse gas emissions particularly where gaseous diffusion processes are used to enrich the uranium.⁴⁶

3.3.2.6. Transportationrelated Releases

Given the long distance between the uranium mines and mills in northern Saskatchewan and the uranium refinery in Blind River, Ontario, air emissions resulting from trucking the unrefined

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uranium 4,000 km are also included in this section. Transport emissions are based on 2003 uranium production figures of 10.5 kilotonnes (kt), which requires approximately 1,035 trips per year to the uranium refinery.⁴⁷

Emissions associated with this transportation activity are summarized in **Table 3.11**.

3.3.3. Releases to Water

Uranium processing results in the release of a liquid effluent. Depending on the specifics of the process, the International Atomic Energy Agency (IAEA) has estimated that 3,000 to 10,000 m³ of effluent is released per 1,000 tonnes of uranium processed.⁵¹ The primary releases to water from fuel production are ammonia, nitrate ion, and uranium. Specific information on releases to water from uranium refining and processing in Canada is very limited in the pubic domain. This makes it challenging to assess the impacts to water resulting from this stage of the life cycle of nuclear power.

There are no engineered (i.e., not accidental) releases to groundwater from the Blind River refinery. The facility releases process effluent through a diffuser (which dilutes the effluent 100-fold) into Lake Huron. At the Port Hope conversion facility, only cooling water is directly discharged to the lake. Process water is treated and re-processed.⁵² Fabrication facilities discharge their liquid effluents to municipal sanitary sewer systems,⁵³ where persistent pollutants such as heavy metals will either be discharged to surface waters or accumulated in sewage sludge. The latter may be landfilled or applied to farm fields.

3.3.3.1. Uranium Discharges

Table 3.12 lists annual uranium loadings to sanitarysewers from uranium processing facilities in Canada.

Table 3.12:

Uranium Released to Water From Uranium Processing Facilities 54, 55, 56, 57, 58

Loadings given for the Port Hope conversion facility are due to both production of UO_2 for CANDU fuel, and for UF_6 for other types of nuclear fuel.

The 2004 Environment Canada and Health Canada PSL assessment report on releases of radionuclides from nuclear facilities concluded that releases of uranium and uranium compounds from Canadian uranium refining and conversion facilities were not entering the environment in quantities or concentrations that would have an immediate or long-term harmful effect on the environment or its biological diversity.⁵⁹ The report did not assess the human health impacts of uranium releases from these facilities.

3.3.3.2 Ammonia

Due to its solubility in water, ammonia can damage natural habitats through nitrogen eutrophication and acidification processes. Acidification can lead to increased levels of toxic metals, which can then leach into surface water, affect aquatic life, and accumulate in the food chain.⁶⁰ Ammonia dissolved in water is classified as a toxic substance for the purposes of the *Canadian Environmental Protection Act*.⁶¹

The majority of ammonia releases emitted in the fuel processing stages are emitted via the air as stack releases. Cameco's Port Hope conversion facility also releases ammonia to surface water and reported 'direct discharge' releases of 0.013, 0.036, and 0.255 tonnes in 2004, 2003 and 2002, respectively. The facility is not a major source of ammonia releases relative to other facilities in Canada.⁶²

3.3.3.3. Nitrate lons

Nitrates are nutrients. Excessive loadings of nitrates can play a role in the eutrophication of water bodies.⁶³ High concentrations in drinking water are associated with health impacts.⁶⁴

Table 3.13 reports nitrate ion releases to water, as reported to the NPRI. The facility is not a major source of nitrate releases relative to other facilities in Canada.⁶⁵

No other nitrate ion releases were reported to the NPRI from uranium refining, conversion or fabrication facilities.

Note: 'n/a' indicates that the measurement is not applicable.

	Refining (Blind River) (kg U)	Conversion (Port Hope) (kg U)	Fabrication (kg U)
2004	2.8	N/a	1.07
2003	2.9	N/a	3.88
2002	6.2	3.9	1.94
2001	7	4.5	1.13

	Refinery (Blind River) (tonnes)
2004	6.329
2003	4.755
2002	6.235

 Table 3.13: Nitrate Ion Releases from Cameco's Blind

 River Refinery⁶⁶

Table 3.14: Phosphorus Releases from Cameco'sBlind River Refinery

	Blind River Refinery(tonnes)
2004	0.171
2003	0.157

3.3.3.4. Phosphorus

Aquatic algae and plants use phosphorus for nutrition making it difficult to detect as it is quickly absorbed by plants. In most freshwater bodies, phosphorus is the primary nutrient that limits plant and algae growth. Excessive phosphorus levels can lead to changes in numbers and types of plants, decline of oxygen levels in the water and increased buildup rates of dead organic material.⁶⁷

Table 3.14 outlines phosphorus releases to surface water as direct discharge, as reported to the NPRI for the Blind River Refinery. Again, no other facilities reported phosphorous releases. The facility is not a major source of phosphorous relative to other facilities in Canada.⁶⁸

3.3.4. Landscape Impacts

Cameco Corporation only reports its land use in Canada aggregated for its entire operations, which include mining, milling, refining, processing, nuclear fabrication facilities and other undisclosed operations. Reported numbers include 20 km² disturbed land, 4 km² rehabilitated land, and 83 km² total owned and leased land. Land use per uranium processing facility is generally not available in the public domain.⁶⁹

At the Blind River Refinery, Cameco owns 636 acres of land and has leased an additional 481 acres from the town. The refinery physically occupies 28 acres of land.⁷⁰

3.3.5. Occupational and Community Health

3.3.5.1. Occupational Health

Uranium refining, conversion, and fabrication in Canada employs approximately 900 workers. The unique occupational health and safety risk in the sector is radiation exposure.

3.3.5.1.1. Radiation Exposure

Workers at uranium fabrication facilities receive doses of radiation due to their proximity to and contact with uranium. **Table 3.15** lists annual average skin, extremity and whole body doses at all uranium processing facilities in Ontario. The whole body effective dose limit for nuclear energy workers, as regulated by the CNSC, is 50 mSv in one year and 100 mSv in five years. The annual effective radiation dose limits to skin, hands and feet are each 500 mSv. For comparative purposes, annual effective dose limits to members of the public for skin, hands and feet are each 50 mSv, while the limit for the whole body is 1 mSv.⁷¹

Whole body doses and skin doses are measured by dosimeters worn on the breast pocket, while extremity doses are measured with dosimeter rings or bracelets worn for one week each quarter.

Table 3.15 indicates that, if nuclear energy workers were subject to the same standards as the general public, workers at the GE fabrication facility in Toronto would consistently face higher than acceptable levels of radiation exposure to their extremities. Production at the GE Canada Toronto fabrication facility requires the handling of open source UO₂ material, while production at the Peterborough facility, where pellets are mounted into bundles, requires handling of UO₂ sheathed inside zirconium.⁸⁰

Table 3.15 also indicates that workers at all facilities regularly receive whole body doses above the levels considered acceptable to the general public. Finally **Table 3.15** indicates that Cameco does not measure exposure to extremities at its facilities. Extremities are exposed to higher rates of radiation than are other parts of the body.

3.3.5.1.2. Workplace Safety

From 1995–2005, the uranium processing industry in Ontario reported one fatality, 102 lost time accidents, and 464 medical aid claims with no lost time. The fatality was a result of asbestos exposure.⁸¹

Table 3.15: Average Annual Facility Doses (mSv) to Workers at Canadian Uranium Processing Facilities

Facility		2003 (mSv)			2002 (mSv)			2001 (mSv)			2000 (mSv)	
	Skin	Extremities	Whole Body	Skin	Extremities	Whole Body	Skin	Extremities	Whole Body	Skin	Extremities	Whole Body
Cameco Refinery, Blind River ^{72,73, 74}	3.6	m/n	1.6	2.5	m/n	1.3	2.1	m/n	1.1	2.1	m/n	1.1
Cameco Conversion Facility, Port Hope ⁷⁵	n/a	m/n	n/a	1.3	m/n	0.8	1.4	m/n	0.8	1.4	m/n	0.9
Zircatec Fabrication Facility, Port Hope ^{76,77}	5.91	29.63	1.42	4.36	31.29	1.84	3.18	31.72	1.81	2.40	24.92	1.57
GE Canada Fabrication Facility, Toronto ^{78,79}	25.5	64.1	4.07	23.1	63.1	3.99	25	67.1	3.73	28.8	170.8	3.78
GE Canada Fabrication Facility, Peterborough ^{78,79}	2.75	35.21	1.72	3.5	48.3	2.04	1.5	5.8	0.97	2.1	20.1	1.6
Note: 'n/m' indicates that the factor wa	s not measured	d; 'n/a' indicates that	the measurem	ient is not app	olicable.							

3.3.5.2. Community Health Impacts

In addition to nuclear energy workers, the processing of nuclear fuel also has health impacts on the surrounding community. The focus of this section is on radiation dose limits and linkages to increased cancer rates resulting from discharges of radioactive material. Only limited information on this subject was found in the public domain.

3.3.5.2.1. Radiation Doses

The public radiation dose limit prescribed in Radiation Protection Regulations is 1 mSv/year. Annual doses to the most exposed individuals living near uranium conversion and refinery facilities are estimated to be 0.0025–0.2 mSv. For individuals living near fabrication facilities, the dose range is 0–0.17 mSv.⁸² This dose is in addition to the radiation dose obtained from natural sources. These levels are below the CNSC annual public dose limit of 1,000 microsieverts (μ Sv).⁸³

Respirable uranium particulate in air has been monitored at three locations around the Port Hope fuel facility since 1998. The highest average particulate measured at one of these stations was 0.0074 μ g/m^{3.84} The corresponding radiation dose from inhalation can be calculated at 0.013 mSv/ per year. ⁸⁵

The 2004 Environment Canada and Health Canada PSL assessment report on releases of radionuclides from nuclear facilities concluded that ionizing radiation from Canadian uranium refining and conversion facilities was not entering the environment in quantity or concentration that may have an immediate or long-term harmful effect on the environment or its biological diversity.⁸⁶ However, the assessment did not assess the human health impacts of radiation from these facilities.

In February 2002, the CNSC, after holding a twoday public hearing, issued a five-year license renewal for Cameco Corporation of Saskatoon, Saskatchewan to operate its uranium processing facility in Port Hope.

In a dissenting statement—unprecedented in the history of the commission and of its predecessor, the Atomic Energy Control Board—Commission Member Dr. C.R. Barnes disagreed with the five-year duration of the license term (rather than the three-year terms issued earlier):

Commission Member Dr. C. R. Barnes concurred with the other Members that the licensee meets the requirements of section 24(4) of the Nuclear Safety and Control Act (NSCA) and therefore that a licence should be issued.

Dr. Barnes, however, disagreed with the majority view on the duration of the licence term. In Dr. Barnes' view, a maximum licence term of three years should be approved. Dr. Barnes held that a five-year licence should be reserved for facilities where the effects have been demonstrated to be well characterized and where public concerns about health and safety are not high. Dr. Barnes found that a five-year licence in this case would not adequately address the significant remaining concerns of the public about the health effects of the facility in combination with the past uranium contamination in the community. Furthermore, Dr. Barnes was also concerned about the current lack of environmental effects monitoring in the vicinity of the facility. Dr. Barnes is of the view that bringing the matter of the licence renewal before the Commission in three years time, as opposed to a status report, will have a greater influence on ensuring the licensee maintains close attention to the design and implementation of the environmental effects monitoring program and the need to continue to address the significant remaining concerns of the people potentially affected.87

As noted earlier, Port Hope is designated as an Area of Concern under the Great Lakes Water Quality Agreement due to sediments contaminated with uranium and thorium-series radionuclides, heavy metals, and PCBs in the harbour.⁸⁸ An estimated 3.5 million m³ of similarly contaminated historic waste associated with the conversion facility remains within the town.⁸⁹ In the Mill Street/Madison Street area located southeast of the facility, for example, the Ontario Ministry of the Environment found a uranium concentration of 135 parts per million (ppm) in 1986. The concentration had decreased to 40 ppm in 1997.⁹⁰

The current uranium deposition rate has been monitored at five test plots; three have ben monitored by Cameco and two by the Ontario Ministry of the Environment (MOE). At the Cameco soil test plots, a maximum increase in soil uranium concentration of 0.58 ppm per year was observed for 1998—the first year of monitoring. At MOE's Town Hall plot, the uranium concentration in soil dropped from 0.85 ppm to 0.56 ppm in 1997 and 0.58 ppm in 1998. At MOE's Marina plot, the uranium concentration increased from 0.85 ppm to 1.78 ppm in 1997 and 3.93 ppm in 1998. In Ontario, normal background levels for uranium in soil are up to 2 ppm.⁹¹

Planning Envelopes	Work Packages
Materials shipping, receiving and storage areas	 Remove product/yellowcake inventories. Decontaminate and remove equipment, tools, conveyors, hoists, etc.
Digester process area	 Remove contents and loose contamination from primary and secondary digesters Dismantle digester vessels Remove ancillary piping, valves and electrics Remove other equipment and tools
Solvent extraction process area	 Remove contents of vessels and piping Decontaminate and dismantle feed tanks Decontaminate and dismantle column trains Decontaminate and dismantle settling tanks Dismantle ancillary piping, valves, electrical and conveyance systems
Reactor areas	 Remove contents of denitrification reactors Decontaminate and dismantle reactor vessels Decontaminate and remove reaction gas scrubber system Remove active drains
Effluent management systems	 Remove contents of effluent neutralization vessels Remediate effluent monitoring and treatment lagoons Remediate storm water management lagoon Remove final effluent discharge line Decontaminate sumps Decontaminate and remove raffinate evaporators Decontaminate and remove liquor evaporators
Emission control system	 Remove baghouse filter system Remove central vacuum system
Solid waste management areas	 Decontaminate uranium scrap area Decontaminate and remove refuse incinerator Decontaminate drum cleaning and processing area Remove inventory and decontaminate low-level storage area
Maintenance and trades shops	 Remove tools and equipment Remove other materials and stores Remove work benches, furniture, etc. Dismantle mechanical and electrical rooms
Administrative offices and labs	 Remove equipment, furniture and fixtures Decontaminate laboratories and remove equipment
Chemical tank farm	 Remove inventory Dismantle and dispose of tanks
Building surfaces and structure	 Decontaminate interior floors, walls and ceilings as required Decontaminate exterior surfaces as required Remove heating, ventilation and air conditioning (HVAC) ductwork Remove plumbing, electrical and other services Demolish structures
Site	 Remove waste piles and other potentially contaminated materials Remove contaminated soil and asphalt Grade and revegetate immediate area Complete final release survey

 Table 3.16: Typical Nuclear Fuel Processing Facility Decommissioning Work Packages⁹⁵

In March 2001, an agreement was reached between the federal government, the town of Port Hope and adjacent municipalities on the cleanup of wastes and development of facilities for their long-term management. This \$260 million project,⁸⁸ called the Port Hope Area Initiative (PHAI), is being managed by the Low Level Radioactive Waste Management Office (LLRWMO) of the Ministry of Natural Resources.⁹²

Concerns have been raised by community residents about the health impacts of exposure to historical radioactive and other materials in Port Hope, as well as emissions from the current uranium processing industry. A Health Canada study in 2000 found that Port Hope cancer rates were comparable to that of the rest of Ontario.⁹³ These findings have been challenged by independent researchers and community residents.⁹⁴

3.3.6 Facility Decommissioning

To date, no uranium refining, conversion, or fuel fabrication facilities in Canada have been decommissioned. Decommissioning would be a complex and potentially expensive undertaking, and would generate substantial amounts of low- and medium-level radioactive wastes.

Table 3.16, taken from CNSC Regulatory Guide G-219, outlines typical work packages that would be completed for the decommissioning of nuclear fuel processing facilities.

The CSNC requires that, as part of its license, Cameco make a financial guarantee for future decommissioning of its Port Hope conversion facility. Cameco submitted to the CNSC an irrevocable standby Letter of Credit for \$33.8 million devoted to decommissioning costs.⁹⁶ A financial assurance of \$14.6 million is held against the Blind River refinery.⁹⁷ An assurance of \$3.3 million is held against the Zincatec Port Hope fuel fabrication facility.⁹⁸ The adequacy of these assurances to cover actual decommissioning costs has been questioned.⁹⁹

3.4 Conclusions

The publicly available information on the impacts of uranium refining, conversion and fuel fabrication is limited.

The major impacts identified in this chapter include the generation of low and intermediate levels of radioactive wastes, estimated in the range of 990 tonnes of solid wastes and 9,000 m³ of liquid wastes per year, as a result of fuel production for domestic use.

Releases of uranium from refining and conversion facilities to the atmosphere and to water do occur. Releases of other radionuclides may also occur. Total annual releases range between 17 and 60 kg of uranium to the air, and 6.8 kg to surface waters. Environment Canada concluded in 2004 that releases of uranium and ionizing radiation from uranium refining and conversion facilities were not toxic to the environment for the purposes of the *Canadian Environmental Protection Act*. The assessment did not consider human health impacts.

Releases to the atmosphere of lead (approximately 10 kg per year) are reported from the Port Hope conversion facility. Minor releases to the air of persistent organic pollutants, hydrogen fluoride, and criteria air pollutants are reported from the Blind River refining and Port Hope conversion facilities, as are minor discharges of ammonia, nitrates and phosphorous to surface waters.

Atmospheric releases of nitrogen oxides, PM and GHGs are associated with transportation of uranium from mill sites in northern Saskatchewan to the Blind River refinery, and then on to the Port Hope Conversion facility.

Total air releases from uranium refining, conversion and fuel fabrication operations and associated transportation activities are summarized in **Table 3.17**.

Total releases of NOx from uranium processing facilities, including transport from northern Sas-

 Table 3.17: Total 2003 Air Releases Considering Transportation and Production

Uranium (kg)	Dioxins and Furans (grams TEQ)	Hexachloro- benzene (grams)	Sulphuric Acid (tonnes)	Hydrogen Fluoride (tonnes)	Ammonia (tonnes)	NOx (tonnes)	PM (tonnes)	GHGs (tonnes)
17.85	0.007	0.233	0.22	0.537	21.4	135.66	7.78	79,610 - 116,303

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katchewan, are roughly equivalent to the annual exhaust of 29,600 cars.¹⁰⁰ In 2004 the total releases of PM from uranium processing facilities (including transport from northern Saskatchewan) were roughly equivalent to the annual exhaust of 9,000 cars.¹⁰¹ Total GHG emissions in the same year were equivalent to those produced annually by 22,700 to 33,100 cars.

The production of enriched uranium fuel, for use in light water, rather than CANDU, reactors is associated with much higher greenhouse gas emissions particularly where gaseous diffusion processes are used to enrich the uranium. Uranium enrichment is also associated with much higher atmospheric releases of uranium, and the production of depleted uranium (DU) waste streams.

No refining, conversion or fuel fabrication facilities have been decommissioned to date. Decommissioning would likely result in the generation of large amounts of hazardous and low- and medium-level radioactive wastes. The adequacy of the financial assurances provided by facility operators for decommissioning, waste management and long-term care costs has been questioned.

The occupational exposure of workers to radiation in refining, conversion and, particularly, fuel fabrication stages, although within the range acceptable for nuclear energy workers, exceeds levels considered unacceptable to members of the public.

Major debates exist about the extent of the community health impacts of refining, conversion and fuel fabrication facilities, particularly in relation to the Port Hope conversion facility. There is a history of significant contamination of the surrounding community with radionuclides, heavy metals and other contaminants associated with the facility.

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4. Phase III: Power Plant Operation

Summary of Findings

The major environmental impacts of nuclear power plant operation in Canada include the following:

- The generation of approximately 85,000 waste fuel bundles each year. These fuel bundles are extremely radioactive, contain other toxic materials, and require isolation from the environment for one million years. No long-term management strategy for these wastes is in place. A strategy proposed by the Nuclear Waste Management Organization is currently under consideration by the federal government.
- The generation of between 5,500 and 7,000 m³ of low-level radioactive wastes per year (Ontario only). No national strategy for the management of these wastes is in place or under development. A wide range of radiological and hazardous pollutants has been released to the atmosphere as a result of the incineration of these wastes at the Bruce Western Waste Management facility. A new incinerator installed at the in 2003 has reduced emissions of hazardous, but not radiological, pollutants.
- The generation of waste heavy metals and other hazardous wastes (e.g., asbestos) from facility operation and maintenance and refurbishment activities.
- The generation of large amounts of radioactive and hazardous wastes from reactor decommissioning (expected to occur 30 years after the end of the plant's life). The costs of decommissioning Ontario's existing reactors have been estimated at \$7.474 billion (present value \$6.263 billion). The decommissioning fund maintained by OPG currently has a balance of \$4.211 billion. The province of Ontario provides a financial guarantee for any shortfall between the value of the decommissioning fund and the actual costs of decommissioning.
- Routine and accidental releases of tritium oxide and carbon-14 to surface water and groundwater from nuclear generating facilities. The health and environmental significance of these releases is highly disputed. The current permissible levels of tritium in drinking water in Ontario and are substantially higher than those permitted in the United States or European Union.
- Discharges of hydrazine, a recognized carcinogen, to surface waters. Ontario nuclear facilities are the most significant source of such discharges in Canada.
- Significant discharges of metals (copper, zinc and chromium) have resulted from scouring and corrosion of boilers and heat exchangers in Ontario nuclear facilities.
- The use of large amounts of cooling water. The Darlington and Pickering facilities alone used approximately 8.9 trillion litres in 2003, approximately 19 times the City of

Toronto's annual water consumption Adverse impacts on fish arising from the thermal impacts of large cooling water discharges have been noted.

- Routine and accidental releases of tritium oxide, noble gases, iodine-131, radioactive particulates, elemental tritium and carbon-14 to the atmosphere. Recent studies have suggested that the community health impacts of these releases, particularly with respect to their effects on infants and children, may be more significant than previously thought.
- Atmospheric releases of hydrazine, an extremely hazardous pollutant. Ontario nuclear generating facilities are the most significant source of such releases in Canada.
- Minor releases of criteria air pollutants and GHGs as a result of the testing of fossil fuelpowered emergency generating equipment at nuclear generating facilities.
- Large-scale flaring of hydrogen sulphide (H_2S) (up to 2,000 tonnes per year) in the past due to the production of the heavy water moderator (deuterium oxide) for CANDU reactors.

Releases of radioactive materials to the environment as a result of accidents, and a potential, although low probability, of catastrophic accidents at nuclear generating facilities. Catastrophic accidents at generating facilities would have impacts over a much wider area and longer term, than accidents involving any other electricity generating technology, particularly as a result of the large-scale release of radiation and radionuclides. It has been estimated that the monetized value of the off-site environmental, health and economic impacts of a major accident at the Darlington generating facility, for example, would exceed \$1 trillion (1991 \$Cdn).

4.1. Introduction

After the uranium fuel has been processed, it can be used to produce power in a nuclear plant. There are currently five commercial nuclear power generating stations in Canada, all of which use CANDU reactors: three in Ontario (Pickering – eight reactors, Darlington – four reactors, and Bruce – eight reactors); one in Quebec (Gentilly-2 – one reactor) and one in New Brunswick (Point Lepreau – one reactor).

4.2. Overview of CANDU Nuclear Power Plant Operation

Once the uranium is processed into fuel bundles, the bundles are delivered to nuclear plants and placed in the reactors, where a chain fission reaction (splitting of atoms) takes place. Neutrons slowed by a heavy water moderator collide with uranium-235 atoms and split these atoms into radioactive by-products. This reaction releases radiation, fast neutrons and heat. This heat generates steam to rotate turbines, which in turn drive an electrical generator, thereby generating electricity.

About 220 megawatt hours (MWh) are generated per kilogram of uranium in the reactor.¹ Each fuel bundle stays in the reactor for 12 to 18 months,² after which it is removed from the reactor. Nuclear fuel waste management issues are discussed in detail in Chapter 5.

4.3. Impacts of CANDU facility operation

Nuclear power plant operation has several biophysical and socio-economic impacts. These are summarized in **Table 4.1**. Key issues include the generation of highly radioactive spent fuel wastes, low- and medium-level radioactive wastes, and routine and accidental releases of radionuclides, particularly tritium, to surface water, groundwater, and the atmosphere. The issues of facility performance and costs in Canada are discussed in

Waste Generation	Atmospheric Releases	Water Impacts	Landscape Impacts	Occupational Health and Community Health Impacts
 Waste fuel generation Low-level radioactive waste from plant maintenance, operations, and refurbishments Hazardous wastes from plant maintenance, operations and refurbishments 	 Releases of tritium oxide, noble gases, iodine-131, radio- active particulate, and carbon-14 Releases of ammo- nia and hydrazine Releases of GHGs, carbon monoxide, oxides of nitrogen, SO₂ and PM from standby generator testing Releases of radiological con- taminants, heavy metals, persistent toxic pollutants and criteria air pollutants from on-site incinera- tion of low- and medium-level wastes Historic releases of H₂S from heavy water production 	 Routine and accidental tritium oxide and gross beta-gamma activity releases Large water use for cooling; thermal impacts on receiving ecosystems Large historic water use for heavy water production Releases to water as a result of erosion of boiler tubes (chromium at Darlington; copper and other materials at Pickering and Bruce) Hydrazine and ammonia discharges 	 A 1,000 megawatt equivalent (MWe) nuclear generation station requires up to 4 km² 	Occupational Health • Radiation exposure and hazards Community Health • Tritium releases to atmosphere, surface water and groundwater • Radiation exposure through food, air, water, and so on • Risk exposure to catastrophic events

Figure 4.1: Breakdown of Spent Nuclear Fuel Bundle Accumulation by Source, 2003 (adapted from the Low Level Radioactive Waste Management Office)⁸



detail in Chapter 6. It is important to note that the poor performance of a large portion of the Ontario reactor fleet has resulted in major increases in reliance on electricity sources with high environmental impacts of their own, particularly domestic and imported coal-fired generation.

4.3.1. Waste Generation

The operation of nuclear power plants results in the generation of a number of waste streams, including waste nuclear fuel, low and intermediate level radioactive wastes, and other hazardous wastes, particularly heavy metals and asbestos.

4.3.1.1. Nuclear Fuel Waste

When a spent fuel rod is removed from a CANDU nuclear reactor, it is extremely radioactive; an unprotected person standing within a metre of such a bundle would die within an hour.³ The radioactivity decreases with time. After one year, it decreases to about 1per cent of its initial value; after 100 years, it decreases to about 0.01per cent of its initial value. But it takes one million years for the level of radioactivity of spent fuel to return to that of natural uranium.⁴ Spent fuel also gives off heat and contains toxic chemical elements such as heavy metals.⁵ It must therefore be isolated and contained over extremely long periods of time. Waste nuclear fuel also contains materials (uranium and plutonium) that can be used in nuclear weapons production, and therefore must be secured against access for such uses.

About 85,000 used fuel bundles are generated every year by Canadian nuclear reactors.⁶ As of 2003, there were 1.7 million fuel bundles in storage at nuclear generators and this number continues to grow.⁷ **Figure 4.1** shows the breakdown of spent nuclear fuel bundle accumulation by source. More than 85per cent of the total is in Ontario.

The spent fuel waste is currently stored at the reactor sites. Its continued storage at the reactors is not considered a long-term solution due to the unsuitability of the sites and the fact that the communities hosting the reactors did not agree to, or anticipate, the long-term storage of nuclear waste at these sites.⁹

A full discussion of waste nuclear fuel management impacts and issues is provided in Chapter 5.

4.3.1.2. Low-level and Medium-level Radioactive Wastes

Low- and medium- (or intermediate-) level waste (LILW) is waste that can be dangerous for either short or extended periods of time. It can include waste with an amount of radioactivity that requires shielding containers. LILW is further segregated into "shortlived" (less than 30 years) and "long-lived" (more than 30 years) categories depending on the half-lives of the radionuclides within.¹⁰

Low-level wastes are items with low levels of radioactivity that have been contaminated during clean-up and/or maintenance activities at power generation facilities and can include mops, rags, floor sweepings, clothing, and tools. Low-level waste is considered by the CNSC as safe enough to handle without any radiation protection. Intermediate-level wastes are of a level where shielding is required to protect workers and can include ion exchange resins, filters, and irradiated core components.¹¹

Nuclear power plants produce about 80per cent of the low-level radioactive waste that is accumulated in Canada each year.¹² Wastes generated as a result of plant operations and maintenance include paper, plastic, rubber, cotton, wood, organic liquids, plastic PVC suits, fibreglass, metal pieces, empty drums, filters, light bulb cable, used equipment, construction debris, absorbents (sand, vermiculite, sweeping compound), ion exchange resins, reactor core components and retubing wastes, radioactive drain wastes and chemical cleaning solutions. Low- and medium-level wastes are also generated as a result of reactor refurbishment projects. The waste generated at Ontario's nuclear power plants is collected at the Western Waste Management Facility at the Bruce station site in Kincardine. Some of the wastes are compacted (e.g., paper, plastic PVC suits, rubber, fibreglass, metal pieces and empty drums) and some are incinerated (e.g., paper, plastic, rubber, cotton, wood and organic liquids).¹³ Compacted and nonprocessable wastes are stored onsite at the Bruce facility.

The incinerator operated at the western waste management facility until 2001 was associated with releases to the air of radiological contaminants, particularly tritium, iodine-131 and carbon-14,¹⁴ and a wide range of non-radiological pollutants, including heavy metals (e.g. cadmium, lead and mercury), dioxins and furans, hydrogen chloride, particulate matter, and sulphur and nitrogen oxides. Emission levels of non-radiological pollutants, particularly hydrogen chloride and dioxins and furans were often well in excess of Canadian Council of Ministers of the Environment, Ontario Ministry of the Environment and United States Environmental Protection Agency guidelines.¹⁵

A new incinerator was commissioned in 2003. This resulted in substantial reductions in emissions of non-radiological pollutants, with emissions of dioxins and furans, measured as Toxicity Equivalent (TEQ) being reduced from 177 to 0.017 ng TEQ/Rm³ from 1999 to 2005. Similarly mercury emissions were reduced from 4.7 to <0.015 ug/Rm³ 1998 to 2005.¹⁶ Emissions of some radiological contaminants, particularly particulates and iodine-131 were also reduced, although emissions of tritium and carbon-14 remained unchanged.¹⁷

No overall national strategy is in place for the management of low and intermediate level radioactive wastes in Canada. OPG is advancing proposals for a permanent underground repository for low- and intermediate- level wastes at the Bruce site.¹⁸ The proposal has been subject to calls for a full panel review under the *Canadian Environmental Assessment Act* rather than the CNSC, in light of the following considerations:¹⁹

• The potentially infinite lifetime of the facility

- The diversity of radioactive materials in many different physical and chemical forms that would be deposited in the facility
- The inadequacy of existing categories of radioactive waste materials to adequately describe the

 Table 4.2: Low-level Waste Generated by Ontario Nuclear Power Plants²⁰

Year	Low-level Radioactive Waste Generated (m ³)	Low-level Radioactive Waste Stored – After Processing (m ³)
2003	5,556	3,219
2002	6,983	3,238
2001	6,060	3,334
2000	4,585	2,065

long-term hazards

- The lack of a detailed inventory of radionuclides involved with an emphasis on questions of longevity, mobility and radiotoxicity
- The new and hitherto unexamined categories of radioactive waste that may be emplaced in the proposed deep geologic repository as a result of retubing, refurbishment and decommissioning activities.

Table 4.2 summarizes the volumes of low-level radioactive waste generated at Ontario's nuclear power plants for the years 2000 to 2003.

Based on the total electricity production by Ontario's nuclear power plants between 2000 and 2003, an average of 93 m³ of low-level radioactive waste is created per TWh of electricity generated.

No strategies are currently in place for the longterm management of low and intermediate level wastes produced at nuclear generating facilities in Quebec and New Brunswick.²¹

4.3.1.3. Other Wastes

4.3.1.3.1. Metals

Nuclear generating facilities report large-scale transfers of heavy metals, particularly lead, mercury and cadmium off-site for disposal and recycling. Lead and mercury are classified as toxic substances for the purposes of the *Canadian Environmental Protection Act*.²²

Transfers reported to the NPRI over the past few years are outlined below.

Lead

The amounts of lead being disposed of are substantial. Among 443 reporters of transfers of lead to disposal or recycling in Canada in 2002, Bruce Power, for example, was the thirteenth largest source and Pickering the twenty-third largest source.²⁴

Bruce Power reports that their sources of lead emis-

Total	Bruce (kg)	Pickering (kg)	Darlington (kg)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	13,729.000	110.500	60,000.000	-	-
2003	123,564.000	82,100.000	30,000.000	-	-
2002	160,647.000	41,000.000	-	-	-

Table 4.3: Off-site Releases of Lead from Nuclear Power Plant Facilities (Off-site Disposal and Recycling)²³

Note: '-' indicates that data was not reported.

Table 4.4: Off-site Releases of Cadmium from Nuclear Power Plant Facilities (Off-site Disposal)²⁶

Total	Bruce (kg)	Pickering (kg)	Darlington (kg)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	176.000	-	-	-	-
2003	168.800	-	-	-	-
2002	168.800	-	-	-	-

Note: '-' indicates that data was not reported.

Table 4.5: Off-site Releases of Mercu	y from Nuclear Power Plant Facilities	(Off-site Disposal and	Recycling)28
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Total	Bruce (kg)	Pickering (kg)	Darlington (kg)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	-	156.060	0.354	-	-
2003	-	1.600	0.656	-	-
2002	8.792	1.676	5.110	-	-

Table 4.6: Off-site Releases of Chromium from Nuclear Power Plant Facilities (Off-site Disposal and Recycling)³⁰

Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (tonnes)	Point Lepreau (tonnes)
2004	0.007	-	-	-	-
2003	0.000	-	-	-	-
2002	0.008	-	-	-	-

sions are from large electric storage batteries (used to provide power to reactors and systems and to power vehicles and signs) and lead radiation shielding materials.²⁵ The release of lead and its compounds as reported in the NPRI database for the Pickering plant was from lead acid batteries, which were sent to recycling.⁴⁷

Cadmium

Bruce Power reports that one of the primary sources of waste cadmium is from waste rechargeable batteries.²⁷

Mercury

Bruce Power reports their off-site releases of mercury are from the disposal of electrical relays, switches, industrial lighting bulbs and fluorescent tubes.²⁹ The Pickering plant reports its off-site mercury releases are generated from fluorescent lights sent to disposal.⁴⁷

Chromium

Bruce Power reports that the main source of chromium is the erosion of condenser tubes where chromium is an alloying element in the tube metal.³¹

4.3.1.3.2. Hazardous Wastes

Nuclear generating facilities produce both liquid and solid hazardous wastes as a result of facility operations, maintenance and refurbishment activities. Other wastes result from recycling and disposal of ancillary supplies. The Pickering station, for example reported substantial off-site disposal of asbestos in 2001 and 2002 (see **Table 4.7**).

The Pickering facility was the sixth largest source of transfers of asbestos to disposal in Canada in 2001.³²

Table 4.7: Off-site Releases (i.e. disposal) of Asbestos

 from the Pickering Nuclear Generating Station

Year	Off-Site Disposal (tonnes)
2004	8.6
2003	8.2
2002	44
2001	83
2000	0
1999	11

Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	1.9 x 10 ²	3.7 x 10 ²	1.7 x 10 ²	3.1 x 10 ²	1.7 x 10 ²	1.5 x 10 ²	1.0 x 10 ²
2002	1.5 x 10 ²	4.3 x 10 ²	2.3 x 10 ²	2.8 x 10 ²	1.9 x 10 ²	1.8 x 10 ²	1.3 x 10 ²
2001	2.3 x 10 ²	4.2 x 10 ²	3.1 x 10 ²	2.7 x 10 ²	2.4 x 10 ²	1.9 x 10 ²	1.4 x 10 ²

Table 4.8: Radioactive Air Releases of Tritium Oxide from Nuclear Power Plant Facilities³⁴

Table 4.9: Radioactive Air Releases of Iodine-131 from Nuclear Power Plant Facilities³⁵

Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	2.1 x 10- ⁶	3.2 x 10-⁵	6.4 x 10- ⁵	9.7 x 10-⁵	1.4 x 10-4	n/d	n/d
2002	n/a	4.9 x 10-⁵	6.7 x 10-⁵	9.8 x 10-⁵	1.5 x 10-4	1.4 x 10-7	n/d
2001	n/a	2.8 x 10-⁵	7.8 x 10-⁵	1.0 x 10-4	1.3 x 10-4	n/d	n/d

Note: 'n/d' indicates that radioactive releases were not detected; 'n/a' indicates that the measurement is not applicable.

Table 4.10: Radioactive Air Releases of Elemental Tr	ritium from Nuclear Power Plant Facilities ³⁶
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Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	-	-	-	-	6.6 x 101	-	-
2002	-	-	-	-	5.6 x 101	-	-
2001	-	-	-	-	1.8 x 102	-	-

Note: '-' indicates that data was not reported.

4.3.2. Atmospheric Releases

The operation of nuclear power plants and their associated facilities and equipment results in the release to the atmosphere of a range of air pollutants, including radionuclides, other hazardous air pollutants, smog precursors and greenhouse gases.

4.3.2.1. Radionuclides

The operation of a nuclear generation station results in gaseous radioactive emissions containing tritium in the form of tritium oxide, noble gases, iodine-131, radioactive particulate, and carbon-14.³³ Amounts of radionuclides released vary by nuclear generating station. The annual releases from Canadian facilities are summarized in the following tables.

The significance and health implications of these releases are discussed in section 4.3.6.2.

Total	Bruce-A (TBq)	Bruce- (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	1.4 x 10 ¹	51	2.8 x 10 ²	2.0 x 10 ²	1.3 x 10 ¹	0.71	4.6
2002	n/a	56	2.7 x 10 ²	2.0 x 10 ²	1.5 x 10 ¹	0.69	3.2
2001	n/a	61	2.8 x 10 ²	2.1 x 10 ²	1.8 x 10 ¹	1.9	5.9

Table 4.11: Radioactive Air Releases of Noble Gases from Nuclear Power Plant Facilities³⁷

Note: 'n/a' indicates that the measurement is not applicable.

Table 4.12: Radioactive Air Releases of Radioactive Particulate from Nuclear Power Plant Facilities³⁸

Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	2.9 x 10- ⁶	1.1 x 10-4	3.4 x 10-4	1.6 x 10-⁵	6.9 x 10-⁵	5.4 x 10- ⁶	n/d
2002	4.7 x 10- ⁶	1.1 x 10-4	3.6 x 10-4	2.0 x 10- ⁵	8.7 x 10-⁵	5.0 x 10- ⁶	n/d
2001	4.1 x 10-6	1.4 x 10-4	3.5 x 10-4	2.6 x 10-⁵	5.6 x 10-⁵	8.3 x 10- ⁶	n/d

Note: 'n/d' indicates that radioactive releases were not detected.

Table 4.13: Radioactive Air Releases of Carbon-14 from Nuclear Power Pla	ant Facilities ³⁹
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Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	0.51	4.3	1.1	2.6	3.5	0.39	0.21
2002	0.39	2.1	0.19	1.8	2.8	0.37	0.29
2001	0.39	2.7	0.16	6.3	3.5	0.40	0.22

Total	Bruce (kg)	Pickering (kg)	Darlington (kg)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	4.420	-	0.241	-	-
2003	5.977	-	-	-	-
2002	25.869	-	-	-	-

Table 4.14: Lead Releases to Air from Nuclear Power Plant Facilities

Note: '-' indicates that data was not reported.

Table 4.15: Mercury Releases to Air from Nuclear Power Plant Facilities

Total	Bruce (kg)	Pickering (kg)	Darlington (kg)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	-	0.040	0.021	-	-
2003	-	0.070	0.049	-	-
2002	0.571	0.053	0.027	-	-

Note: '-' indicates that data was not reported.

Table 4.16: Cadmium Releases to Air from Nuclear Power Plant Facilities

Total	Bruce (kg)	Pickering (kg)	Darlington (kg)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	1.408	-	-	-	-
2003	1.826	-	-	-	-
2002	1.880	-	-	-	-

Note: '-' indicates that data was not reported.

Table 4.17: Chromium Releases to Air from Nuclear Power Plant Facilities⁴⁰

Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (tonnes)	Point Lepreau (tonnes)
2004	0.003	-	-	-	-
2003	0.003	-	-	-	-
2002	0.003	-	-	-	-

Note: '-' indicates that data was not reported.

4.3.2.2. Heavy Metals

Lead

Table 4.14 summarizes the lead releases from the five Canadian nuclear power generating facilities as reported to the NPRI.

Mercury

Table 4.15 summarizes the mercury (and its compounds) releases from the five Canadian nuclear power generating facilities as reported to the NPRI.

Cadmium

Table 4.16 summarizes the cadmium (and its compounds) releases from the five Canadian nuclear power generating facilities as reported to the NPRI.

Chromium

Table 4.17 summarizes the chromium releases from the five Canadian nuclear power generating facilities as reported to the NPRI.

Bruce Power reports much higher emissions of heavy metals to the NPRI than do other facilities. Bruce Power indicates that a large portion of these emissions come from the combustion of fossil fuels from the facility's Bruce Alternate Steam Supply (BASS) Plant. The BASS Plant supplies building heating steam and process steam to the off-site Bruce Energy Centre industrial park. The portion of the Bruce facility's heavy metal emissions reported to NPRI attributed to the plant include⁴¹

- 99per cent of nickel
- 91per cent of lead
- 90per cent of mercury
- 88per cent of chromium
- 77per cent of cadmium.

4.3.2.3. Criteria Air Contaminants

The testing of diesel and gas turbine-powered standby and emergency generators at nuclear generating facilities contains carbon monoxide (CO), NOx, SO₂, PM and CO₂.⁴² These generators are tested monthly, with the resultant air emissions displayed in Table **4.18**.⁴³

Bruce Power reports much higher emissions of

criteria air pollutants to the NPRI than shown in **Table 4.18**. In particular, for 2004, the Bruce facility reported to the NPRI releases of 16.2 tonnes of CO, 662 tonnes of SO₂, 145 tonnes of NOx and 48 tonnes of total PM. A large portion of these emissions come from the combustion of fossil fuels from the BASS Plant. The portion of the Bruce facility's criteria air pollutant emissions reported to NPRI attributed to the BASS Plant include⁴⁴

- 98per cent of Total PM (not including road dust)
- 98per cent of SO₂
- 85per cent of CO
- 76per cent of NOx

Bruce Power reports their total PM air emissions to be from various forms of vehicular traffic (passenger vehicles, truck and heavy work equipment) and road dust. Vehicular traffic is stated to contribute 98per cent of all non-road-dust PM.⁴⁵

4.3.2.4. Ammonia and Hydrazine

The steam released from a nuclear power plant contains ammonia and hydrazine.⁴⁶ Ammonia is a breakdown product of hydrazine, which is used in the boilers to prevent boiler corrosion.⁴⁷ Hydrazine is an extremely hazardous chemical; it is a recognized carcinogen, and suspected reproductive, developmental, cardiovascular, neurological and respiratory toxicant.⁴⁸ Gaseous ammonia is classified as a toxic substance for the purposes of the *Canadian Environmental Protection Act.*⁴⁹

The air releases of ammonia reported to the NPRI by nuclear power plants over the past three years are summarized in **Table 4.19**.

Table 4.20 summarizes the air release of hydrazine (and its salts) reported to the NPRI by nuclear power plants over the past three years. Ontario nuclear generating facilities are the only electricity generating facilities in Canada to report discharges of hydrazine to the air.⁵⁰

Based on annual electricity production by Ontario's nuclear power stations and annual releases reported in the NPRI database, an average of 0.5 tonnes of ammonia and 0.03 tonnes of hydrazine are released per TWh of electricity produced at Ontario's nuclear plants.

 Table 4.18: Annual Releases to Air from Ontario's Nuclear Generating Stations

Total	CO (tonnes)	NOx (tonnes)	Total PM (tonnes)	SO ₂ (tonnes)	
2004	5.50	44.27	1.16	14.74	
2003	4.40	91.27	1.73	18.71	

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Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (tonnes)	Point Lepreau (tonnes)
2004	4.944	5.325	12.216	-	-
2003	10.738	5.700	8.595	-	-
2002	3.005	6.500	8.315	-	_

Table 4.19: Ammonia Releases to Air from Nuclear Power Plant Facilities

Note: '-' indicates that data was not reported.

Table 4.20: Hydrazine (and its Salts) Releases to Air from Nuclear Power Plant Facilities

Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (tonnes)	Point Lepreau (tonnes)
2004	0.112	0.010	0.030	-	-
2003	0.042	0.040	0.047	-	-
2002	0.032	0.035	0.035	-	_

Note: '-' indicates that data was not reported.

4.3.3. Water Impacts

The operation of nuclear generating stations results in both routine and accidental radioactive liquid releases in the form of tritium oxide and gross beta-gamma activity.³³ There have also been reported releases of heavy water, which is used as a moderator in the reactor.

Erosion of boiler tubes has resulted in the release of chromium to water at Darlington. Large amounts of copper and zinc have been discharged as a result of the scouring of brass condensers at Pickering.

Nuclear power plants require a large amount of cooling water for steam condensation in their cooling loop. Water is used in the cooling loop and returned to the lake or river from which it was drawn, but at an elevated temperature.

4.3.3.1. Radioactive Releases of radioactive contaminants to water

4.3.3.1.1. Releases to Surface Waters

Releases to surface water of radionuclides occur as part of the normal operation of nuclear power plans. The contaminants of primary concern are tritium oxide and carbon-14.

Tritium is a low energy beta emitter not strong enough to be considered an external radiation hazard (radiation that can penetrate human skin). Tritium can, however, become incorporated into organic molecules and irradiate targets like DNA that can induce mutations and malignant change.⁵¹

Both the USEPA and CNSC report the health risks due to tritium exposure to be minimal. However, tritium is known to adversely affect reproductive outcomes and increase the risk of cancer. The main area of concern for tritium exposure is through ingestion either via contaminated drinking water or tritium in the food chain.⁵²

Currently, the permissible level of tritium in drinking water in Ontario is 7000 Bq/L of water, which was reduced from the previous level of 40,000 Bq/L. This revision reflected the uncertainty surrounding 'safe' levels of tritium in drinking water. The Advisory Committee on Environmental Standards (ACES) in Ontario recommended that acceptable tritium levels be immediately reduced to 100 Bq/L, and ultimately to 1 Bq/L.⁵³ The European Union's current standard is 100 Bq/L and the U.S. federal standard is 740 Bq/L.⁵⁴ In June 2006 the City of Toronto's Board of Health adopted a resolution requesting that the government of Ontario revisit the recommendations of ACES and consider a more protective standard.⁵⁵

Releases of tritium oxide, beta-gamma activity and carbon-14 to surface waters reported to the CNSC by

Canadian nuclear power plants between 2001 and 2003 are summarized in the tables below.

4.3.3.1.2. Groundwater Contamination

In addition to routine releases to surface waters, significant groundwater contamination has occurred at Canadian nuclear facilities. In July 1997, it was revealed that Ontario Hydro, the predecessor to OPG, had failed to report tritium contamination of groundwater on the Pickering site for a period of 20 years. In 1979, 2,150,000 Bq/L of tritium were found in groundwater, and in 1994 Ontario Hydro found 700,000 Bq/L.⁵⁹

4.3.3.1.3. Accidental Releases

Major releases of tritium oxide and other radioactive releases to surface waters have occurred during accidents at Canadian nuclear generating facilities.

In August 1992, a tube break at Pickering-1 caused the release of 2,000 L of heavy water contaminated with 2,300 trillion becquerels of tritium into Lake Ontario. A nearby drinking water plant was shut down and elevated levels of tritium (as high as 195 Bq/L⁶⁰) were found in Toronto drinking water.⁶¹ (Normal levels of tritium are about 7 to 11 Bq/L.) In April 1996, there was a heavy water leak at Pickering-4. Fifty trillion becquerels of tritium were released into Lake Ontario. Levels of tritium in local drinking water reached 100 times background levels.⁶²

Table 4.21: Radioactive Water Releases of Tritium Oxide from Nuclear Power Plant Facilities⁵⁶

Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	6.0 x 10 ¹	8.0 x 10 ²	6.8 x 10 ¹	1.9 x 10 ²	1.0 x 10 ²	3.5 x 10 ²	8.1 x 10 ¹
2002	6.4 x 10 ¹	3.5 x 10 ²	7.7 x 10 ¹	2.1 x 10 ²	6.9 x 10 ¹	5.0 x 10 ²	1.4 x10 ²
2001	1.3 x 10 ¹	1.5 x 10 ²	1.3 x 10 ²	2.0 x 10 ²	9.4 x 10 ¹	4.5 x 10 ²	1.5 x10 ²

Table 4.22: Radioactive Wat	ter Releases of Beta-Ga	amma Activity from Nucle	ear Power Plant Facilities ⁵⁷
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Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	8.8 x 10-4	6.1 x 10- ³	3.0 x 10- ³	7.0 x 10- ³	7.3 x 10- ³	8.6 x 10-4	1.0 x 10- ³
2002	8.1 x 10-4	3.0 x 10- ³	2.9 x 10- ³	1.4 x 10- ²	8.5 x 10- ³	1.3 x 10- ³	3.0 x 10- ³
2001	7.0 x 10- ⁴	2.4 x 10- ³	2.1 x 10- ³	1.1 x 10- ²	5.6 x 10- ³	1.2 x 10- ³	1.3 x 10- ³

Table 4.23: Radioactive Water Releases of Carbon-14 from Nuclear Power Plant Facilities⁵⁸

Total	Bruce-A (TBq)	Bruce-B (TBq)	Pick-A (TBq)	Pick-B (TBq)	Darlington (TBq)	Gentilly-2 (TBq)	Pt. Lepreau (TBq)
2003	1.7 x 10- ³	6.5 x 10- ³	-	1.1 x 10- ²	1.2 x 10- ³	3.0 x 10- ²	1.8 x 10- ³
2002	1.4 x 10- ³	7.1 x 10- ³	-	1.5 x 10- ²	1.7 x 10- ³	2.6 x 10-2	3.4 x 10- ³
2001	6.4 x 10- ³	3.1 x 10- ³	-	3.3 x 10- ³	3.0 x 10- ³	3.4 x 10- ²	2.8 x 10-3

Note: '-' indicates that data was not reported.
Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (kg)	Point Lepreau (kg)
2004	0.001	-	0.430	-	-
2003	0.001	-	0.430	-	-
2002	0.001	-	0.450	-	-

Table 4.24: Water Releases of Chromium from Nuclear Power Plant Facilities

Note: '-' indicates that data was not reported.

Table 4.25: Water Releases of Hydrazine (and its Salts) from Nuclear Power Plants⁶⁸

Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (tonnes)	Point Lepreau (tonnes)
2004	2.491	0.610	0.292	-	-
2003	0.393	0.650	0.240	-	-
2002	1.383	0.497	0.281	-	-

Note: '-' indicates that data was not reported.

Table 4.26: Water	Releases	of Ammo	nia from	Nuclear	Power	Plants ⁶⁹

Total	Bruce (tonnes)	Pickering (tonnes)	Darlington (tonnes)	Gentilly-2 (tonnes)	Point Lepreau (tonnes)
2004	9.258	1.100	3.568	-	-
2003	5.003	0.720	5.492	-	-
2002	3.917	0.670	5.155	_	_

Note: '-' indicates that data was not reported.

4.3.3.2. Metals Discharges

Erosion of boiler tubes results in the release of chromium to water (as has occurred at Darlington).⁶³ In May 1997 it was revealed that the Pickering generating station had released more than 1,000 tonnes of copper and zinc into Lake Ontario from the time of the plant's opening. The releases were the result of scouring of brass condenser tubes in the plant's head exchanger systems.⁶⁴ The condensers at the operating units at Pickering were subsequently replaced with titanium and stainless steel parts to avoid release of copper and zinc.

Discharges of chromium from nuclear power facilities reported to the NPRI between 2002 and 2004 are summarized in **Table 4.24**.

4.3.3.3. Ammonia and Hydrazine Discharges

Nuclear facilities have also reported discharges of ammonia and hydrazine to the NPRI as summarized in **Tables 4.25** and **4.26**. This is a result of the use of hydrazine, which is used in the boilers to prevent corrosion. Hydrazine is an extremely hazardous chemical; it is a recognized carcinogen, and a suspected reproductive, developmental, cardiovascular, neurological and respiratory toxicant.⁶⁵ Ammonia dissolved in water is classified as a toxic substance for the purposes of the *Canadian Environmental Protection Act.*⁶⁶

The Ontario nuclear generating facilities are the only electricity generating facilities in Canada to report to the NPRI discharges of hydrazine to surface waters.⁶⁷

4.3.3.4. Water use and thermal impacts

Nuclear power plants require a large amount of cooling water for steam condensation in their cooling loop. For every kilowatt-hour of electricity they produce, nuclear power plants require between 205 and 228 litres of cooling water.⁷⁰ This means that to power the average Ontario household with nuclear generated electricity for one year requires enough water to fill well over half an Olympic-size swimming pool.⁷¹ With total generation of 42.3 tWh in 2003, the Darlington and Pickering facilities⁷² alone would have used at least 8.9 trillion litres of cooling water, approximately 19 times the annual water consumption of the City of Toronto.⁷³

Water is used in the cooling loop and returned to the lake or river from which it was drawn, but at an elevated temperature. Given the large amounts of water required for cooling, the impact on local fish populations can be significant (see Cooling Water Effects case study).

4.3.4. Landscape Impacts

Considering only the area required for the plant and supporting infrastructure, a 1,000 MWe nuclear generation station requires approximately one to four square kilometres.⁷⁴ While the area required for a nuclear generating plant can be viewed as relatively small compared to many other electricity sources, the life cycle land impacts from nuclear power must be included to make it a fair comparison. Specifically, this not only includes the land required for mining, milling and processing of uranium, but also the impacts to land and biota, particularly outside of uranium mine lease areas.

4.3.5 Heavy Water Production and Use

CANDU reactors require heavy water in operation; other reactor designs use light (i.e., ordinary) water. Heavy water (or deuterium oxide) plays a very important role as a neutron moderator within the reactor. Deuterium is a non-radioactive hydrogen isotope containing an additional neutron, making it twice as heavy as hydrogen.⁷⁵ Deuterium oxide is highly effective as a neutron moderator for its low neutron absorption rate, which allows a chain reaction to occur with natural uranium fuel (with water as the moderator, enriched uranium is required).⁷⁶

Heavy water production is expensive, representing approximately 20per cent of the capital costs of CANDU reactors. However heavy water can be re-used. CANDU reactors need about one metric tonne of heavy water for every megawatt of capacity. CANDU designs under development are intended to require less heavy water.⁷⁷

Heavy water is not radioactive and is only toxic to humans if consumed in large quantities (a level of 10per cent body weight).⁷⁸ Canadian heavy water production used ordinary water as the feed source with a deuterium concentration of ~150 ppm and was produced through the Girdler-Sulphide process. The Girdler-Sulphide process is expensive to operate and requires the use of large quantities of highly toxic H₂S gas. The Girdler-Sulphide process at Bruce required 340,000 pounds of feed water to generate one pound of heavy water.⁷⁹

The last Canadian Girdler-Sulphide process was shut down in 1997. Atomic Energy of Canada Limited (AECL) is currently developing a wet-proofed catalyst heavy water production process that is intended to be less environmentally harmful than the Girdler-Sulphide process.⁷⁷

The Bruce Heavy Water Plant (BHWP) was operated by OPG (formerly Ontario Hydro) and was in operation from April 1973 to March 1998. BHWP produced over 16,000 megagrams (or 16,000 metric tonnes) of reactor grade heavy water. By the end of 1993, the plant had produced enough heavy water to meet Ontario Hydro's needs; one of two operating enriching units was decommissioned, thus reducing plant capacity to 50per cent to support external markets. On August 30, 1997, Ontario Hydro decided to permanently shut down the entire BHWP. Flaring of excess H_2S commenced November 6, 1997 and ended

Case Study: Cooling Water Effects on Fish at the Bruce Power Plant

The Bruce power plant draws large volumes of water from Lake Huron, uses this water in its cooling loop, and returns the water to Lake Huron at an elevated temperature. It is estimated that the plume of elevated temperature water may have warmed 30–50per cent of the local spawning habitat of white-fish.¹ It has been shown in empirical studies that the magnitude of warming from the Bruce generating station may have substantial negative impacts such as increased egg mortality and advanced hatching dates that no longer coincide with zooplankton (food) development.²

on January 23, 1998, flaring a total of 619.9 megagrams (Mg) H_2S (with 1Mg left in storage).⁸⁰ If combusted to SO_2 , under the best case scenario of complete combustion, 619.9 Mg of H_2S would result in 1,166 Mg of SO_2 emissions.⁸¹ In reality, complete combustion is rarely realized and there would be residual H_2S and other sulphur compounds released other than SO_2 .

Historically, the BHWP flared substantial amounts of H_2S during operation. In 1997, over 2,000 tonnes of H_2S was flared. In the early 1980s, an annual average of 1,500 tonnes of H_2S was flared. In the mid-eighties, H_2S flaring was decreased to 500 tonnes annually. From 1987 to the plant's closure in 1997, H_2S flare discharges were approximately 200 tonnes per year.⁸²

A nearby farmer reported central nervous system disorders as well as loss of livestock (300 sheep and lambs) due to H_2S contamination, substantiated by the McMaster University Occupational Health Clinic and a Cornell University atmospheric study.⁸³

4.3.6. Occupational and Community Health

4.3.6.1. Occupational Health Impacts

4.3.6.1.1. Radiation Exposure

In 2002 Ontario's nuclear generating stations employed approximately 11,700 people, including full-time, part-time and contract workers engaged in normal operations and refurbishment projects.

Workers at nuclear generating stations are generally exposed to less radiation than their colleagues working in mining and milling or fuel processing facilities (see Chapters 2 and 3). **Table 4.27** lists average annual whole body doses to workers at Ontario nuclear power reactors. These doses are below the acceptable rate of exposure for nuclear energy workers or members of the public.⁸⁴ It is important to note

Table 4.27:

Average Annual Whole Body Radiation Doses to Workers at Ontario Nuclear Power Reactors⁸⁵

Year	Average annual dosage (mSv)
2002	0.84
2001	0.91
2000	0.81
1999	0.92
1998	0.80

that these are aggregated figures, with the implication that there may be individuals being exposed at higher or lower levels than the average levels.

4.3.6.1.2. Workplace Safety

Between 1995 and 2005, the nuclear generating industry in Ontario reported four fatalities and 508 lost time injuries. The fatalities were attributed to asbestos exposure of electricians, mechanics and pipefitters.⁸⁶

4.3.6.2. Community Health Impacts 4.3.6.2.1. Radiation

The operation of nuclear generating stations results in gaseous radioactive emissions containing tritium in the form of tritium oxide, noble gases, iodine-131, radioactive particulate, and carbon-14, as well as radioactive liquid releases in the form of tritium oxide and gross beta-gamma activity.⁸⁷ Amounts of radionuclides released vary by nuclear generating station.

Members of the public may receive radiation by eating food grown or fish caught near a nuclear power station, drinking water from local sources and being exposed to radioactive material in the air and in the environment.⁸⁸ Maximum annual doses for an individual residing near a nuclear power plant in Ontario, based on environmental transfer models, are estimated at 0.002–0.02 mSv/year.⁸⁹

The CNSC publishes a radiation index based on an analysis of air, water, milk, fish and vegetation as well as actual measurements of emissions from areas surrounding nuclear generating stations. The index represents an estimate of the dose of radiation to the most exposed individuals living near the station. The index conclusions for Ontario from 1998–2004 are shown in **Table 4.28**.

There is considerable scientific debate over the significance of community health impacts arising from incidental radiation releases from nuclear power plants. Plant operators note that the annual radiation indexes indicate levels of exposure for the public well below the CNSC annual public dose limit of 1,000 mSv, and that the exposures resulting from proximity to nuclear power generating facilities are trivial relative to natural and other human-made sources.⁹¹

However, recent work has suggested that ongoing exposure to relatively low levels of radioactivity present in the vicinity of nuclear power plants may present increased health risks to infants and children that had not been previously well understood.⁹² Major concerns continue to be raised regarding the adequacy of Ontario's existing standards for tritium levels in drinking water.⁹³

Table 4.28:

Annual Radiation Index at Ontario Nuclear Reactors (1998-2004)⁹⁰

Year	Pickering (µSv)	Darlington (µSv)	Bruce (µSv)
2004	6.5	1.4	2.5
2003	5.1	1.1	2.3
2002	6.3	1.1	2.8
2001	6.4	1.1	2.6
2000	5.0	2.0	3.3
1999	12.6	3.0	2.4
1998	15.7	4.1	4.6

0.5per cent of the amount guaranteed.⁹⁸ In effect the government of Ontario would assume any shortfall between the value of the decommissioning fund and the actual cost of reactor decommissioning. The Bruce Power Lease Agreement provides that OPG retain responsibility for the eventual decommissioning of the Bruce nuclear reactors.⁹⁹

Table 4.29, taken from the CNSC Regulatory Guide G-219, summarizes typical work packages that would be required for decommissioning a nuclear power plant facility.

4.3.7. Facility Decommissioning

Nuclear generating stations must be decommissioned at the end of their operational lives. This is a major undertaking. As yet, no power reactor in Canada has been fully decommissioned.

The radioactive materials within the generating stations, including spent fuel, and even parts of the plant and machinery that have been exposed to radiation, present serious hazards. The plants must be carefully dismantled and decontaminated, a very costly process that can take more than 30 years. In fact, the first stage of decommissioning may involve shutting down the reactor and doing nothing for five to ten years to let the radioactive materials cool down.⁹⁴ This approach has been taken, for example, with the Douglas Point facility.

Of all stages in the nuclear life cycle, decommissioning of nuclear reactors will produce the most significant quantities of waste. These include highly radioactive materials from the reactor core, as well as other materials and building components contaminated during reactor operation.⁹⁵

The total facility decommissioning costs for OPG's nuclear facilities have been estimated at \$7.474 billion (2003 dollars - estimated present value \$6.263 billion), assuming that reactors would be decommissioned 30 years after their end of life.⁹⁶ A decommissioning fund has been established by OPG. As of June 2006, the fund had a balance of \$4.211 billion.⁹⁷

The Ontario Nuclear Funds Agreement provides that the government of Ontario undertake financial guarantees to the CNSC for OPG's reactor decommissioning liabilities in return for an annual fee of

4.4. Facility Performance

It is important to note that the performance of a large portion of the Ontario reactor fleet has been well below expectations over an extended period of time. The situation resulted in major increases in reliance on electricity sources with high environmental impacts of their own, particularly domestic and imported coalfired generation. The implementation of the Ontario Nuclear Asset Optimization Plan in 1997, involving the shut-down of all eight units at the Pickering and Bruce A stations, led to increased reliance on the province's five coal-fired generating plants for base load electricity supply. According to the OPG, the shut-down was a result of failures in "managerial leadership, culture and standards, people and performance, processes and procedures, plant hardware and design, organization and resources, and labour relations."101

Issues of facility performance and its consequences are discussed in detail in Chapter 6.

4.5. Accidents, Catastrophic Failures and Liability

4.5.1. Accidents

While CANDU reactors are often portrayed as having an inherently safe design, significant accidents have occurred at Canadian facilities. As reactors age the risk of incidents increases.

The experience at the Pickering A Facility, the oldest major nuclear generating facility in Canada, illustrates the types of accidents that may occur. Reported accidents have included the following:¹⁰²

Table 4.29: Typical	Decommissioning	Work Packag	ges ¹⁰⁰
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Planning Envelopes	Work Packages
Calandria vault	 Dismantle calandria internals and shells Decontaminate vault Segment and remove calandria vault
Reactor building	 Remove steam generators Remove primary heat transport pumps and piping Remove moderator dump tanks Dismantle and remove emergency core cooling system Remove fuelling machine and ducts Dismantle and remove internal concrete structures and shielding Remove steel walkways, ladders and stairs Dismantle containment structures and floor slab
Vacuum building and ducts	• Dismantle structures (decontaminate as necessary)
Reactor auxiliary bay	 Remove inventory of irradiated fuel Drain and decontaminate bays Segment and remove bays Remove control centre equipment Remove standby generators Demolish structure
Turbine hall	 Remove turbine generators Remove other electrical and ancillary equipment Demolish structure
Turbine auxiliary bay	 Remove condenser Remove condenser water circulating and service pumps/piping Remove de-aerator Remove feedwater heaters, piping and other equipment Raise structure
Service buildings	 Remove inventory of liquid and solid wastes Decontaminate, dismantle and remove waste management equipment Remove equipment from and decontaminate maintenance shops Remove equipment from and decontaminate laboratories Remove other equipment and materials from stores Demolish structure
Heavy water treatment and storage facility	 Remove inventory of heavy water Remove other equipment and materials Decontaminate and dismantle structures
Water treatment system	 Remove pumphouse Remove water treatment equipment Dismantle structures
Administration building	 Remove contents Dismantle structures
Site	 Remove services, roads, and so on Complete final radiological and contaminants survey Grade and landscape

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- On August 1, 1983, Pickering reactor 2 had a loss of coolant accident after a pressure tube suffered a metre-long rupture. The station was shut down and the four reactors at Pickering A were eventually retubed at a cost of about \$1 billion.
- On November 22, 1988, an operator error damaged 36 fuel bundles. The cooling system was contaminated by radioactive iodine that was vented into the environment over several weeks following the accident.
- On September 25, 1990, Pickering reactor 2 experienced large power shifts in the reactor core. Staff spent two days trying to stabilize it before shutting it down. The AECB later criticized the utility for not shutting down immediately.
- On August 2, 1992, Pickering reactor 1 had a heavy water leak from a heat exchanger that resulted in a release of 2,300 trillion Bq of radioactive tritium into Lake Ontario.
- On April 15, 1996, Pickering reactor 4 had a heavy water leak from a heat exchanger that resulted in a release of 50 trillion Bq of tritium into Lake Ontario.

4.5.2. Catastrophic Failures

Nuclear generating facilities are unique among electricity generating systems in that they have the potential to release large amounts of radioactive materials to the environment in the event of a catastrophic failure. The 1986 Chernobyl accident has been the most extreme example of such an event to date.¹⁰³

It has been noted that, in comparison with the majority of the world's commercial reactors, which are pressurized, light water reactors (PWR), the CANDU heavy water reactor design has several inherent safety advantages. A CANDU reactor has relatively large inventories of both (heavy) water and uranium. If cooling of the reactor core is interrupted for whatever reason, the thermal mass of the (heavy) water and uranium will slow the rate of temperature increase in the reactor core. The separation of cooling water and moderator water in the CANDU design is also seen as a safety advantage.¹⁰⁴

At the same time, it has also been noted that the CANDU design uses a relatively large amount of zirconium alloy in the fuel bundles, almost four times as much as is present in some light water reactors. This large zirconium inventory gives CANDU reactors an inherent safety disadvantage. Zirconium reacts vigorously and exothermically (produces heat) with steam at the high temperatures that would be experienced during a core damage accident. This reaction produces hydrogen gas that can quickly reach explosive concentrations within the reactor containment building. 105

Another area of concern is the potential seriousness of a loss of coolant accident. In light water reactors, the power level declines if the cooling water is lost. In CANDU reactors, like the RBMK reactor in Chernobyl, however, power levels increase when coolant is lost. An accident of particular concern for CANDU reactors involves a loss of coolant accident together with a failure of the reactor shutdown system. If the fast shutdown fails, the power level can rise dramatically. A violent disruption of the reactor core can occur within four to five seconds and release significant quantities of radioactive materials.¹⁰⁶

Assessments of the potential for major accidents at Canadian nuclear generating stations have suggested a very low probability of severe core damage or major releases of radioactive materials. The probability of severe core damage at the Darlington station, for example, has been estimated at 1.5 per 100,000 years of reactor operation.¹⁰⁷ The reliability of these estimates has been challenged.¹⁰⁸ Catastrophic accidents at generating facilities would have impacts over a much wider area and longer term than accidents involving any other electricity generating technology, particularly as a result of the large-scale release of radiation and radionuclides.

Modelling undertaking in the early 1990s estimated that a major accident at the Darlington nuclear generating station east of Toronto, would result in an average collective radiation does of 2.7 million person-Seivert. By comparison, the Chernnobel accident led to a collective radiation does of at least 600,000 to 1.2 million person-Seivert. The monetized value of the offsite consequences of a severe accident at the Darlington facility was estimated to be at least \$1 trillion (1991 \$Cdn).¹⁰⁹

4.5.3. Accident Liability

Currently, the liability of OPG and/or Bruce Power in the event of a nuclear accident is limited to \$75 million by the Nuclear Liability Act. Additionally, manufacturers of nuclear reactor components are exempt from all liability.¹¹⁰ The \$75 million limitation was first stipulated in 1976 and was to be reviewed every five years.¹¹¹

In 2001, officials from Natural Resources Canada pointed out that, accounting for inflation, an equivalent amount would be \$250 million, while international standards at the time were approximately \$650 million.¹¹² The Senate Standing Committee on Energy, Environment and Natural Resources, in its 2002 review of nuclear safety issues in Canada, concluded that, considering emerging international standards, a requirement for liability coverage by operators in the range of \$3 billion would be more appropriate.¹¹³

4.6. Conclusions

Perhaps the most significant environmental impact of nuclear power plant operation in Canada is the generation of approximately 85,000 waste fuel bundles each year. These fuel bundles are extremely radioactive, contain other toxic materials, and require isolation from the environment for one million years. No long-term management strategy for these wastes is in place, although a strategy has been proposed by the Nuclear Waste Management Organization. As of 2003, 1.7 million waste fuel bundles were in storage at nuclear generating facilities in Canada.

In addition, the Ontario facilities generate between 5,500 and 7,000 m³ of low-level radioactive wastes per year as a result of plant operations, maintenance and refurbishment activities. There is no national management strategy for low and intermediate radioactive wastes in place or under development in Canada. OPG is pursing the development of a deep underground disposal facility the low- and intermediate- wastes generated at Ontario facilities.

A wide range of pollutants, including radiological contaminants, persistent organic pollutants, heavy metals and criteria air pollutants have been released to the atmosphere as a result of the incineration of a portion of the low- and intermediate-level wastes generated in Ontario at the Bruce facility. A new incinerator, commissioned in 2003 has subsequently reduced emissions of non-radiological pollutants. Large amounts of waste heavy metals and other hazardous wastes (e.g., asbestos) are also generated as a result of facility operation, maintenance and refurbishment activities.

Large amounts of radioactive and hazardous wastes will be generated from reactor decommissioning, which is expected to occur 30 years after the end of a facility's operational lifetime. The costs of decommissioning Ontario's existing reactors have been estimated at \$7.474 billion (present value \$6.263 billion). The decommissioning fund maintained by OPG currently has a balance of \$4.211 billion. The province of Ontario provides a financial guarantee for any shortfall between the value of the decommissioning fund and the actual costs of decommissioning.

There are routine and accidental releases of tritium oxide and carbon-14 to surface water and groundwater

from nuclear generating facilities. The health and environmental significance of these releases is highly disputed. Ontario's current drinking water quality standards for tritium are substantially weaker than those in place in the United States and European Union.

In addition, Ontario nuclear facilities are the most significant source of discharges of hydrazine, a recognized carcinogen, to surface waters in Canada. There have also been significant discharges of metals (copper, zinc and chromium) as a result of scouring and corrosion of boilers and heat exchangers at the facilities.

Nuclear generating facilities use large amounts of cooling water. The Darlington and Pickering facilities, for example, used at least 8.9 trillion litres in 2003, more than 19 times the annual water consumption of the City of Toronto. Adverse impacts on fish arising from the thermal impacts of large cooling water discharges have been noted.

Routine and accidental releases of tritium oxide, noble gases, iodine-131, radioactive particulates, elemental tritium (Darlington only) and carbon-14 to the atmosphere occur in the course of plant operations. Recent studies have suggested that the community health impacts of these releases, particularly with respect to their effects on infants and children, may be more significant than previously thought.

Ontario nuclear generating facilities are the most significant source of atmospheric releases of hydrazine, an extremely hazardous pollutant, in Canada. Minor releases of criteria air pollutants and GHGs occur as a result of the testing of fossil fuel-powered emergency generating equipment at nuclear generating facilities.

Historically, large scale (up to 2,000 tonnes per year) flaring of H_2S was associated with the production of the heavy water moderator for CANDU reactors. The process also used large amounts of water.

There is a history of accidents, particularly at the Pickering facility, resulting in releases of radioactive materials to the environment, and a potential, although, low probability, of catastrophic accidents at nuclear generating facilities. Catastrophic accidents at generating facilities would have impacts over a much wider area and longer term than accidents involving any other electricity generating technology, particularly as a result of the large-scale release of radiation and radionuclides. The monetized value of the off-site consequences of a severe accident at the Darlington facility, for example, has been estimated to be at least \$1 trillion (1991 \$Cdn).

5. Phase IV: Waste Fuel Management

Summary of Findings

- Approximately 85,000 waste fuel bundles are generated each year in Canadian nuclear reactors. It is estimated there will be 3.6 million waste fuel bundles in storage in Canada by 2033.
- Waste nuclear fuel is extremely hazardous, and remains hazardous over extremely long time frames. It requires management over the very long term (i.e., one million years) to protect human health and the environment and to address the risks of the use of uranium and plutonium contained in waste fuel for nuclear weapons or other damaging purposes.
- No strategy for managing waste nuclear fuel in Canada is in place. The Nuclear Waste Management Organization (NMWO) has proposed a "phased adaptive management strategy" culminating in storage in a deep geological repository, with a total cost of \$24 billion. The proposed strategy would require more than 300 years to be fully implemented. The NWMO proposal is now under consideration by the federal government.
- OPG will be responsible for more than ninety per cent of the waste fuel generated in Canada. OPG's current waste fuel management fund and contributions to the NWMO's trust fund are not sufficient to cover the estimated cost of OPG waste fuel management, with the difference in costs being guaranteed by the province of Ontario. The current arrangements between OPG and the province of Ontario do not reflect the increased amounts of waste fuel that would be generated as a result of the proposed extensions of reactor lifetimes or the construction of new reactors.
- Short-term risks of waste fuel management are seen to be largely associated with the large-scale transportation of waste fuel that would flow from a central repository strategy. As a result of the risk of accidents or other incidents, waste transportation is generally regarded as the most dangerous and highest risk aspect of nuclear waste management.
- The most direct short-term impacts of waste fuel management are the air pollution associated with transportation of waste fuel to a central storage facility and long-term disposal facility. If transportation occurs by road, it would involve 19,000 shipments over a 30-year period, or 53 shipments per month.
- The long-term risks arising from waste fuel disposal are ultimately unknown, given timeframes over which a permanent repository would be required to last.
- Waste nuclear fuel management presents major ethical challenges. Failures of present management options may leave future generations with very large costs and risks.

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5.1. Introduction

About 85,000 used fuel bundles are generated every year by Canadian nuclear reactors.¹ As of 2003, there were 1.7 million fuel bundles in storage at nuclear generators and this number continues to grow.² If existing nuclear facilities continue to operate to the end of their normal operating lives, it is projected that there will be 3.6 million bundles in storage by 2033.³

When a spent fuel rod is removed from a CANDU nuclear reactor, it is extremely radioactive: an unprotected person standing within a metre of such a bundle would die within an hour.⁴ The radioactivity does decrease with time: after one year, it decreases to about 1per cent of its initial value; after 100 years, it decreases to about 0.01per cent of its initial value. However, it takes one million years for the level of radioactivity of spent fuel to return to that of natural uranium.⁵ Spent fuel is also physically extremely hot, and contains toxic chemical elements such as heavy metals.⁶

Spent fuel is currently stored on-site at reactor locations. More than 50 years after the initial decision to develop nuclear power in Canada, a long-term arrangement for the management of waste fuel has yet to be established. Many commentators make strong distinctions between the issue of the management of existing wastes, and the establishment of strategies to manage additional wastes generated by the operation or construction of new nuclear generating facilities. Managing existing wastes is considered the responsibility of current generations; they are the ones who benefited from the energy production that led to their generation. Establishing strategies to manage additional wastes generated in a condition of knowledge of the risks and costs that such choices may impose on future generations, in contrast, is open to serious ethical challenge.7

Fission Product	Half-Life (Years)
Krypton-85	11
Strontium-90	29
Technetium-99	210,000
Tin-126	210,000 ⁸
lodine-129	16,000,000
Cesium-135	2,300,000
Cesium-137	30

Table 5.1: Fission Products in Waste Fuel¹⁰

5.2. What is Waste Nuclear Fuel?⁸

In the CANDU system used in Canada each reactor fuel bundle contains approximately 19 kg of natural uranium, in the form of high-density UO_2 ceramic pellets.⁹ Fuel bundles typically remain in reactors for approximately 18 months. By the time they are removed, the fissile materials are partially depleted, and neutron absorbing fission products and actinides have built up in the fuel bundle.

Waste fuel bundles contain three types of radionuclides: fission products, actinides and waste fuel activation products.

Fission products are formed when neutrons hit and split uranium 235 atoms. The most significant fission products are listed in **Table 5.1**.

Fission products generate large amounts of radiation and heat. As a result, the fuel bundles have to be handled remotely, and must be shielded and cooled when first removed from the reactor. When initially removed from a reactor, each fuel bundle gives off more than 25,000 watts of heat energy.

The second type of radionuclides are actinides, nuclides of heavy elements that absorb neutrons, but do not split. The main actinides contained in used fuel are listed in **Table 5.2**. Actinides tend to be highly radioactive and have long half-lives.¹¹

Finally, waste fuel contains activation products, the result of neutron reactions with materials in the fuel cladding rather than the fuel itself. Examples of some of these products are shown in **Table 5.3**.

The overall composition of waste fuel compared with fresh fuel is summarized in **Table 5.4**. About 30 per cent of the energy derived from the fuel bundles is

Actinide	Half-Life (Years)
Uranium-235	710,000,000
Uranium-236	23,000,000
Uranium-238	4,500,000,000
Plutonium-239	24,000
Plutonium-240	6,600
Plutonium-242	360,000
Neptunium-237	2,100,000
Americium-241	460
Thorium-232	1,400,000,000

Table 5.2: Actinides Contained in Waste Fuel¹²

Tal	ble	5.3:	Waste	Fuel	Activation	Products ¹³
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lsotope	Half-Life (Years)	
Carbon-14	5,700	
Chlorine-36	300,000	
Zirconium-93	1,500,000	

Table 5.4:

Composition of Waste Fuel Compared to Fresh Fuel¹⁴

Component	Composition of Fresh Fuel, %	Composition of Used Fuel, %
Uranium-235	0.72	0.23
Uranium-236	0	0.07
Uranium-238	99.28	98.58
Plutonium-239	0	0.25
Plutonium-240	0	0.10
Plutonium-241	0	0.02
Plutonium-242	0	0.01
Fission products	-	0.74

Figure 5.1:

Total Radioactivity Per Bundle of Used Fuel¹⁵



derived from the fissioning of plutonium.

Radioactive decay continues when the used fuel is removed from the reactor, causing emissions of radiation and heat at decreasing rates and changing composition over time. Of the 350 different isotopes present, approximately 200 are radioactive. The total radioactivity per bundle of used fuel is summarized in **Figure 5.1** below.

The level of radioactivity declines rapidly at first, then tails off. Activity declines to that of natural uranium and its associated decay products after approximately one million years. It is assumed that waste fuel will need to be isolated from the environment until its radioactivity reaches that level.

In addition to the hazards of radioactivity and heat, used nuclear fuel has the potential to release chemically toxic elements, including heavy metals. Uranium itself has greater chemical toxicity than radiotoxicity; once it decays it becomes lead, which is also toxic. Other elemental products of radioactive decay are rare, but do exist, and little is known about their environmental behaviour. In addition to the trace elements that appear from the decay of radionuclides, there are a number of trace elements present in the fuel cladding or containment vessels. If these contaminants move into ground or surface waters or enter the atmosphere and are taken up by organisms, they can cause harm.¹⁶

In addition to the immediate safety, health and environmental risks posed by waste nuclear fuel, the waste fuel contains materials (uranium-235 and plutonium-239) that can be used for nuclear weapons production. These fissile materials are highly attractive materials for countries or individuals seeking to develop nuclear weapons.¹⁷ Although the radioactivity of fresh waste fuel makes the extraction of plutonium difficult, after some decades radioactivity falls sufficiently to make extraction relatively easy.¹⁸ This implies that waste fuel not only needs to be isolated from the environment for a time period of approximately one million years, but also needs to be kept secure from deliberate human disturbance for weapons development purposes.

These environment and security requirements clearly present major technological and managerial challenges. Waste nuclear fuel management also presents major ethical issues. The risks and costs of potential failures of any management strategy chosen may fall on generations far into the future, rather than on current generations that consumed the electricity associated with the production of the waste fuel.

5.3. Current Management Practices

All nuclear fuel waste in Canada is currently in "interim storage" at reactor sites pending a resolution of the question of long-term management. When the waste fuel is removed from the reactor it is initially placed in water-filled pools, where it remains until its heat and radioactivity decline. After approximately seven to ten years, the bundles are transferred to dry storage at the reactor sites. Dry storage containers are designed to last 50 years, at which point they have to be replaced.¹⁹

The total amounts of waste fuel in storage in Canada as of December 2002 are shown in **Table 5.5**.

5.4. Waste Fuel Management Initiatives

The decision to pursue the development of nuclear energy for electricity generation in Canada was made in the 1950s, and the construction of generating facilities was initiated in the 1960s.²¹ However, serious consideration of how to manage waste fuel did not begin in Canada until the early 1970s, after the first generating facilities had been brought into service.

In 1977, the federal department of Energy, Mines and Resources engaged an expert group under the chair of Dr. Kenneth Hare to examine potential management methods. The resulting "Hare Report" concluded that deep burial in the Canadian Shield would

Storage Location	Licencee	Fuel Bundles in Reactor	Used Fuel Bundles in Wet Storage	Used Fuel Bundles in Dry Storage	Total Fuel Bundles
ONTARIO					
Bruce A (1)	Bruce Power Corporation	0	354,567	0	354,567
Bruce B (1)	Bruce Power Corporation	24,679	356,519	0	381,198
Pickering (2)	Ontario Power Generation	96,796	393,690	99,106	529,552
Darlington (3)	Ontario Power Generation	24,960	211,932	0	238,892
Douglas Point (13)	AECL	0		22,256	22,256
Chalk River Laboratories (12)	AECL	0		4,853	4,853
(used fuel from Rolphton Nuclear Power Demonstration)					
QUEBEC					
Gentilly 1 (14)	AECL	0	0	3,213	3,213
Gentilly2 (4)	Hydro-Quebec	4,560	37,181	48,000	89,741
NEW BRUNSWICK					
Point Lepreau (5)	New Brunswick Power	4,560	40,482	52,920	97,962
MANITOBA					
Whiteshell Laboratories (15)	AECL	0		360	360
(used fuel from Douglas Point and non-standard waste)					
TOTAL		95,515	1,394,371	230,708	1,720,594

Table 5.5: Waste Nuclear Fuel in Storage in Canada, December 2002²⁰

be the best approach. Subsequently, Atomic Energy of Canada was given responsibility for researching and developing "disposal in a deep underground repository in intrusive igneous rock."²²

Between 1978 and 1995, \$538 million was spent on research and development of the concept of deep geological disposal of nuclear waste in hard rock. Of this amount, about \$370 million was federal funding to AECL and about \$133 million was provided by Ontario Hydro, with the balance coming from other sources, primarily foreign waste management research agencies.²³

In the late 1980s the concept of deep disposal in the Canadian Shield was examined under the federal Environmental Assessment Review Process. The Environmental Assessment Panel, chaired by Blair Seaborn, tabled its report in 1988. The panel's conclusions were as follows:²⁴

- "From a technical perspective, safety of the AECL concept has been on balance adequately demonstrated for a conceptual stage of development, but from a social perspective, it has not.
- As it stands, the AECL concept for deep geological disposal has not been demonstrated to have broad public support. The concept in its current form does not have the required level of acceptability to be adopted as Canada's approach for managing nuclear fuel wastes."

The panel identified a number of additional steps that would be required to develop an approach for managing nuclear fuel wastes in a way that could achieve broad public support. These included

- issuing a policy statement on managing nuclear fuel wastes;
- initiating an Aboriginal participation process;
- creating a nuclear fuel waste management agency (NFWMA);
- conducting a public review of AECB regulatory documents using a more effective consultation process;
- developing a comprehensive public participation plan;
- developing an ethical and social assessment framework; and
- developing and comparing options for managing nuclear fuel wastes.

The panel made the following basic recommendations to governments with respect to a management agency:

• that an NFWMA be established quickly, at arm's length from the utilities and AECL, with the sole purpose of managing and co-ordinating the full

range of activities relating to the long-term management of nuclear fuel wastes;

- that it be fully funded in all its operations from a segregated fund to which only the producers and owners of nuclear fuel wastes would contribute;
- that its board of directors, appointed by the federal government, be representative of key stakeholders;
- that it have a strong and active advisory council representative of a wide variety of interested parties;
- that its purposes, responsibilities and accountability, particularly in relation to the ownership of the wastes, be clearly and explicitly spelled out, preferably in legislation or in its charter of incorporation; and
- that it be subject to multiple oversight mechanisms, including federal regulatory control, with respect to its scientific-technical work and the adequacy of its financial guarantees; to policy direction from the federal government; and to regular public review, preferably by Parliament.

The panel concluded that, until these steps had been completed and broad public acceptance of a nuclear fuel waste management approach had been achieved, the search for a specific site should not proceed.

In 2002, the *Nuclear Fuel Waste Act* was enacted, establishing a Nuclear Waste Management Organization (NWMO) to investigate approaches to managing Canada's spent nuclear fuel. The legislation mandated the NWMO to make a recommendation to the federal Minister of Natural Resources regarding a preferred option for the long-term management of waste nuclear fuel by November 2005. In developing its recommendation, the NWMO was directed by the legislation to consider three options:

- deep geological disposal in the Canadian Shield, based on the AECL concept
- storage at reactor sites
- centralized storage, either above or below ground. Between 2002 and 2005, the NWMO conducted an extensive study during which it engaged in consulta-

tions with experts, stakeholders and citizens, including Aboriginal peoples.

In November 2005, the NWMO submitted its final study report to the Minister of Natural Resources, recommending a strategy of Adaptive Phased Management. ²⁵ This strategy consists of three phases:²⁶

1. Preparing for central used fuel management (30 years)

- Continue use of storage reactor sites while final disposal site is selected and approved
- Construct underground characterization facility
- Conduct shallow storage if this option is chosen
- Central storage and technology demonstration (30 years)
 - Begin transport of used fuel to shallow storage (if selected)
 - Research and demonstrate to confirm site suitability
 - Assess site, technology and timing for used fuel placement
 - · Construct and licence deep repository
- 3. Long-term containment in a deep geological repository in the Canadian Shield or Ordovician sedimentary rock (beyond 60 years)
 - Move used fuel from shallow storage or reactors for repackaging
 - · Place used fuel in containers in deep repository
 - · Monitor and maintain access
 - Leave future generation to decide when to close repository and the nature of ongoing monitoring.

The planned timeframes for this project are extremely long. The geological storage phase would not begin for approximately 60 years, and complete implementation of the strategy would extend over more than 300 years.²⁷ The NWMO states that it intends to seek an informed, willing community host for the central facilities. NWMO estimates that the cost of its proposed waste fuel management strategy will be \$24 billion (2002\$) over the life of the project, equivalent to \$6.1 billion in 2004 dollars.²⁸

The NWMO's recommendations highlight the fact that Canada remains a considerable distance from actual implementation of a long-term strategy to manage waste nuclear fuel. A number of factors need to be taken into account before this situation can be addressed.

First, the NMWO's recommendations have yet to be accepted by the federal government. The process of seeking approvals for specific activities or facilities cannot begin until the recommendations are accepted. The proposed strategy would be subject to a comprehensive environmental assessment under the *Canadian Environmental Assessment Act.*²⁹

Secondly, the legitimacy of the NWMO's recommendations and consultative processes has been challenged by a range of stakeholders. It has been pointed out that the NWMO board is made up of representatives of Canada's nuclear corporations, in contradiction to the Seaborn Panel's recommendation of an arm's length agency. The NWMO's failure to address the question of the long-term role of nuclear energy in Canada has also been a significant source of criticism.³⁰

In addition, the NWMO will need to identify a willing community in Ontario, Quebec, New Brunswick or Saskatchewan to host a permanent facility. This may be a significant challenge. Saskatchewan Premier Lorne Calvert has indicated that he will not allow a nuclear waste disposal facility to be located in Saskatchewan.³¹ Ontario Natural Resources Minister David Ramsay has stated: "We don't like the idea of nuclear waste coming to Northern Ontario."³² Manitoba has enacted legislation prohibiting the storage of high-level radioactive waste generated outside of Manitoba, and prohibiting the establishment of high-level waste disposal facilities within the province.³³

5.5. Impacts of Waste Fuel Management

Table 5.6 provides an overview of the environmental and health impacts and risks associated with nuclear waste fuel management. The potential impacts outlined are based on the premise of the implementation of the NWMO's recommendations.

It is important to note that the NWMO itself points out that the performance of its proposed disposal technology in the long term is unknown due to the impossibility of forecasting thousands of years into the future.³⁴ It is known, for example, that in the long-term disposal facilities would be exposed to environmental stresses beyond current experience, such as ice ages and other extreme climate changes.³⁵

In the short term, the largest environmental and occupational and community health risks associated with waste fuel management are seen to involve the potential for accidents during the transport of waste fuel to centralized management facilities.³⁶ Waste fuel transportation would itself be an enormous undertaking, estimated to require more than 50 truck trips per month over a period of 30 years.³⁷ The air pollution emissions associated with these transportation activities would constitute the most direct short-term environmental impact of waste fuel management activities.

Table 5.6: An Overview of the Environmental and Health Impacts and Risks Associated with	
Nuclear Waste Fuel Management	

Waste	Atmospheric	Water Impacts	Landscape	Occupational and Community
Generation	Releases		Impacts	Health Impacts
• Radiation, heat toxicity, pro- liferation risks from waste fuel	 Impacts of transportation to central stor- age and dis- posal sites 	 Risk of ground- water impacts from deep geological storage (no prospective locations identified) 	 Storage of spent fuel means perma- nent footprint of facilities 	 Transportation accidents Radiation exposure of workers Potential impacts on communities chosen for waste storage facility Potential environmental health and weapons proliferation risks to future generations

Table 5.7: Estimate of Requirements for Transportation of Spent Fuel to a Deep Geological Storage Facility⁴²

Transportation Mode	Number of Shipments/Month for the First 30 Years	Total Number of Shipments for the First 30 Years	
Road	53	19,000	
Rail/Road	5 rail and 36 road	1,800 rail 12,960 road	
Water/Road	2 water and 26 road	720 water 9,360 road	

5.5.1.Waste Generation

About 85,000 used fuel bundles are generated every year by Canadian nuclear reactors.³⁸ As of 2003, there were 1.7 million fuel bundles in storage at nuclear generators and this number continues to grow.³⁹ Waste fuel accumulates from generating stations as well as from research, prototype and demonstration reactors. If current rates of generation continue, it is expected that there will be 3.6 million bundles in storage by 2033, when the existing reactor fleet reaches the end of its normal operative life.⁴⁰ These figures do not account for the wastes that might be generated by new reactors, or reactors whose operating live is extended via refurbishment projects.

Wastes will also be generated from the construction of permanent disposal facilities. If a deep underground repository is located in the Canadian Shield, construction may generate significant amounts of waste rock that may be subject to acid drainage.

5.5.2. Atmospheric Impacts

Canada's current approach to nuclear waste fuel management does not require transportation. All nuclear waste fuel is stored on the site where it is generated and is moved to the appropriate location on-site by a specialized transporter.

The Adaptive Phased Management plan proposed by the NWMO requires transportation of the spent fuel from the reactor sites to a centralized storage facility, if used, and from there to a deep geological storage facility. Air emissions associated with this management plan would thus be due to transportation and construction of the various facilities.

The transportation of nuclear waste fuel under the Adaptive Phased Management plan could occur by road, rail or water depending on the location of the eventual site for the storage facility. The transportation requirements estimated by the NWMO, based on a fuel inventory of 3.6 million bundles, are summarized in **Table 5.7**. The table makes it clear that transportation activities would be an enormous undertaking of their own, potentially requiring more than 50 truck trips per month over a period of 30 years.

The figures are given for the first 30 years, based on the amount of time required to transport the currently accumulated waste plus the ongoing waste accumulation. One shipment by truck contains 192 bundles and has a full payload of approximately 42.5 tonnes. One shipment by water contains 5,760 bundles, which would weigh 1,015 tonnes including transport containers.⁴¹

To facilitate initial cost-benefit studies of the Adaptive Phased Management approach, the NWMO used four illustrative examples of sites for nuclear waste disposal. The sites are not repre-

sentative of actual sites, but rather represent a range of characteristics. The four sites have average distances of 260, 1,000, 2,000 and 3,500 km from current reactor sites.⁴³ Based on the waste inventories and hypothetical distances to the example waste disposal sites, estimates of total distances traveled to move all currently accumulated waste plus ongoing waste accumulation ranges from five million to 69 million kilometres.

It is worth noting that the scenario of five million total kilometres is unlikely to gain public acceptance as this waste disposal site would be located in an area with a medium population density near the Bruce Peninsula in southern Ontario.⁴⁴ We suggest that, were a suitable disposal site actually found, total distance traveled for disposal would likely be closer to the 69 million kilometre estimate.

5.5.2.1. GHG, PM, and NOx Emissions

Estimates of GHG, PM, and NOx air emissions resulting from transport of spent nuclear to disposal sites are reported below in **Table 5.8**. Emissions are calculated for the shortest (five million) and longest (69 million) transport distances estimated by the NWMO.

To provide some context, the GHG and PM emissions in **Table 5.8** are roughly equivalent to the annual emissions of 1,500 and 21,000 passenger cars for the transportation of spent fuel five million and 69 million kilometres, respectively. The NOx emissions are equivalent to the amount annually produced by 10,500 and 145,000 passenger cars for five million and 69 million kilometre distances, respectively.

In addition to their air pollution impacts, waste transport is generally regarded as the most dangerous, highest risk aspect of nuclear waste management.⁴⁹ The greater the distance transportation nuclear fuel waste is transported, the greater the potential for an accident that would release radioactivity to the surrounding area. The issue of accident risks and their potential consequences is discussed in section 5.5.5.2.

Table 5.8:
Total Estimated Air Emissions from Spent Fuel Transport Scenarios ^{45,46,47,48}

Scenario: Total km Traveled	GHG Emissions (tonnes)	PM Emissions (tonnes)	NOx Emissions (tonnes)
Closest: 5 million	5,200	1.1	39.2
Farthest: 69 million	72,000	15.0	540.0

5.5.3. Water Impacts

One of the principle critiques of the deep geological disposal option that would form the final phase of the strategy proposed by the NWMO is that no container can remain water resistant for hundreds of years. It has been argued that groundwater will find its way into any vault, no matter how well designed, and will eventually seep in and out of the nuclear waste containers, carrying radioactive substances. This contaminated groundwater may then show up in wells, springs, lakes, and rivers, and be taken up into the food chain. It has also been pointed out that it is impossible to predict geological activity with certainty even in stable geological formations.⁵⁰

NWMO argues that below 500 metres groundwater is very saline, reducing and old, and therefore can be considered stagnant over the period of concern for a repository facility (i.e., one million years). The plutonic rock of the Canadian Shield has this attribute, as do bedded salts and shales.⁵¹ However, the validity of modeling groundwater behaviour over such long time periods has been challenged.⁵² It has also been pointed out that the option provides little flexibility for future generations to influence the management of waste fuel, or to make fundamental changes without incurring considerable additional costs.⁵³

5.5.4. Landscape Impacts

A deep geologic storage facility will have to be a permanent structure that will last virtually in perpetuity. Therefore, the land footprint of such a facility will also be permanent. However, the area of land required for these facilities will likely be fairly small. Of greater concern is the potential impact to land, human health and surrounding biota in the event of a catastrophic failure.

Depending on the location of the facility and the mode of transportation chosen, the construction of roads or rail lines would also impact the land and associated biophysical environment.

5.5.5. Occupational and Community Health Impacts

5.5.5.1. Occupational Health

Implementation of the nuclear waste management strategy is expected to bring significant employment and income benefits to the local host economic region. It is estimated that thousands of jobs could be created as a result of the implementation of the strategy.⁵⁴ However, like all jobs in the nuclear industry, these jobs would require exposure to a higher rate of radiation than is acceptable for the general public.

A 2002 report for the State of Nevada estimated that there could be significant radiation exposure to drivers and others involved in the transport of nuclear waste, as well as to members of the public on transportation routes. While the report is specific to the proposed plan to transport nuclear fuel waste in Nevada, it is nonetheless instructive in highlighting the potential exposure risks associated with transport of nuclear fuel waste in Canada. The likely radiation exposures risks associated with waste fuel transportation as identified by the report are outlined in **Table 5.9**.

It is instructive to compare these values with the regulated dose limits for the general public and nuclear energy workers of 1.0 mSv/year and 50 mSv/ year, respectively, as referenced in **Table 1.3**.

In general, one hour of exposure to an average transport cask at a distance of two metres exposes an individual to a dose equivalent to a whole body X-ray.⁵⁷

5.5.5.2. Accident Risks

The transportation of waste nuclear fuel is generally regarded as the most dangerous, highest risk aspect of nuclear waste management due to the risk of accidents. 58

In the event of a transportation accident, people who live and work near the accident site will be severely impacted. The NWMO argues that radioactive materials have been transported around the world for 40 years, with no accidents that have resulted in the release of "significant amounts" of radioactivity.⁵⁹ However, because the frequency of transport required under the Adaptive Phased Management plan is currently unknown, the eventual plan may result in more transportation than currently occurs; this will bring with it an increased likelihood of accidents.

The State of Nevada contracted Radioactive Waste Management Associates (RWMA) to model the consequences of a nuclear waste fuel transportation accident. Using accident scales of the US Department of Energy (DOE), the RWMA chose to focus on category five accidents (out of six categories) because these were deemed to be severe, yet realistic scenarios, and chose Los Angeles as the site of a potential urban accident. The RWMA found that the results of such an accident would be serious. Inhalation of radioactive particles would expose thousands of individuals to 1,000 mSv, several hundred times the background radiation in the area. Radioactive particles could infiltrate ventilation systems, causing particles to settle on furniture and in rugs. Radioactive particles would also be spread by vehicles and people traveling away from the accident site. The RWMA found that this type of accident would overwhelm the emergency response team and the medical community in Los Angeles.⁶⁰

It has also been pointed out that waste fuel storage installations at individual reactor sites may be targets for attack by groups or individuals motivated by political purpose, insanity or both. It has been suggested that, if such an attack were to lead to the loss of water from a spent-fuel pool with high density racks, it could lead to a runaway zirconium-air or zirconium-steam reaction;

Table 5.9: Estimated Exposure Risks Associated with Spent Nuclear FuelTransport in Nevada⁵⁵

Occupation	Exposure
Safety inspector	85 mSv/year
Truck driver	40 mSv/year
Service station attendant on transport route	5-10 mSv/year
Driver in gridlock with transport truck (4 hours)	0.04–0.08 mSv/hour
Average American exposure from X-rays, etc. ⁵⁶	0.65 mSv/year

the resulting heat production and fuel degradation would release a large amount of radioactive material to the atmosphere.⁶¹

5.6. Proliferation and Security Issues

While it cannot ignite or explode, nuclear waste fuel is highly radioactive and contains uranium and plutonium. These fissile materials are highly attractive materials for countries or groups seeking to develop nuclear weapons.⁶²

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In 1996, researchers from the US National Nuclear Security Administration estimated that "a determined group of six persons who have familiarized themselves with the unclassified technical literature could produce 1 significant quantity (SQ)⁶³ of plutonium metal eight weeks after the receipt of a sufficient quantity of spent enriched uranium fuel."⁶⁴ Since CANDU fuel is made of un-enriched uranium, 2.5 tonnes of it would be necessary to produce 1 SQ of plutonium. Obtaining such an amount of spent CANDU fuel would not be easy. Nevertheless, there are significant security risks associated with the transportation and storage of spent nuclear fuel.^{65,66}

The weapons proliferation risks associated with nuclear waste fuel were demonstrated on May 17, 1974, when India detonated a 12-kiloton nuclear bomb in the Rajasthan desert. The bomb was allegedly built using plutonium derived from Canadian-supplied uranium in a Cirus research reactor donated by Canada in 1956 under the Commonwealth "Colombo Plan" aid program to promote economic and social development in south and southeast Asia.⁶⁷

In the Adaptive Phased Management plan proposed by the NWMO, used nuclear fuel bundles will have to be transported from the reactor sites to the centralized facilities, and then from the centralized facilities to the geological storage facilities, thereby increasing the associated risks of transporting nuclear waste.

While the NWMO claims that transport of nuclear fuel waste is safe, the potential impacts of a terrorist attack or sabotage are substantial. A 1999 study for the US DOE found that the casks in which nuclear fuel is transported are vulnerable to high energy explosive attacks, as could be used in a terrorist attack.⁶⁸ Similarly, the NWMO report states that the casks are designed to withstand "expected accident conditions."⁶⁹ These do not include terrorist attacks.

The draft Environmental Impact Statement prepared by the US DOE for nuclear waste transport in Nevada estimated that a successful attack on a truck cask in an urbanized area would result in about 15 cancer fatalities among those exposed to the release of radioactive materials. An analysis prepared for Nevada by the Radioactive Waste Management Agency (RWMA) estimated sabotage impacts could be significantlygreater than the US DOE's estimate (6–104 fatalities). Under worst case weather conditions, the RWMA estimated 4–165 latent cancer fatalities would result. Further, cleanup costs and other economic impacts were estimated at \$3.1–13.5 billion (2000\$) for average weather conditions, and \$10.1–20.9 billion (2000\$) for worst case weather conditions.⁷⁰ Concern has also been raised that long-term waste disposal sites could become the "plutonium mines of the future," raising longer-term proliferation issues. Plutonium in fresh spent fuel is "protected" by the high radioactivity of the material. This radioactivity decays after some decades, whereupon separation of plutonium from the fuel would become relatively simple.⁷¹

5.7. Costs and Liabilities

Between 1978 and 1995, \$538 million was spent on research and development of the concept of deep geological disposal of nuclear waste in hard rock. Of this amount, about \$370 million was federal funding to AECL and about \$133 million was provided by Ontario Hydro, with the balance coming from other sources, primarily foreign waste management research agencies.⁷²

NWMO estimates that the cost of its proposed long-term spent fuel management strategy will be \$24 billion over the life of the project (2002\$).⁷³

The project will be funded by the entities that operate nuclear reactors and produce waste nuclear fuel. Under the provisions of the Nuclear Waste Fuel Management Act, a trust fund has been established for this purpose, to which OPG has deposited \$500 million, Hydro Quebec and New Brunswick Power have each deposited \$20 million, and AECL has deposited \$10 million.74 Additionally, each utility makes an annual contribution to the fund: \$100 million by Ontario Power Generation), \$4 million by Hydro Quebec, \$4 million by New Brunswick Power and \$2 million by AECL. These funds can not be used until a construction license has been granted for the waste management approach recommended by the NWMO. As of November 2005, total contributions to the trust fund were \$880 million.75

It is expected that OPG will responsible for over 90 per cent of the waste nuclear fuel generated in Canada⁷⁶ OPG has provided the CNSC with a financial guarantee of \$4.5 billion (present value as of January 1, 2005) for waste fuel management, reflecting the estimated present value costs of these activities.⁷⁷ OPG's Used Fuel Fund currently contains assets of \$2.985 billion,⁷⁸ while OPG had contributed \$807 million to the NWMO trust as of March 31, 2005. Under the Ontario Nuclear Funds Agreement, the provincial government has provided an unconditional guarantee to cover any shortfall between OPG's financial guarantee to CNSC, and the funds contained in OPG's Used Fuel Fund and its contributions to the NWMO trust. As of 2005, the province's guarantee to the Canadian Nuclear Safety Commission on behalf of OPG amounted \$1.5 billion.⁷⁹

More generally, the Ontario Nuclear Funds Agreement, between OPG and the province of Ontario, limits OPG's liabilities for disposal of used nuclear fuel. The agreement stipulates that OPG will be responsible for all costs (present value as of January 1, 1999) up to \$4.6 billion; will share costs on an equal basis between \$4.6 and \$6.6 billion; and will be responsible for 10 per cent of the costs incurred between \$6.6 and \$10 billion, while the province will be responsible for any costs above \$10 billion.⁸⁰

It is important to note that the \$4.5 billion waste fuel management cost estimates on which the ONFA was based were premised on the 1.76 million waste fuel bundles estimated to be generated by the end of 2005.⁸¹ In contrast, NWMO's estimated \$6.1 billion present value cost of the its proposed waste management strategy, was premised on a 40 year reactor life, resulting the generation of 3.6 million waste fuel bundles of which 3.3 million would be the responsibility of OPG.⁸² The provisions of the ONFA will need to be adjusted to reflect the increased waste management costs associated with the extension of reactor lifetimes, or the province of Ontario may be faced with significant additional liabilities for waste fuel management.

The NWMO's cost estimates also assume that the long-term management project would be completed on budget and on time. Given the complexity and long timelines associated with the project, these may not be valid assumptions.

5.8. Conclusions

Approximately 85,000 waste fuel bundles generated each year in Canadian nuclear reactors. It is estimated there will be 3.6 million waste fuel bundles in storage in Canada by 2033. This figure assumes that no new nuclear generating facilities are brought into service. Waste nuclear fuel is extremely hazardous. It remains hazardous over extremely long time frames, and requires management over the very long term (i.e., one million years) to protect human health and the environment, and to address the risks of the use of uranium and plutonium contained in waste fuel for nuclear weapons or other damaging purposes.

No strategy is in place for managing nuclear waste fuel in Canada. The NMWO, established in 2002, has proposed a "phased adaptive management strategy" extending over more than a century, and culminating in storage of the fuel in a deep geological repository, with a total estimated cost of \$24 billion. The strategy has yet to be approved by the federal government, and a location for a permanent facility has not been identified. The strategy would require more than 300 years to fully implement.

OPG will be responsible for more than ninety per cent of the waste fuel generated in Canada. OPG's current waste fuel management fund and contributions to the NWMO's trust fund are not sufficient to cover the estimated present value cost of OPG waste fuel management, with the difference in costs being guaranteed by the province of Ontario. The current arrangements between OPG and the province of Ontario do not reflect the increased amounts of waste fuel that would be generated as a result of the proposed extensions of reactor lifetimes or the construction of new reactors.

The short-term risks related to waste fuel management are seen to be largely associated with the large-scale transportation of waste fuel that would flow from a central repository strategy. Waste transportation is generally regarded as the most dangerous and highest risk aspect of nuclear waste management due to the risks of accidents or other incidents during transportation.

The most direct short-term impact of waste fuel management is the air pollution associated with transportation of waste fuel to a central storage facility and long-term disposal facility. If transportation occurs by road, it would involve 19,000 shipments over a 30-year period, or 53 shipments per month.

The long-term environmental risks associated with waste fuel management are unknown, given the timeframes over which a permanent repository would be required to last.

Waste nuclear fuel management presents major ethical challenges. Failures of the management options chosen in the present may burden future generations with very large costs and risks arising from wastes associated with past generations' consumption of energy produced from nuclear facilities. Strong distinctions are made between the management of existing wastes—current generations benefited from the energy production that led to their generation and thus have a responsibility to ensure their safe management—and the generation of additional wastes via the operation or construction of new nuclear generating facilities—there is knowledge of the risks and costs that such choices may impose on future generations.

6. Sustainability Challenges

Summary of Findings

In addition to its physical environmental impacts, particularly waste generation, nuclear energy faces a number of unique economic, security and policy challenges.

- Generating facilities are subject to very high capital costs, and long construction times. This makes it difficult for such facilities to compete for private capital investments against potential investments that will bring more rapid and secure returns. Extraordinary financial guarantees by governments have been required to overcome these barriers.
- In Ontario there is a history of serious delays and cost overruns on nuclear generating facility projects, accounting for a large portion of the "stranded debt" left by Ontario Hydro.
- The Ontario CANDU fleet has been plagued by performance and maintenance problems. In recent years average fleet capacity has been in the 50 per cent range rather than the expected 85–90 per cent range. Reactors expected to have operational lifetimes in the range of 40 years turned out to require major refurbishments after approximately 25 years of service. Refurbishment projects themselves have run seriously over budget and behind schedule.
- The poor performance of the Ontario CANDU fleet has had major collateral environmental and health impacts. These impacts have been the result of the province's increased reliance on domestic and imported coal-fired generation to replace electricity that would have been provided by out-of-service nuclear units.
- Fuel costs are emerging as a significant issue for the nuclear industry. World uranium prices have increased by more than a factor of six since 2001 and are expected to continue to rise.
- The question of the long-term fuel supply for nuclear generating facilities is emerging as a concern, particularly if there is a large-scale expansion of reliance on nuclear generation. Current Canadian uranium reserves are estimated to be sufficient for 40 years at present levels of consumption, significantly less than other non-renewable fuels such as natural gas and coal. Efforts to increase the available fuel supply through the reprocessing of waste fuel, or the use of fast breeder reactors (FBRs), are seen to present serious waste management, technological and weapons proliferation risks.

- The potential for, and reality of, links between technologies and materials used for nuclear energy production and nuclear weapons have always given rise to concerns about nuclear weapons proliferation. The concerns about these connections have grown in the past few years as a result of nuclear programs in Iran, North Korea, India and Pakistan. Large-scale reprocessing activities to increase the nuclear fuel supply would seriously exacerbate the existing weapons proliferation risks.
- Nuclear generating facilities and their associated waste fuel storage facilities have been identified as potential targets for attack by groups or individuals motivated by political purposes, insanity or both. The consequences of such an attack, if successful, would be uniquely severe, as an attack could result in the release of large amounts of radioactive material to the atmosphere.

6.1. Introduction

In addition to the direct physical impacts of nuclear power generation, a number of issues are unique to nuclear power generation and must be considered in any assessment of the technology's environmental and economic sustainability.

Four key areas to be considered in this regard:

- Generating facility construction costs and timelines
- Generating facility performance and maintenance costs
- Uranium fuel supply and costs
- Weapons proliferation and security issues

6.2. Generating Facility Costs

Nuclear generating stations have a history of capital and maintenance cost overruns and low performance rates. This section outlines the history of these issues in Ontario's nuclear generation stations: Pickering, Bruce and Darlington. The large contribution of Ontario Hydro's nuclear program to Ontario's electricity debt is also discussed.

6.2.1. Facility Construction

The construction of a nuclear generating facility is a large engineering and financial undertaking.

Estimated environmental impacts of the manufacturing and construction of a typical nuclear power plant have been estimated as presented in **Table 6.1**.

Assuming a 700 MW reactor with a capacity factor of 85 per cent producing an annual output of 5,212,200 MW, and a lifetime of 25 years with a lifetime output of 130,305,000 MWh, this would imply emissions of 265,000 tonnes of CO_2 , 300 tonnes of SO_2 , 5,551 tonnes of NO2 and 378 tonnes of PM associated with plant construction.

As shown in **Table 6.2**, the actual construction costs of all five of Ontario's nuclear generating stations significantly exceeded their original estimated costs. All facilities were completed significantly behind schedule.

The high capital costs and long construction timelines for nuclear generating facilities present major challenges in attracting private investments in such projects. Most other electricity generating technologies have construction times of 2–4 years, compared to nuclear facilities with a minimum of 7–11 years before construction is completed and revenue generation can begin.⁸ Other generating technologies also have much lower initial capital costs per MW of capacity.⁹ Investments that bring more rapid and more secure returns are likely to be more attractive to private capital, particularly in the context of competitive electricity markets.¹⁰

Table 6.1:

Conventional Pollutant Emissions from Nuclear Plant Construction¹

CO ₂ (kg/MWh)	SO ₂ (kg/MWh)	NOx (kg/MWh)	PM (kg/MWh)
2.03	0.0023	0.0426	0.0029

Table 6.2: Cost Overruns	for Ontario	Nuclear Generating	g Stations ^{2,3,4,5,6, 1}
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Nuclear Generating Station	Estimated Capital Cost (\$billions)	Actual Capital Cost (\$billions)	Per Cent Overrun
Pickering A	0.508 (1965\$)	0.716 (1971\$)	40
Pickering B	1.585 (1974\$)	3.846 (1986\$)	140
Bruce A	0.930 (1969\$)	1.8 (\$ of the year)	90
Bruce B	3.929 (1976\$)	5.994 (\$ of the year)	50
Darlington	3.950 (1978\$) 7.4 (1993\$) projected from estimate in 1981	14.4 (1993\$)	270

The scale of the investments needed for nuclear facilities, whose designs cannot be easily scaled up or down, presents additional challenges.¹¹ AECL's proposed Advanced CANDU Reactor, for example, is a 700 MW unit, with an implied capital cost of \$1.5 billion per unit, or \$6 billion for a four-unit facility, excluding waste management and decommissioning costs.¹²

Most studies done on nuclear economics conclude that new plants built by the private sector, with investors bearing the full brunt of risks, are not economic without subsidy. The industry's legacy of cost growth, technological problems, cumbersome political and regulatory oversight, and the newer risks brought about by competition and terrorism makes it an unattractive investment otherwise.¹³ It has been estimated that privately financing a next generation CANDU 6 reactor in Ontario, including realistic expectations of return on capital, would result in an electricity cost per kilowatt-hour of 20.9 cents—three times that of a new combined cycle natural gas fired plant (7 cents), and almost 2.5 times the cost of renewable power (8.6 cents).¹⁴

6.3. Facility Performance and Maintenance¹

Ontario's CANDU nuclear generating stations have also been plagued by poor performance. The performance of nuclear generating stations is best described by capacity factors, or actual electricity production expressed as a percentage of what the generating station could have produced had it operated at full capacity for the entire period (i.e., if there had been no maintenance outages). **Table 6.3** lists the capacity factors of Ontario's nuclear generating stations between 1990 and 2004.

Between 1990 and 1996, Ontario's nuclear generating stations experienced particularly poor performance; the Pickering A station had an average capacity factor of 55per cent during this time and the Bruce A station had an average capacity factor of 52per cent. Capacity factors for Bruce B and Pickering B were 77per cent and 79per cent, respectively.

In response to regulatory concerns regarding the declining performance and safety of Ontario's nuclear

plants, an external review team was commissioned by Ontario Hydro to examine the utility's nuclear operations.²⁰ The team performed an Independent Integrated Performance Assessment (IIPA) of Ontario Hydro's nuclear generating stations. The results were presented to the Ontario Hydro Board in 1997. The IIPA recommended improvements to "managerial leadership, culture and standards, people and performance, processes and procedures, plant hardware and design, organization and resources, and labour relations."²¹

In response, Ontario Hydro adopted a Nuclear Asset Optimization Plan (NAOP). Under the plan, six generating units (Pickering units 1, 2 and 3, and Bruce units 1, 3, and 4) were taken out of service for repair and overhaul. Pickering 4 was already shut down the previous year during a maintenance check and the Bruce 2 reactor had been shut down in 1995 and was not restarted. Investments of between \$5 and \$8 billion over four years on the refurbishment of the laid-up units were announced.²² It was planned that the units would begin to return to service in 2000.²³

In practice, Ontario Hydro's successor, OPG, encountered major problems in bringing the units back into service. All of the refurbishment projects were subject to major cost overruns and delays. As shown in **Table 6.4**, the first units were not brought back into service until 2003.

The refurbishment of Pickering Units 2 and 3 has been abandoned by OPG as "uneconomical."³¹ Under an agreement signed with the Government of Ontario in October 2005, Bruce Power will invest \$4.25 billion to restart Units 1 and 2, refurbish Unit 3 when it reaches the end of its operational life and replace the steam generators in Unit 4. In exchange, the Government of Ontario provided price guarantees for power generated from the Bruce A units, and agreed to share up to 75 per cent of the cost overruns associated with the refurbishment project.³²

In addition to the direct impacts of the refurbishment projects noted in Chapter 4, the removal of the Pickering and Bruce A facilities from service for refurbishment had major collateral environmental impacts. Ontario Hydro and OPG have relied on their coalfired generating facilities (Lakeview [Mississauga], Nanticoke, Lambton, Thunder Bay and Atikokan) to replace the power supplies lost as a result of the taking out of service of the nuclear generating units. This led to major increases in emissions of smog and acid rain precursors, heavy metals, and GHGs from these facilities. Between 1995 and 2001, their GHG emissions increased by a factor of 2.3, and emissions of smog and acid rain precursors SO₂ and NOx had doubled

¹ The discussion in this section is adapted from D. Martin, "Ontario's Nuclear Generating Facilities: A History and Estimate of Unit Lifetimes and Refurbishment Costs, Appendix 2 in M. Winfield, M. Horne, T. McGlenaghan, and R. Peters, *Power for the Future: Torwards a Sustainable Electricity System for Ontario* (Toronto: Pembina Institute and Canadian Environmental Law Association, 2004).

	Pickering A	Pickering B	Bruce A	Bruce B	Darlington
1990	39.6	77.7	48.2	81.1	n/a
1991	55.9	89.7	64	88	n/a
1992	61.3	74.1	55.9	78.3	n/a
1993	80.3	81.6	34.3	67.2	81.7
1994	71.7	84.6	48.7	80.4	86.9
1995	41.7	83.4	52.9	77.2	90.1
1996	36.3	50.1	58.9	82.6	84.4
1997	72.5	59.1	39.3	78.5	61.5
1998	0	72.7	85.6	70.2	85.4
1999	0	77	0	75	83.3
2000	0	56.9	0	80.3	86.8
2001	0	73.2	0	80.3	85.6
2002	0	80.9	0	75	90.2
2003	70.319	67.8	n/a	85	81.7
Average	37.8	73.4	37.5	78.5	83.4

 Table 6.3: Average Capacity Factors at Nuclear Generating Stations, 1990–2003^{15,16,17,18}

Note: 'n/a' indicates that the measurement is not applicable.

Table 6.4: Performance Record and Restart Costs for Pickering A and Bruce A^{24, 25,26,27,28}

Reactor	Date of Closure for Maintenance	Date of Restart	Estimated Restart Cost (\$billions)	Actual Cost of Restart (\$billions)
Pickering 1	1997	2005	0.213 million29	1.01630
Pickering 2	1997	uneconomical		_
Pickering 3	1997	uneconomical		_
Pickering 4	1996	2003	0.458 (1999 esti- mate)	1.25
Bruce 1	1997	pending	4.25	_
Bruce 2	1995	pending		
Bruce 3	1998	2003	0.34 (2001\$)	0.72
Bruce 4	1998	2003		

and increased by a factor of 1.7, respectively.³³ The increased emissions from the Lambton, Nanticoke, and Lakeview facilities in particular have significantly exacerbated the severe air quality problems regularly experienced in southern Ontario³⁴ and emerged as a major political issue in the province.

The NAOP highlighted additional problems with CANDU reactors. Traditionally, Ontario Hydro and OPG have depreciated nuclear power plants assuming that their life expectancy was 40 years.³⁵ However, due to the demonstrated need for retubing and major rehabilitation at much earlier dates, a reasonable estimate of CANDU lifetime is no more than 25 years, in the absence of major rehabilitation efforts. This estimate is consistent with the position of the Canadian Nuclear Association, which has stated,

It is . . . assumed that all [CANDU] nuclear power plants will have to undergo one major rehabilitation program after 25 years in service to get to the full 40 years. Hence the forecast on availability depends on whether or not a decision is made to rehabilitate a nuclear plant after 25 years in service.³⁶

Ontario Hydro also estimated periods shorter than 40 years, when major rehabilitation of nuclear plants would be needed for removal and replacement of fuel channels and steam generators. Thus, it can be seen from **Table 6.5** that these dates were also typically close to 25 years after the initial date of first commercial operation.

There is a secondary question as to the period of time that reactors can be expected to operate *after* they have undergone retubing and/or rehabilitation. The only evidence in this regard is the experience of the Pickering A station that was retubed from 1983 to 1993. After being retubed, the Pickering reactor 1 lasted 10 years 3 months before being shutdown at the end of 1997; the Pickering reactor 2 lasted 9 years 1 month; the Pickering reactor 3 lasted 6 years 4 months; and the Pickering reactor 4 lasted 4 years 9 months.

After being shut down on December 31, 1997, the Pickering reactor 4 was restarted in October 2003, but OPG has not said how long it expects the reactor to operate. However, in 1997, as part of the Nuclear Asset Optimization Plan, Ontario Hydro suggested that the Pickering reactor 4 would require pressure tube replacement in March 2017.³⁸ This study has therefore assumed that the most recent rehabilitation of Pickering 4 will give a life expectancy of 13 years to the reactor, i.e., until 2016. This estimate is consistent with the estimate of Torrie Smith Associates.³⁹ This is also consistent with the position of the Ontario Power Generation Review Committee, which estimated a lifetime of 8–14 years for Pickering reactor 1 after refurbishment.⁴⁰

The OPG Review Committee has suggested that the Pickering B plant may last longer.⁴¹ It has suggested that the rehabilitation dates for the Pickering reactors are 2012–2016, or 29–30 years after the start of commercial operation. The committee's estimates for Bruce B and Darlington are consistent with the estimates of this study. The OPG Review Committee suggests that rehabilitation dates for Bruce B are 2009–2017 (24–30 years post-commercial operation), and for Darlington are 2013–2020 (21–27 years postcommercial operation).

When Ontario's electricity sector was restructured in 1999, Ontario Hydro had \$30.5 billion of debt and \$7.6 billion of other liabilities.⁴² To keep OPG solvent, \$19.433 billion of Ontario Hydro's accumulated debt or unfunded liabilities associated with electricity generation facilities was transferred to the Ontario Electricity Financial Corporation (OEFC) as "stranded debt" or "unfunded liabilities." Of this amount, \$15.147 billion was nuclear related.⁴³

Table 0.3. Tuel Chamile and Steam Generator Replacement Times	Table 6.	5. Fuel	Channel	and Steam	Generator I	Replacement Times ³³
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Station	Time Period	Years Past Commercial Operation
Bruce A Reactors 3 & 4	2001-2008	23-29
Pickering B	2008-2012	25-26
Bruce B	2010-2013	26–26
Darlington	2016-2019	26-26

6.4. Fuel Supply and Costs

6.4.1. Uranium Costs

Part of original the rationale for developing nuclear energy in Ontario was that it relied on a domestically sourced fuel that was not subject to political risks and international price shifts.⁴⁴

This is no longer the case. Uranium mining has ceased in northern Ontario, and the uranium mined in north Saskatchewan is an internationally traded commodity, like other non-renewable primary energy sources such, oil, natural gas and coal. Uranium prices are a function of the state of supply and international demand. They have risen dramatically in recent years, and are expected to remain high for the foreseeable future.

Global demand for uranium currently exceeds supply. A significant portion of uranium supply is from inventories of decommissioned nuclear warheads purchased by the United States from the Russian government under an agreement signed in 1993.⁴⁵ At the time, these inventories created a surplus, which resulted in a short-term decrease in the price of uranium.

However, as these inventories have diminished, the price of uranium has risen. **Figure 6.1** displays international uranium price patterns over the past two decades, illustrating an increase in the price of uranium since the beginning of 2001 by a factor of more than six, from $U_{10}^{0.08}$ per kg $U_{3}O_{8}$ to $U_{10}^{0.08}$ per kg in July 2006.⁴⁶

Globally, uranium forecasts out to 2008 indicate a shortage as demand outstrips supply, leading to continued price increases.48

6.4.2. Uranium Supply

Based on known global uranium reserves of slightly more than three million tonnes,⁴⁹ and assuming reactor requirements holding fairly steady at about 60,000 tonnes per year, it can be estimated that there is a 50year supply of uranium available for use. However, at current levels of nuclear energy production, some estimates have placed the world's high-grade uranium reserves at 19 years.⁵⁰ The uranium resources in northern Saskatchewan are generally estimated to be sufficient for approximately 40 years at current levels of consumption.⁵¹ The projected global supply constraints have lead to increased uranium exploration in Canada, particularly in the Athabasca Basin of northern Saskatchewan.⁵²

Although sufficient for several decades, at current levels of consumption the estimated Canadian reserves of uranium would run out before Canadian reserves of fossil fuels, particularly natural gas (estimated at 77 years)⁵³ and coal (estimated at several hundred years).⁵⁴ Significant environmental impacts, including additional waste rock and tailings generation would accompany the expansion of uranium mining activities, particularly as lower grade uranium deposits would have to be exploited. As noted in chapter 2, a number of jurisdictions maintain bans on the establishment of new uranium mines as a result of concerns over their potential environmental and health impacts.

It has been suggested that in the long term uranium extracted from sea water or thorium might be employed in alternative fuel cycles. However, these technologies do not currently exist commercially and face major challenges in terms of their technical and economic feasibility on a large scale.⁵⁵

Some countries reprocess spent fuel, which is estimated to displace about 2,000 tonnes of uranium demand per year. Reprocessing of spent uranium has the potential to expand the supply of fuel for nuclear reactors.⁵⁶



Figure 6.1: Price of Uranium $(US/kg U_3O_8)^{47}$

However, reprocessing was ruled out as an option in the Canadian context by the NWMO. The NWMO assessment concluded that the costs of building the necessary industrial capacity to undertake reprocessing would imply the need to commit to an expanded and multigenerational nuclear fuel cycle. NWMO also noted that reprocessing would increase the generation of high-level nuclear wastes.⁵⁷ The present process (the Purex process originally developed for the Manhattan project) produces 14 m³ of waste per tonne of spent fuel processed.⁵⁸

Additionally, as reprocessing plants separate plutonium from spent fuel, reprocessing would also carry with it risks of spreading technology and materials that could be used for the production of nuclear weapons.⁵⁹ In addition, large-scale reprocessing would involve extensive shipments of materials between reactors, reprocessing plants, and fresh fuel production facilities, most involving plutonium-containing materials. The proliferation concerns associated with such reprocessing led the United States government to prohibit the reprocessing of spent fuel from civilian plants in the 1970s.⁶⁰

In the longer term, it has been suggested that fast breeder reactors, which are able to convert uranium 238 into plutonium, thereby producing more fuel than they burn, could be used to provide additional fuel. It is generally thought that FBRs are several decades from even a prototype stage, and would be extremely expensive to build and operate.

Research on FBRs was abandoned in the United States in the 1970s after they were shown to increase the risks of nuclear weapons proliferation.⁶¹

6.5 Weapons Proliferation Risks

The use of nuclear energy for power production arose from the US Manhattan project during World War II. The early development of nuclear energy was largely for military purposes, and surrounded by great secrecy.

In the 1950s Canada decided not to pursue nuclear weapons development, but to focus on potential energy production instead.⁶⁷ Other countries, such as France, the Soviet Union and the United Kingdom, were more interested in the development of nuclear weapons, and designed their first reactors so that they could be used for plutonium production.⁶⁸ Even in Canada, the division between civilian and military uses was not complete, with Canadian uranium mines providing materials for United States and British weapons programs until the mid-1970s.⁶⁹ the technologies and materials associated with nuclear energy remains a subject of major global concern.

The Treaty on the Non-Proliferation of Nuclear Weapons, opened for signing in 1968 and ratified by 189 nations, requires non-nuclear weapons states (including Canada) to accept and maintain IAEA safeguards with respect to all source and special nuclear material, including spent fuel. Under the treaty, only the five nations that possessed nuclear weapons at the time of the creation of the treaty are permitted to possess nuclear weapons: United States, France, United Kingdom, Russia and China. These states are prohibited from transferring nuclear weapon technology to non-nuclear weapons states, and have entered into voluntary safeguard agreements arising from their civil nuclear power programs.⁷⁰

In practice it has proved difficult to keep nuclear energy and weapons development technologies apart. India's 1974 nuclear bomb, developed in part using Canadian-supplied technology and uranium, demonstrated this clearly.⁷¹ Recent high-profile international concerns regarding the nuclear weapons potential of the Iranian,⁷² North Korean,⁷³ Indian,⁷⁴ and Pakistani⁷⁵ nuclear programs continue to highlight the proliferation risks of nuclear energy technologies. A worldwide expansion of the role of nuclear energy, particularly if it involves large-scale reprocessing or FBRs, would likely exacerbate these risks.⁷⁶

In addition to the potential for the use of nuclear materials in the construction of nuclear weapons, materials generated through nuclear programs can be used to produce a "dirty" bomb—a conventional explosive combined with radioactive materials, intended to disperse radionuclides over a wide area, presenting radiation hazards and significant clean-up challenges.⁷⁷

For its part, the Environmental Audit Committee on the UK Parliament raised a number of concerns regarding the British government's nuclear expansion proposals in an April 2006 Report. The committee's key concerns included the following:⁷⁹

- The extremely long time frames for the construction of nuclear power plants relative to other options, meaning that they could make little contribution to shorter term gaps in generating capacity.
- The degree to which subsidies for nuclear facilities would displace investments in energy efficiency and low-impact renewable energy sources.
- The costs of plant decommissioning and longterm waste management.

The potential for weapons development offered by

6.6. Unique Security Risks

Nuclear generating facilities and their associated waste fuel storage facilities have been identified as potential targets for attack by groups or individuals motivated by political purposes, insanity or both. The consequences of such an attack, if successful, would be uniquely severe, as an attack could result in the release of large amounts of radioactive material to the atmosphere. Such outcomes would make response and recovery extremely difficult, and would result in environmental and health impacts over a much wider area than any other type of electricity generating facility subject to such an attack.⁸⁰

6.7. Conclusions

In addition to its physical environmental impacts, particularly waste generation, nuclear energy faces a number of unique economic, security and policy challenges.

Nuclear generating facilities are subject to very high capital costs, and long construction times. This makes it difficult for facilities to compete for private capital against potential investments that will bring much more rapid and secure returns on investment. Extraordinary financial guarantees by governments, for which taxpayers have assumed much of the financial risk, have been required to overcome these barriers. In addition, in Ontario there is a history of serious delays and cost overruns on nuclear generating facility generating projects, accounting for more than 75 per cent of the \$19 billion "stranded debt" left by Ontario Hydro. This history again reduces the attractiveness of nuclear projects to private sector investors.

The Ontario CANDU fleet has been plagued by performance and maintenance problems. In recent years the average fleet capacity factor has been in the 50 per cent range rather than the expected 85–90 per cent range. Reactors expected to have operational lifetimes in the range of 40 years turn out to require major refurbishments after approximately 25 years of service. Refurbishment projects themselves have run seriously over budget and behind schedule. Refurbished units appear to have operational lifetimes of 8–14 years. It had been thought units could be refurbished in indefinite cycles, although the OPG decision not to refurbish Pickering units 2 and 3 suggests there are limits to the degree to which this can be done.

The poor performance of the Ontario CANDU fleet has had major collateral environmental and health impacts. These impacts have been the result of the province's increased reliance on domestic and imported coal-fired generation from 1997 onwards to replace electricity that would have been provided by out-of-service nuclear units.

Origins of Canada's Nuclear Program

In 1942, a joint British-Canadian laboratory was started in Montreal, Quebec to design a nuclear reactor. The Zero Energy Experimental Pile (ZEEP), an experimental 10-watt research reactor, was constructed in Chalk River, Ontario, under the administration of the National Research Council. ZEEP had two purposes: to advance Canada's civilian nuclear program and to produce plutonium for the United States military in the production of nuclear weapons.⁶² The ZEEP reactor, the first large-scale nuclear reactor ever built in Canada, began operating in 1945 and continued to operate for 25 years.

Building on the knowledge acquired from the ZEEP reactor, the National Research Experimental (NRX) research reactor was started in 1947 at Chalk River, Ontario. Similar to the ZEEP reactor, the NRX reactor employed unenriched uranium as a fuel and heavy water as a moderator.⁶³ Also in 1947, a facility was built to extract plutonium from fuel rods irradiated in the ZEEP and NRX reactors for British and American nuclear weapons production.⁶⁴ In 1952, the Chalk River project became the crown corporation Atomic Energy of Canada Limited (AECL).

Nuclear power was mentioned for the first time in Ontario Hydro's 1954 annual report, and plans were put place to build a 20 MW demonstration power plant.⁶⁵ The 20 MW Nuclear Power Demonstration (NPD), a prototype CANDU reactor, was declared in-service at Rolphton, Ontario in 1962. In 1968, the first prototype commercial reactor–Douglas Point, in Kincardine, Ontario–was declared in-service. Located on the Bruce Peninsula on Lake Huron, the Douglas Point facility had a lifetime capacity factor of 50per cent over less than 18 years.⁶⁶ However, a major program of nuclear generating facility construction followed. Five commercial nuclear power stations with a total of 22 reactors were brought into service between 1971 and 1993 in Ontario, Quebec and New Brunswick.

Report of the UK Sustainable Development Commission⁷⁸

According to the UK Sustainable Development Commission (SDC), nuclear power is not the answer to tackling climate change or security of supply.

In response to the government's current Energy Review, the SDC nuclear report draws together the most comprehensive evidence base available and finds that there is no justification for bringing forward a new nuclear power program at present.

Based on eight new research papers, the SDC report gives a balanced examination of the pros and cons of nuclear power. Its research recognizes that nuclear is a low carbon technology, with an impressive safety record in the UK. Nuclear could generate large quantities of electricity, contribute to stabilising CO₂ emissions and add to the diversity of the UK's energy supply.

However, the research establishes that, even if the UK's existing nuclear capacity was doubled, it would only give an 8 per cent cut on CO_2 emissions by 2035 (and nothing before 2010). This must be set against the risks.

The report identifies five major disadvantages of nuclear power:

1. **Long-term waste** – no long-term solutions are yet available, let alone acceptable to the general public; it is impossible to guarantee safety over the long-term disposal of waste.

2. **Cost** – the economics of nuclear new-build are highly uncertain. There is little, if any, justification for public subsidy, but if estimated costs escalate, there's a clear risk that the taxpayer will have to pick up the tab.

3. **Inflexibility** – nuclear would lock the UK into a centralised distribution system for the next 50 years, at exactly the time when opportunities for microgeneration and local distribution networks are stronger than ever.

4. **Undermining energy efficiency** – a new nuclear program would give out the wrong signal to consumers and businesses, implying that a major technological fix is all that is required, weakening the urgent action needed on energy efficiency.

5. **International security** – if the UK brings forward a new nuclear power program, it cannot deny other countries the same technology. Having lower safety standards, other countries run higher risks of accidents, radiation exposure, proliferation and terrorist attacks.

On balance, the SDC finds that these problems outweigh the advantages of nuclear. However, the SDC does not rule out further research into new nuclear technologies and pursuing answers to the waste problem, as future technological developments may justify a re-examination of the issue.

SDC Chair, Jonathon Porritt, says:

It's vital that we get to grips with the complexity of nuclear power. Far too often, the debate is highly polarised, with NGOs claiming to see no advantages to nuclear at all, and the pro-nuclear lobby claiming that it's the only solution available to us.

Instead of hurtling along to a pre-judged conclusion (which many fear the Government is intent on doing), we must look to the evidence. There's little point in denying that nuclear power has benefits, but in our view, these are outweighed by serious disadvantages. The Government is going to have to stop looking for an easy fix to our climate change and energy crises—there simply isn't one.

Concluding with advice on a future energy strategy, the SDC report establishes that it is indeed possible to meet the UK's energy needs without nuclear power. With a combination of a low-carbon innovation strategy and an aggressive expansion of energy efficiency and renewables, the UK would become a leader in low-carbon technologies. This would enhance economic competitiveness whilst meeting the UK's future energy needs.

Fuel costs are emerging as a significant issue for the nuclear industry. World uranium prices have increased by a factor of more than six since 2001 and are expected to continue to rise. The question of security of long-term fuel supply for nuclear generating facilities is also emerging as a concern, particularly if there is a large-scale expansion of reliance on nuclear generation. Current Canadian uranium reserves are estimated to be sufficient for 40 years at present levels of consumption. This is significantly less than other non-renewable fuels, such as natural gas and coal. Efforts to increase the available fuel supply through the reprocessing of waste fuel or the use of fast breeder reactors are seen to present serious technological challenges and waste management and weapons proliferation risks.

The potential for, and reality of, links between technologies and materials used for nuclear energy production and nuclear weapons have always given rise to concerns about nuclear weapons proliferation. Concerns about these connections have grown in the past few years as a result of nuclear programs in Iran, North Korea, India and Pakistan. Large-scale reprocessing activities to increase the nuclear fuel supply would seriously exacerbate existing weapons proliferation risks.

Nuclear generating facilities and their associated waste fuel storage facilities have been identified as potential targets for attack by groups or individuals motivated by political purposes, insanity or both. The consequences of such an attack, if successful, would be uniquely severe, as it could result in the release of large amounts of radioactive material to the atmosphere.

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7. Summary and Conclusions

7.1. Introduction

This study examined the environmental impacts of the use of nuclear energy in Canada, through each of the four major stages of nuclear energy production: uranium mining and milling; uranium refining; conversion and fuel fabrication; nuclear power plant operation; and waste fuel management.

The impacts of nuclear energy production were examined in terms of waste generation, atmospheric releases of pollutants and contaminants, impacts on water quality and water use, landscape and ecosystem impacts, and occupational and community health impacts. In addition, a number of challenges to the long-term sustainability of the use of nuclear energy for electricity production were examined. These challenges included power plant construction costs and timeframes, power plant performance and associated maintenance costs, the fuel supply security and costs, and security and weapons proliferation risks.

7.2. Overview of Impacts

The annual physical impacts associated with nuclear energy production in Canada are summarized in **Table 7.1** using data for 2003 (except as noted), the most recent year for which relatively complete data is available. The figures for emissions and waste generation from mining, milling and refining, conversion and fuel fabrication facilities are adjusted to reflect the portion of these activities (approximately 16 per cent) associated with fuel production for domestic energy production purposes. The major findings are discussed in the following sections.

7.2.1. Waste Generation

The most significant short-term environmental impact of the use of nuclear power to produce electricity in Canada is the generation of a number of extremely large waste streams that contain a wide range of radioactive and hazardous contaminants, and some of which also represent serious security, accident and weapons proliferation risks. All of these waste streams will require care and management over very long time frames for safety, environmental and security reasons.

Major waste streams arise at each stage of the nuclear power production process. The key waste streams include the following:

Uranium mining and milling

- An estimated 575,000 tonnes of tailings per year, of which 90–100,000 tonnes can be attributed to uranium production for domestic energy purposes. Uranium mill tailings contain a range of radionuclides, heavy metals and other hazardous contaminants.
- Waste rock production may be as high as 18 million tonnes per year, of which 2.9 million tonnes can be attributed to mining for domestic energy production purposes.
- It is estimated that there are more than 213 million tonnes of uranium mine tailings in facilities in Canada.

Refining and conversion operations

• IAEA figures suggest that nearly 1,000 tonnes of solid wastes and 9,000 m³ of liquid wastes are produced per year as a result of uranium refining, conversion and fuel production for domestic energy generation purposes. Information on the precise character and fate of these wastes is not publicly available.

Power Plant operation

• Approximately 85,000 waste fuel bundles are

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Occupational and Community Health	Dust and radon exposure risks among miners. Elevated incidence of lung cancer reported among workers exposed to radon, particularly uranium miners. Continuing reported mortalities among uranium mine workers from historical exposure to silica, solvents, asbestos and radiation. Nuclear energy workers in min- ing and RCF sectors typically exposed to annual effective radiation. Nuclear energy workers in min- ing and RCF sectors typically exposed to annual effective radiation. Nuclear energy workers in min- ing and refease acceptable to general public. Particularly high occupational exposure occurs in the fuel fabrication process. Conventional occupational risks associated with mining and industrial plant operation. Community health risks from lichen-caribou-human food chain in vicinity of mines. Significant elevated cancer risks from consumption of caribou. Health impacts of RCF facilities contested in host communities, particularly Port Hope.
Land Impacts	Uranium mine surface lease areas total 13,866 ha. Impacts on water and biota extend beyond lease areas. RCF facilities and power plants typically occupy 1–4 km2. Evidence of contamination of biota and environment with radionuclides (uranium, radi- um-226, lead-210) as a result of wind- blown dust from mine sites. Approximately two million litres of ammonium nitrate by product from conversion sold as agricultural fertilizer. Estimated to contain up to 20 kg of uranium.
Water Impacts	Releases of radionuclides (uranium, lead-210, thorium-230), heavy metals (arsenic, nickel, lead, selenium, vanadium), other metals (copper, zinc) to surface waters from mine sites. Total discharges to surface waters in 2003 included over 240 kg of uranium, 13 kg of molybdenum, 11 kg of arsenic, 30 kg of nickel, and 6.5 kg of selenium. Extensive groundwater contamination at mine, TMF and waste rock storage areas (radionuclides (lead-210, polonium-210, radium-226), metals (arsenic, lead, cobalt, nickel, molybdenum, lead, manganese) and ions (potassium, magnesium, chlo- ride, calcium, sodium, sulphates)) Ammonia releases from mines and mills: 10 tonnes. Effluent from uranium mine/ mill operations found to be 'toxic' to non-human biota by Environment Canada.
Atmospheric Releases (annual, except as noted)	Radiation • Radon releases from mine ventilation systems, open pit mines, and mill operations. Radioactive contaminants • Radon, uranium and lead- 210 in dust from mine/mill sites. • 2.9 kg uranium from refining and conversion facilities. • Routine releases from power plants, principally of tritium oxide, as well as Carbon-14, lodine –131, noble gases, elemental tritium and radioactive particulate. • Releases, principally of tritium from incineration of low- and intermediate-level wastes from Ontario plant operations and maintenance.
Waste Generation (annual, except as noted)	Roughly 90–100,000 tonnes mine tailings. Radioactive, toxic, and acidic. Will require perpetual care. Up to 2.9 million tonnes waste rock. Potentially acid generat- ing, with leaching of radionu- clides and heavy metals. Will require perpetual care. Approximately 85,000 waste fuel bundles. Waste fuel man- agement issue unresolved. Environmental, health, security and weapons proliferation risks associated with waste fuel. Will require secure isolation for one million years. Generation of 990 tonnes solid wastes, and 9,000 m3 liquid wastes fuel production for domestic purposes. Approximately 6,000+ cubic metres low-level radioactive wastes from power plant operation and maintenance. Heavy metals, asbestos and hazardous wastes from main- tenance and refurbishment projects.

Waste Generation (annual, except as noted)	Atmospheric Releases (annual, except as noted)	Water Impacts	Land Impacts	Occupational and Community Health
Very large amounts of radio- active (high-, intermediate- and low-level) and hazardous wastes will be generated as a result of eventual facility decommissioning.	Conventional air pollutants • Sulphur oxides • 7,000 tonnes from Rabbit Lake acid plant alone. • Nitrogen oxides • 178–195 tonnes (mill, RCF and power plant operation). • Farticulate matter • 6 tonnes (excluding PM from mine sites). • Particulate matter • 6 tonnes from mill opera- tions. • Y9 tonnes from mill opera- tions. • Additional releases of (nitrogen and sulpur oxides and PM from incineration of low- and intermediate- level radioactive wastes in Ontario. Hazardous air pollutants • Releases of dioxins, furans, hexachlorobenzene from refining and conversion, incineration of low- and intermediate-level wastes (releases reduced signifi- cantly with new incinerator commissioned in 2003).	Discharges of ammonia, nitrates and phosphorous from refining and conversion facilities. Discharges of approximately 1 kg uranium from refining and conversion facilities. Routine and large accidental releases of tritium oxide, beta- gamma activity and carbon-14 from power plant operations. Tritium oxide contamination of groundwater at Pickering site. Historical discharges of copper and zinc (1,000+ tonnes) from Pickering plant. Minor discharges of chromium from Bruce and Darlington from all plants. Ontario nuclear generating facilities are the only power plants in Canada reporting discharges of hydrazine (1.3 tonnes total). Groundwater contamination risks with long-term waste stor- age and underground disposal. Major impacts on ground and surface water storage and flows from mining operations.		Impacts of routine releases of radionuclides to atmosphere and surface waters from power plant operations. Risks of large-scale community radiation exposure due to accidents/incidents at electricity generating facilities or during transportation of waste fuel.

Waste Generation (annual, except as noted)	Atmospheric Releases (annual, except as noted)	Water Impacts	Land Impacts	Occupational and Community Health
	 Heavy metals Releases of lead, mercury, cadmium reported from power plants. Larger lead releases (10.8 kg in 2004) occasionally reported from Port Hope conversion facility. Releases from incineration of low- and intermediate radioactive level wastes in Ontario. Heavy metals may be contrained in dust from mine sites but not reportable to NPRI. Hydrazine Ontario nuclear plants are the only NPRI reported source of releases to the attmosphere in Canada. Ammonia Ammonia Ammonia Bydrogen Fluoride Minor releases from Refining/Conversion and fuel fabrication. Hydrogen sulphide Historically significant releases (up to 2,000 tonnes per year) from heavy water production. 228,000+ tonnes arising from plant construction. 	Historical large use of water (5.4 trillion litres) for heavy water production. Large use of water for cooling water at power plants; 8.9 trillion litres/yr used for Picking and Darlington alone. Thermal impacts on fish populations.		

generated by Canadian nuclear reactors each year. As of 2003, 1.7 million bundles were in storage at reactor sites. It is estimated that these wastes will have to be secured for approximately one million years for safety, environmental and security reasons.

- Approximately 6,000 cubic metres of lower level radioactive wastes are generated each year in Ontario as a result of power plant operations, maintenance, and refurbishment. A wide range of radiological and hazardous pollutants has been released to the atmosphere as a result of the incineration of these wastes at the Bruce Western Waste Management facility. A new incinerator installed at the in 2003 has reduced emissions of hazardous, but not radiological, pollutants.
- Power plant maintenance and refurbishment also result in the generation of substantial amounts of additional hazardous wastes, including heavy metals and asbestos.
- Very large amounts of low-, intermediate- and high-level radioactive wastes will be produced as a result of the eventual decommissioning of refining, conversion and fabrication facilities and power plants. The costs of decommissioning Ontario's existing reactors have been estimated at \$7.474 billion (present value \$6.263 billion)

Proposals have been advanced since the 1970s for the management of waste fuel, but no long-term management strategy has been adopted to date. The Nuclear Waste Management Organization (NWMO), established by the federal government in 2002, has proposed a "phased adaptive management strategy," for waste fuel from existing reactors. The proposal is currently under consideration by the federal government. It is estimated that implementation of the NWMO's proposed strategy would have a total cost in the range of \$24 billion, and extend over a period of more than 300 years.

There is no national management strategy in place or under development for low-and intermediate- level radioactive wastes arising from uranium refining and conversion activities and nuclear power plant operation, maintenance and refurbishment and eventual decommissioning. Ontario Power Generation is currently pursuing a proposal for a deep underground repository for low and intermediate-level wastes generated by its facilities.

The effectiveness and adequacy of tailings management facilities at mine sites has been subject to serious question. There is a long history of uranium mine tailings management facility failures in Canada and elsewhere in the world, resulting in severe surface water and groundwater contamination. In addition, windblown dust from tailings management facilities has contaminated surrounding environments and biota. The generation of waste rock and tailings increase proportionally with the use of lower grade uranium ores, as larger amounts of ore would have to be processed to produce the same amount of uranium concentrate. The processing of ore that is 0.01% uranium, for example, would generate approximately ten times the amount of tailings compared to the processing of ore that is 0.1% uranium.

The extremely long timelines over which the wastes arising from reliance on nuclear power will have to be managed, make the projection of future risks and costs arising from their management extremely difficult. The resulting situation raises important challenges to the use of nuclear energy for electricity production. The nature and volume of the wastes generated mean significant or unknown risks and costs will be passed on to generations far into the future as a result of energy consumption in the present.

7.2.2. Water Impacts

The second major area of impacts is water related. These impacts relate to both water quality and the quantities of surface and groundwater affected in the course of the production of nuclear energy.

7.2.2.1. Water Quality Impacts

Severe contamination of surface water and groundwater with radionuclides, heavy metals and other pollutants has arisen from uranium mine tailing management facilities and mine and mill operations. Discharges to surface waters from uranium mines and mills in Canada in 2003 included over 1,500 kg of uranium, 860 kg of molybdenum, 70 kg of arsenic, 185 kg of nickel, 40 kg of selenium, and 10 tonnes of ammonia.

Concentrations of major ions (potassium, magnesium, bicarbonate, chloride, calcium, sodium and sulphate) ranging from 10 to 200 times the levels of un-impacted groundwater have been found in areas near uranium mine tailings management facilities. Effluent from uranium mines and mills was found by Health Canada and Environment Canada to be 'toxic' for the purposes of the *Canadian Environmental Protection Act* in 2004.

Routine and accidental releases of radionuclides to surface waters occur in the course of power plant operations, with tritium oxide and carbon-14 being the key radioactive pollutants of concern. Groundwa-
ter contamination with tritium has also occurred at the Pickering generating facility in Ontario.

Ontario's nuclear power plants are the leading source of discharges of hydrazine, an extremely hazardous pollutant, to surface waters in Canada. Nuclear generating facilities have also been sources of discharges of metals (copper, zinc, and chromium) and ammonia to surface waters. Minor discharges of ammonia, nitrates and phosphorous to surface waters occur from refining and conversion facilities, along with discharges of approximately 1 kg of uranium per year.

7.2.2.2. Water use

Nuclear power generation is a major consumer of water. Uranium mining operations involve extensive dewatering (pumping out of groundwater from the mine to avoid flooding), in the range of at least 16–17 billion litres per year, with the implication of impacts on groundwater and surface water storage and flows. Approximately 5.4 trillion litres of water were historically used in heavy water production for CANDU reactors.

Generating facilities require large amounts of cooling water. The Darlington and Pickering facilities use approximately 8.9 trillion litres of water for cooling purposes per year — more than 19 times the annual water consumption of the City of Toronto. Adverse thermal impacts of cooling water discharges on fish populations in the vicinity of nuclear power plants have been observed.

7.2.3. Atmospheric Impacts

Contrary to frequent statements that nuclear energy provides "emission free" energy, there are releases of criteria, radioactive and hazardous air pollutants and greenhouse gases throughout the nuclear energy production process.

7.2.3.1. Radionuclides

Atmospheric releases of a range of radionuclides occur at all stages of nuclear power production. Atmospheric releases of radon gas occur as a result of mining and milling operations and from tailings management facilities. Windblown dust from mine sites and TMFs contains a range of radionuclides (e.g. uranium, Ra-226, Pb-210, and Po-210). Atmospheric releases (principally uranium) arise from refining and conversion activities. Routine and accidental releases of radiation and radionuclides occur from power plant operations, including tritium oxide, carbon-14, noble gases, iodine-131, radioactive particulate and elemental tritium. The incineration of low and intermediate-level radioactive wastes from power plant operations and maintenance in Ontario has resulted in further atmospheric releases of radionuclides, particularly tritium.

7.2.3.2. Hazardous air pollutants.

Windblown dust from mine sites and TMFs contains a range of heavy metals. In addition, releases of a number of hazardous air pollutants, including dioxins and furans, hexachlorobenzene, heavy metals (principally lead) ammonia and hydrogen fluoride arise from uranium refining and conversion operations. Ontario nuclear power plants are the only NPRI reported source of releases of hydrazine to the air in Canada. A wide range of hazardous air pollutants have been released to the atmosphere as a result of the incineration of low and intermediate-level radioactive wastes at the Bruce Western Waste Management facility. A new incinerator installed at the in 2003 has reduced emissions of hazardous, but not radiological, pollutants Major historical releases (up to 2,000 tonnes per year) of H₂S were associated with heavy water production in Canada for CANDU reactors.

7.2.3.3. Criteria air pollutants

Mining and milling operations are major sources of releases of SO_2 (43,000 tonnes per year from the Rabbit Lake Acid plant alone), VOCs (500 tonnes per year) and NOx (400 tonnes per year). Additional releases of NOx, PM, and sulphuric acid arise from refining and conversion activities, and from the road transportation of uranium from mill sites in northern Saskatchewan to the Blind River refinery in northern Ontario and then on to the Port Hope conversion facility in southern Ontario. Minor releases of criteria air pollutants are associated with the testing of fossil fuel-powered emergency generating equipment at nuclear generating facilities. Further transportation related releases of criteria air pollutants would arise from the long-term management of waste nuclear fuel and other radioactive wastes arising from facility operations, maintenance and decommissioning, particularly if the management strategies for these materials require the movement of wastes from reactor sites to centralized facilities.

7.2.3.4. Greenhouse gas emissions

GHG emissions arise at each stage of the nuclear energy cycle, with power plant construction being the most significant source of releases. Further releases of GHGs occur as a result of the operation of equipment in the uranium mining process, the milling of uranium ore, mill tailings management activities, and refining and conversion operations. The generation of greenhouse gases from mining and milling operations would increase proportionally with the use of lower grade uranium ores, as larger amounts of ore would have to be extracted and processed to produce the same amount of uranium concentrate.

The road transportation of uranium between milling, refining and conversion facilities results in additional releases. As with criteria air pollutants, the management of waste nuclear fuel, and other radioactive wastes could involve significant transportation activities, leading to further generation of greenhouse gases emissions.

In Canada, total GHG emissions associated with uranium mining, milling, refining, conversion and fuel fabrication are between 240,000 and 366,000 tonnes of CO₂ per year. Total emissions associated with the sector, including the emissions associated with power plant construction, are in the range of 468,000 and 594,000 tonnes of CO₂ per year, equivalent to the emissions of between 134,000 and 170,000 cars per year. Total annual GHG emissions associated solely with domestic power production are estimated at between 267,000 and 289,000 tonnes of CO₂ per year. This total is almost certainly an underestimate, due to a lack of complete information. Other recent estimates suggest total GHG emissions associated with nuclear power in Canada are in the range of at least 840,000 tonnes per year.¹⁰

7.2.4. Landscape Impacts

Uranium mine surface lease areas in Canada total 13,866 ha. However, as a result of discharges of contaminants to surface water and groundwater and windblown dust from mine sites containing radionuclides and heavy metals, the impacts of mining and milling activities extend well beyond lease areas.

Similarly, though refining, conversion and fuel production facilities and power plants typically occupy a relatively small area (1–4 km²), they can easily have impacts well beyond their fence lines. For example, 3.5 million m³ of historic wastes associated with the Port Hope conversion facility remain within the town. These wastes contain uranium, radium and their radioactive decay products, together with various heavy metals. Another 872,000 m³ lie in the area immediately west of Port Hope at Welcome and Port Granby. Approximately two million litres of ammonium nitrate by-product from the conversion process are reported to be sold as agricultural fertilizer each

year. This material has been estimated to contain up to 20 kg of uranium.

7.2.5. Occupational and Community Health

The International Agency for Research on Cancer (IARC) lists a number of radionuclides as carcinogenic to humans, including isotopes produced in uranium mining and milling, fuel production and nuclear power plant operations. It has been argued that existing standards in Canada for cancer risks arising from radiological hazards permit much higher levels of acceptable risk than is the case for chemical and other hazards. More broadly, recent research on the effects of even very low levels of ionizing radiation suggests that no level is safe to health. The risk of cancer has been found to be greatest for women and children and to be higher for younger children.

Workers in the mining and refining, conversion and fuel fabrication sub-sectors are routinely exposed to levels of radiation above those which would be considered acceptable to members of the general public. Particularly high levels of occupational exposure occur in the fuel fabrication process.

There is a history of significant occupational health effects, particularly elevated incidences of lung cancer, among uranium miners attributed to radon exposure. Increased mortality among uranium miners is also attributed to exposure to silica, solvents, asbestos and radiation.

The community health impacts of nuclear facility operation is a matter of ongoing controversy. Substantial health risks have been identified in relation to the consumption of certain types of "country" food, particularly caribou, in the vicinity of uranium mine/mill operations, as a result of contamination by radionuclides.

Nuclear generating facility operators argue that the levels of public exposure to radiation arising from facility operations are trivial in comparison to other sources. These claims are contested, as recent studies suggest that health impacts of low-level radiation exposure may be more significant than previously thought, and that children and infants may be particularly at risk from such exposures.

There are long-standing community concerns regarding the health impacts of the Port Hope conversion facility in particular. There are also continuing concerns with respect to the adequacy of existing drinking water standards in Ontario for tritium and carbon-14. The existing Ontario drinking water standard for tritium of 7,000 Bq/L is significantly weaker than the standards adopted in other jurisdictions. The standard in the United States is 740 Bq/L and in the European Union is 100 Bq/L.

Finally, there are low probability, but extremely high impact, risks of large-scale community radiation exposure due to accidents or incidents at electricity generating facilities, at waste fuel storage facilities or during the transportation of waste fuel. It has been estimated, for example, that the monetized value of the off-site environmental, health and economic impacts of a major accident at the Darlington generating facility east of the City of Toronto, would exceed \$1 trillion.

7.2.6. Data and Study limitations

The findings outlined in this section likely underestimate the overall impacts of the use of nuclear energy for electricity production in Canada. The findings are focussed on impacts about which information could be accessed in the public domain.

The lack of coverage by the NPRI of the impact of mining activities up to the primary crushing stage results in significant data gaps with respect to the generation and fate of pollutants related to uranium mining activities. The non-inclusion of radionuclides in the NPRI results in additional information gaps in the reporting of releases and transfers of these substances resulting from uranium mining and milling, refining, conversion and fuel fabrication, and nuclear power plant operation. Virtually no information was available regarding fate of wastes from refining, conversion and fuel fabrication operations.

In addition, key pieces of information were found to be scattered among a variety of sources, ranging from individual facilities and companies to regulatory bodies. In some cases, freedom of information requests were required to obtain basic information on environmental releases of pollutants. In other cases, important information proved to be inaccessible due to business confidentiality issues.

Finally, the study relied on what were likely conservative estimates from industry-related sources in a number of key areas, particularly with respect to the generation of GHG emissions, and waste generation from the refining, conversion and fuel fabrication processes.

7.2.7. Impacts of Non-CANDU type nuclear reactors.

The study findings are focussed on the impacts of the operation of CANDU type reactors as these are the only type of reactor used for electricity generation in Canada. Different types of reactors are associated with different impacts and risks. Light-water reactors, employing enriched uranium fuel, for example, are associated with the generation of lower volumes of waste fuel. However, the process of producing enriched uranium fuel for these types of reactors is associated with much higher emissions of greenhouse gases, particular where gas diffusion based enrichment processes are employed, as well as higher atmospheric releases of uranium and the generation of large volumes of depleted uranium (DU) wastes.

7.3. Sustainability Challenges

In addition to its physical environmental impacts, particularly waste generation, nuclear energy generation faces a number of unique challenges relative to energy generation via other sources.

7.3.1. Capital costs and construction times

Nuclear power generating facilities are subject to very high capital costs and long construction times relative to other electricity supply options. In addition, in Ontario there is a history of serious delays and cost overruns on nuclear generating facility projects, accounting for \$15 billion of the nearly \$20 billion "stranded debt" left by Ontario Hydro. Even with extensive subsidies and financial guarantees provided by government, these costs, timelines and risks make it difficult for nuclear power projects to compete for private capital investments against potential investments that will bring much more rapid and secure returns.

7.3.2. Facility reliability

The Ontario CANDU reactor fleet has been plagued by performance and maintenance problems. Over the past decade the Ontario facilities have had an average operating capacity in the 50per cent range rather than the expected 85–90 per cent range. Reactors expected to have operational lifetimes in the range of 40 years have turned out to require major refurbishments after approximately 25 years of service. Refurbishment projects themselves have run seriously over budget and behind schedule.

The poor performance of the Ontario CANDU fleet has had major collateral adverse effects on the province's environment and the health of its residents. These impacts have been the result of the province's increased reliance on domestic and imported coalfired generation to replace electricity that would have been provided by out-of-service nuclear units. As a result of the shutdown of eight reactors under the 1997 Nuclear Asset Optimization Plan between 1995 and 2001, emissions of GHGs from the province's coal-fired power plants increased by a factor of 2.3, sulphur dioxide emissions by a factor of 2, and nitrogen oxide emissions by a factor of 1.7, significantly exacerbating the severe air quality problems regularly experienced in southern Ontario.

7.3.3. Fuel supply and costs

Uranium supply and costs are emerging as significant issues for reliance on nuclear power for electricity generation. World uranium prices increased more than sixfold since 2001 and are expected to continue to rise. Uranium has experienced a much higher rate of price increase compared to other fuels in recent years, including natural gas. The question of the security of long-term uranium supplies is also an emerging concern, particularly if there is a large-scale expansion of reliance on nuclear generation. Current Canadian uranium reserves are estimated to be sufficient for 40 years at present levels of consumption. Efforts to increase the available fuel supply through the reprocessing of waste fuel, or the use of fast breeder reactors, present serious waste management, technological and weapons proliferation risks.

Other suggested fuel sources, such as thorium or extraction of uranium from seawater, face major technological, environmental and economic hurdles. Significant environmental impacts, including additional waste rock and tailings generation would accompany the expansion of uranium mining activities, particularly as lower grade uranium deposits would have to be exploited.

7.3.4. Weapons proliferation

Nuclear energy's shared origins with nuclear weapons programs raise the potential for and reality of links between technologies and materials used for energy production and for nuclear weapons development. Concerns about these connections have grown in the past few years as a result of nuclear programs in North Korea, Iran, India and Pakistan. Any large-scale expansion of reliance on nuclear energy would carry significant risks of the proliferation of materials and technologies that could be applied to weapons development. India's 1974 nuclear bomb, developed in part using Canadian-supplied technology and uranium, demonstrated this problem clearly.

7.3.5. Security risks

Nuclear generating facilities and their associated waste fuel storage facilities have been identified as potential targets for attack by groups or individuals motivated by political purposes, insanity or both. The consequences of such attacks, if successful, would be uniquely severe; they could result in the release of large amounts of radioactive material to the atmosphere, which could be distributed over a large area. By comparison, the impacts of major incidents or accidents at facilities employing other generating technologies would be short term and largely limited to the facility site itself.

7.4. Conclusions

This study has sought to portray, in as complete a manner as possible based on publicly available and accessible information sources, a picture of the environmental impacts resulting from the use of nuclear energy for electricity generation in Canada. It is intended to inform comparisons with other energy sources.

The study finds that nuclear power, like other nonrenewable energy sources, is associated with severe environmental impacts. In short, the use of nuclear power for electricity generation cannot be considered "clean." Each stage of the nuclear energy production process generates large amounts of uniquely difficultto-manage wastes that will require perpetual care, and that effectively push costs and risks arising from current energy consumption onto future generations. The process also has severe impacts on surface water and groundwater water quality via a range of radioactive and hazardous pollutants, and results in releases to the atmosphere of a wide range of criteria, radioactive and hazardous pollutants as well as GHGs.

In addition, the technology poses unique occupational and community health risks, along with security, accident and weapons proliferation risks not shared by any other energy source.

Nuclear generating facilities suffer from high capital costs and long-construction times, with significant risks of cost overruns and delays. These are major barriers to private investment in the sector, particularly in combination with a history of poor facility performance in Ontario and emerging challenges regarding uranium costs and supply.

In the context of these impacts and risks, nuclear energy cannot be seen as a viable response to GHG emission problems associated with reliance on fossil fuels (e.g., coal) for electricity generation. In addition to the consideration that nuclear power is not itself a GHG emission-free energy source, a future path based on nuclear energy would simply replace one problem (GHG emissions) with a series of different, but equally unacceptable impacts and risks. These impacts and risks encompass everything from facility reliability and waste management to the potential for catastrophic accidents and nuclear weapons proliferation.

As a result, proposals for the retention and expansion of the role of nuclear power must be approached with the greatest of caution. Such proposals must be examined in the full light of their environmental, economic and security implications, not only for Canada, but the rest of the world as well. They must also be examined in the context of the full range of available alternatives. Such an examination is likely to conclude that better options are readily available. These options range from making the most efficient use possible of existing energy resources to expanding the role of lowimpact renewable energy sources that offer far safer, cheaper, more reliable and more sustainable options for meeting society's energy needs.

Abbreviations

ACES - Advisory Committee on Environmental Standards AECB - Atomic Energy Control Board AECL - Atomic Energy of Canada Limited BASS - Bruce Alternate Steam Supply BHWP - Bruce Heavy Water Plant CANDU - CANada Deuterium Uranium CEPA - Canadian Environmental Protection Act CH4 - methane CNSC - Canadian Nuclear Safety Commission CO - carbon monoxide CO_2 – carbon dioxide DU - depleted uranium DU - depleted uranium EPA - Environmental Protection Agency FBRs - fast breeder reactors GE – General Electric GHG - greenhouse gas H₂S – hydrogen sulphide HVAC - heating ventilation and air conditioning IAEA - International Atomic Energy Agency IARC - International Agency for Research on Cancer ICUCEC - Inter-Church Uranium Committee **Educational Cooperative** IIPA - Independent Integrated Performance Assessment IPCC - Intergovernmental Panel on Climate Change LILW - Low- and medium- (or intermediate-) level waste LLRWMO - Low Level Radioactive Waste Management Office MOE - Ministry of the Environment N_2O – nitrous oxide NAOP - Nuclear Asset Optimization Plan NFWMA - nuclear fuel waste management agency NGOs - Non-governmental organization NH4+ - ammonium NOx – nitrogen oxide NOx – nitrogen oxides NPD - Nuclear Power Demonstration NPRI - National Pollutant Release Inventory NRC - National Research Council NRCan - National Resources Canada NRX - National Research Experimental NSCA - Nuclear Safety and Control Act NSERC - Natural Sciences and Engineering Research Council of Canada NWMO - Nuclear Waste Management Organization **OEFC - Ontario Electricity Financial Corporation** ONFA – Ontario Nuclear Funds Agreement OPA - Ontario Power Authority

OPG - Ontario Power Generation PCBs - Polychlorinated Biphenyls PHAI - Port Hope Area Initiative PM - particulate matter PSL - Priority Substance List PVC – Polyvinyl chloride PWR - pressurized light water reactors RWMA - Radioactive Waste Management Associates SDC - Sustainable Development Commission SO_2 – sulphur dioxide SOx – sulphur oxide SQ - significant quantity TEMS - Treatment Effluent Management System TEQ - toxicity equivalent TMA - Tailings Management Area TMFs - tailings management facilities U – uranium U_3O_8 – uranium oxide UF₆ – uranium hexafluoride UO₂ - uranium dioxide UO₃ - uranium trioxide US DOE - US Department of Energy VOCs - volatile organic compounds ZEEP - Zero Energy Experimental Pile

Glossary

Radionuclides – "A radionuclide is an atom with an unstable nucleus. The radionuclide undergoes radioactive decay by emitting a gamma ray(s) and/or subatomic particles. Radionuclides may occur naturally, but can also be artificially produced."²

NB NWMO Discussion document 1 discussion of radiation risks.

Low-level radioactive waste – "Low-level waste includes items that have become contaminated with radioactive material or have become radioactive through exposure to neutron radiation. This waste typically consists of contaminated protective shoe covers and clothing, wiping rags, mops, filters, reactor water treatment residues, equipments and tools, luminous dials, medical tubes, swabs, injection needles, syringes, and laboratory animal carcasses and tissues."³

Radiation – "Radiation in physics is a process of emission of energy or particles."⁴

Radioactivity – "The radioactivity is the property of some atoms to spontaneously give off energy as particles or rays. The atoms that make up the radioactive materials are the source of radiation."⁵

Sieverts (Sv) – The Sievert is the SI derived unit of dose equivalent that reflects the biological effects of radiation.⁶

Gamma radiation - Gamma rays are an energetic form of electromagnetic radiation produced by radioactive decay or other nuclear or subatomic processes.⁷

Becquerels (Bq) – The becquerel (symbol Bq) is the SI derived unit of radioactivity, defined as the activity of a quantity of radioactive material in which one nucleus decays per second.⁸

Radon – A radioactive noble gas formed by the disintegration of radium, radon is one of the heaviest gases and is considered to be a health hazard.⁹

Tailings – Produced during the milling process, tailings consist largely of ground rock particles, water and various amounts of mill chemicals; they also contain radioactive and hazardous constituents.

Endnotes

1. Introduction

- 1 M. Winfield, M. Horne, T. McClenaghan, and R. Peters, *Power for the Future: Towards a Sustainable Electricity System for Ontario* (Toronto: Pembina Institute and Canadian Environmental Law Association, 2004).
- 2 Environment Canada, *Priority Substances List Assessment Report. Releases of Radionuclides from Nuclear Facilities (Impact on Nonhuman Biota)* (Ottawa: Government of Canada, May 2003).
- 3 The Power Authority is an agency created through the *Electricity Restructuring Act, 2004.* The Authority's mandate includes electricity system planning. The Authority is subject to directives issued by the Minister of Energy. www.powerauth ority.on.ca.
- 4 Ontario Power Authority, *Supply Mix Advice Report* (Toronto: Ontario Power Authority, December 2005), pg.49.
- 5 See, for example, G. Scotton, "French Minister Nuclear Pitchman," *Calgary Herald*, March 29, 2006. See also B. Duane, "Implementing Kyoto in Canada: The Role of Nuclear Power," *The Energy Journal* 26, no. 1 (insert year here), pp.107–21.
- 6 The Hon. D. Duncan, Ontario Minister of Energy, *Supply Mix Directive to Ontario Power Authority*, June 13, 2006, 2006).
- 7 Ontario Power Generation is a successor corporation to Ontario Hydro. It owns the province's existing nuclear generating facilities and operates the Picking and Darlington Facilities. www.opg.com.
- 8 Ministry of Energy, "McGuinty Government Delivers a Balanced Plan for Ontario's Electricity Future," news release, June 13, 2006.
- 9 See generally M. Winfield, R. Peters and M. Horne, *The Ontario Power Authority's Supply Mix Advice: A Review and Response* (Toronto: The Pembina Institute, February 2006).
- 10 Dr. D. McKeown, Medical Office of Health, City of Toronto, "Letter to Premier McGuinty Re: Ontario Power Authority's Supply Mix Advice Report," February 3, 2006.
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- 12 Ontario Power Generation (OPG), "Pickering Nuclear," Accessed November 21, 2005 at http://www.opg.com/ops/ N_pickering.asp
- 13 The consortium of Bruce Power consists of Cameco Corporation, TransCanada Corporation, BPC Generation Infrastructure Trust (a trust established by the Ontario Municipal Employees Retirement System), the Power Workers' Union and The Society of Energy Professionals. Source: Bruce Power, "About," Accessed August 29, 2005 at http://www.brucepower.com/bpcms_web/pagecontent.aspx? navuid=111
- 14 Bruce Power, "Annual Review, 2001–2002," Accessed November 21, 2005 at http://www.brucepower.com/bpcms_ web/uc/GetDocument.aspx?docid=945
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7. Summary and Conclusions and Glossary

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